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Conference of the Parties to the   
Minamata Convention on Mercury

Fourth meeting

Online, 1–5 November 2021[[1]](#footnote-2)\*

Item 4 (d) of the provisional agenda[[2]](#footnote-3)\*\*

Matters for consideration or action by the Conference of the Parties: mercury waste: consideration of the relevant thresholds

Technical guidelines on the environmentally sound management of wastes consisting of, containing or contaminated with mercury or mercury compounds

Note by the secretariat

1. Paragraph 3 of article 11 of the Minamata Convention on Mercury stipulates that each party is to take appropriate measures so that mercury waste is managed in an environmentally sound manner, taking into account the guidelines developed under the Basel Convention on the Control of Transboundary Movements of Hazardous Wastes and Their Disposal and in accordance with requirements that the Conference of the Parties is to adopt in an additional annex in accordance with Article 27.
2. In 2015, at its twelfth meeting, the Conference of the Parties to the Basel Convention, in its decision BC-12/4, adopted technical guidelines on the environmentally sound management of wastes consisting of, containing or contaminated with mercury or mercury compounds,[[3]](#footnote-4) The Conference of the Parties to the Minamata Convention, in its decision MC-2/2 on mercury waste thresholds, invited the Conference of the Parties to the Basel Convention to consider reviewing, as appropriate, those guidelines, with additional guidance for certain mercury wastes. The Conference of the Parties to the Basel Convention decided, in its decision BC-14/8, that the guidelines should be updated.
3. The Conference of the Parties to the Minamata Convention, in its decision MC-3/5 on mercury waste thresholds, encouraged the parties and other stakeholders to contribute to the process of updating the guidelines by providing comments on the draft updated guidelines when invited to do so, and invited the appropriate bodies of the Basel Convention to take into account decision MC-3/5 in updating the guidelines.
4. The secretariat of the Minamata Convention participated in the meetings of the small intersessional working group established pursuant to decision BC-14/8 of the Conference of the Parties to the Basel Convention and provided an update on the progress made to the group of technical experts on mercury waste thresholds established pursuant to decision MC-2/2.
5. The draft updated technical guidelines,[[4]](#footnote-5) which were prepared by Japan, as the lead country, in consultation with the small intersessional working group, were submitted to the Conference of the Parties to the Basel Convention. The draft updated technical guidelines are reproduced in the annex to the present note. The changes made to the existing guidelines have been tracked so that the revisions can be easily identified. The annex has not been formally edited.

Annex

Technical guidelines on the environmentally sound management of wastes consisting of, containing or contaminated with mercury or mercury compounds

(Draft updated version of 1 February 2021)

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Abbreviations and acronyms

|  |  |  |
| --- | --- | --- |
| AAS  AFS  ASGM | atomic absorption spectrometry  atomic fluorescence spectroscopy  artisanal and small-scale gold mining | |
| ASTM | American Society for Testing and Materials | |
| AOX | absorbable organic halides | |
| BAT  CAS | best available techniques  Chemical Abstracts Service | |
| CEN | European Committee for Standardization | |
| CFLs | compact fluorescent lamps | |
| CH3Hg+ or MeHg+ | monomethylmercury, commonly called methylmercury | |
| Cl | chlorine | |
| EMS | environmental management system | |
| EN | European standard | |
| EPR | extended producer responsibility | |
| ESM | environmentally sound management | |
| FAO  GEF  GEMS  GHS | Food and Agriculture Organization of the United Nations  Global Environment Facility  Global Environmental Monitoring System  Globally Harmonized System | |
| HCl | hydrochloric acid | |
| HF  HFC | hydrofluoric acid  hearth furnace coke | |
| Hg | mercury | |
| HgCl2 | mercury dichloride | |
| HgO | mercury (II) oxide | |
| HgS | mercury sulphide or cinnabar | |
| HgSO4 | mercury sulphate | |
| HNO3 | nitric acid | |
| IAEA | International Atomic Energy Agency | |
| IATA | International Air Transport Association | |
| ICAO  ICCM  ICP  IIED | International Civil Aviation Organization  International Conference on Chemicals Management  inductively coupled plasma  International Institute for Environment and Development | |
| ILO | International Labour Organization | |
| IMERC | Interstate Mercury Education and Reduction Clearinghouse | |
| IMO | International Maritime Organization | |
| ISO | International Organization for Standardization | |
| JIS | Japanese Industrial Standard | |
| JLT  JSA | Japanese standardized leaching test  Japanese Standards Association | |
| LCD | liquid crystal displays | |
| LED  LFLs | light emitting diode  linear fluorescent lamps | |
| MEE | Ministry of Ecology and Environment (China) | |
| MMSD | Mining, Minerals and Sustainable Development (IIED/WBCSD project) | |
| MSW | municipal solid waste | |
| NEMA  NEWMOA | National Electrical Manufacturers Association  Northeast Waste Management Officials’ Association | |
| NIMD | National Institute for Minamata Disease | |
| NOx | nitrogen oxide | |
| OEWG | Open-ended Working Group (of the Basel Convention) | |
| OECD | Organisation for Economic Co-operation and Development | |
| PACE | Partnership for Action on Computing Equipment | |
| PBB | polybrominated biphenyls | |
| PBDE | polybrominated diphenyl ethers | |
| PCB  PCDDs/PCDFs | polychlorinated biphenyl  polychlorinated dibenzodioxins/polychlorinated dibenzofurans | |
| PRTR | pollutant release and transfer register | |
| PVC | polyvinyl chloride | |
| QA/QC  QSC | quality assurance/quality control  Quicksilver Cancus | |
| QSP | Quick Start Programme | |
| RoHS | Directive 2011/65/EU of the European Parliament and of the Council of 8 June 2011 on the restriction of the use of certain hazardous substances in electrical and electronic equipment (RoHS Directive) | |
| SAICM | Strategic Approach to International Chemicals Management | |
| SETAC | Society of Environmental Toxicology and Chemistry | |
| SO2 | sulphur dioxide | |
| SOP | standard operational procedure | |
| SPC  SPSS | sulphur polymer cement  sulphur polymer stabilization and solidification | |
| S/S | stabilization and solidification | |
| TCLP | toxicity characteristic leaching procedure | |
| TOC | total organic carbon | |
| UNDP | United Nations Development Programme | |
| UNECE | United Nations Economic Commission for Europe | |
| UNEP | United Nations Environment Programme | |
| UNIDO  UNITAR | United Nations Industrial Development Organization  United Nations Institute for Training and Research | |
| USEPA | Environmental Protection Agency (United States of America) | |
| VCM | vinyl chloride monomer | |
| WEEE | waste electrical and electronic equipment | |
| WHO | World Health Organization | |

Units of measurement

|  |  |
| --- | --- |
| μg | microgram |
| mg | milligram |
| g | gram |
| kg | kilogram |
| mg/kg | milligram(s) per kilogram. Corresponds to parts per million (ppm) by mass. |
| L | liter |
| m3 | cubic meter |
| cm3 | cubic centimeter |
| ℃ | degree Celsius |

I. Introduction

A. Scope

1. The present technical guidelines provide guidance on the environmentally sound management (ESM) of wastes consisting of, containing, or contaminated with mercury[[5]](#footnote-6) or mercury compounds[[6]](#footnote-7), hereinafter referred to as “mercury wastes”, pursuant to decisions under the Basel and the Minamata Conventions[[7]](#footnote-8). This document supersedes the Basel Convention *Technical guidelines for the ESM of wastes consisting of, containing or contaminated with mercury or mercury compounds*, adopted by the Conference of the Parties at its twelfth meeting.
2. In addition, the present technical guidelines address provisions referred to in Article 11, paragraph 3 (a) to (c)and Article 3, paragraphs 4 and 5 (b) of the Minamata Convention. Article 11, paragraph 3 (a), of the Minamata Convention provides for a link between the present guidelines and the obligations of Parties to the Minamata Convention to manage mercury waste in an environmentally sound manner. For further details, see section II.B.1 of the present guidelines.
3. The present guidelines focus on mercury wastes categorized as hazardous wastes or other wastes as defined under the Basel Convention and mercury wastes as defined by the Minamata Convention.
4. The following mercury wastes are covered by the present guidelines (see tables 1 and 4 for additional information[[8]](#footnote-10)):

A: Wastes consisting of mercury or mercury compounds (e.g., excess mercury from the decommissioning of chlor-alkali facilities, mercury recovered[[9]](#footnote-11) from wastes containing mercury or mercury compounds or wastes contaminated with mercury or mercury compounds, or surplus stock of mercury or mercury compounds designated as waste);

B: Wastes containing mercury or mercury compounds, with the following subcategories:

B1: Wastes of products containing mercury or mercury compounds that easily release mercury into the environment, including when they are broken (e.g., mercury thermometers, fluorescent lamps);

B2: Wastes of products containing mercury or mercury compounds other than those listed in B1 (e.g., batteries); and

C: Wastes contaminated with mercury or mercury compounds (e.g., residues generated from mining processes, industrial processes or waste treatment processes).

1. Article 11, paragraph 1, of the Minamata Convention provides for a link between the relevant definitions of the Basel Convention, such as its definition of wastes, and wastes covered by the Minamata Convention. Under Article 11, paragraph 2, of the Minamata Convention, only those wastes (a) consisting of, (b) containing or (c) contaminated with mercury or mercury compounds in a quantity above the relevant thresholds defined by the Conference of the Parties to the Convention are defined as mercury wastes. This definition excludes overburden, waste rock and tailings from mining, except from primary mercury mining, unless they contain mercury or mercury compounds above thresholds defined by the Conference of the Parties[[10]](#footnote-14). Wastes of mercury-added products[[11]](#footnote-15) belong to (b) wastes containing mercury or mercury compounds. According to Article 3, paragraph 5 (b), of the Minamata Convention, excess mercury from the decommissioning of chlor-alkali facilities belongs to (a) wastes consisting of mercury or mercury compounds.
2. The definition of wastes under the Basel Convention is described in paragraph 13 below, and the mercury wastes under the scope of the Basel Convention are described in section II.A.2. It should be noted that the there is no specific definition of “mercury wastes” in the Basel Convention. The mercury thresholds described in the Minamata Convention are independent from the provisions on determination of hazardous wastes and requirements for transboundary movements under the Basel Convention.

B. About mercury[[12]](#footnote-16)

1. Mercury is or has been widely used in products such as measuring devices (barometers, hygrometers, manometers, thermometers, sphygmomanometers), switches and relays, fluorescent lamps, batteries, cosmetics, pesticides, biocides, topical antiseptics and dental amalgam, and in manufacturing processes, such as those involving the production of chlor-alkali, acetaldehydes, vinyl chloride monomers (VCM), sodium or potassium methylates or ethylates, polyurethane, gold plating and products containing mercury or mercury compounds. Mercury may also be a by‑product of raw material refining or production processes, such as oil and gas refining and non-ferrous metal production.
2. Mercury emissions and releases can be caused by human activities (i.e., they may be anthropogenic) but may also result from natural sources such as volcanic eruptions. Once mercury is emitted or released into the environment, it persists in the atmosphere (e.g., as mercury vapour, as compounds in particulate form, or on particulates), soil (e.g., as ionic mercury) and water bodies   
   (e.g., methylmercury (MeHg+, or CH3Hg+)). Mercury in the air can be carried around the world, eventually being deposited onto soils, waters or plants. From there, it can re-volatilize into the air, or be transported further by water, or be taken into the food web (UNEP, 2019a). The preamble of the Minamata Convention recognizes that mercury is a chemical of global concern owing to its   
   long-range atmospheric transport, its persistence in the environment once released, its ability to bioaccumulate in ecosystems and its significant negative effects on human health and the environment.
3. Handling, collection, transportation or disposal of mercury wastes can lead to emissions or releases of mercury particularly if carried out improperly.
4. The discharge of wastewater containing mercury into Minamata Bay, Japan, from 1932 to 1968 (Ministry of the Environment of Japan, 2002), the illegal dumping of mercury-contaminated waste in Cambodia in 1998 (Honda et al., 2006; NIMD, 1999) and the case involving Thor Chemicals in South Africa (Lambrecht, 1989) are a few examples of cases in which wastes containing or contaminated with mercury or mercury compounds were not managed in an environmentally sound manner.
5. The provisions of the Minamata Convention are intended, inter alia, to reduce mercury supply and demand. The growing global trend towards phasing out mercury-added products and processes using mercury may cause temporary increases in some mercury wastes, particularly wastes consisting of mercury. In some countries, there has been increased use of some products containing mercury or mercury compounds, such as fluorescent lamps, which are being used to replace incandescent lamps, and liquid crystal display (LCD) backlights containing mercury. However, fluorescent lamps are replaced by light emitting diodes (LEDs) in some countries, which are more energy efficient and mercury-free alternatives.

II. Relevant provisions of the Basel Convention and international linkages

A. Basel Convention

1. General provisions

1. The Basel Convention aims to protect human health and the environment against adverse effects resulting from the generation, management, transboundary movements and disposal of hazardous and other wastes.
2. In Article 2 (“Definitions”), paragraph 1, the Basel Convention defines wastes as “substances or objects which are disposed of or are intended to be disposed of or are required to be disposed of by the provisions of national law”. Article 2, paragraph 4, of the Convention defines disposal as “any operation specified in Annex IV” to the Convention. Annex IV contains two categories of operations: those leading to the possibility of resource recovery, recycling, reclamation, direct reuse or alternative uses (R operations) and those not leading to this possibility (D operations).
3. Article 4 (“General obligations”), paragraph 1, establishes the procedure by which Parties exercising their right to prohibit the import of hazardous wastes or other wastes for disposal are to inform the other Parties of their decision. Paragraph 1 (a) states: “Parties exercising their right to prohibit the import of hazardous or other wastes for disposal shall inform the other Parties of their decision pursuant to Article 13”. Paragraph 1 (b) states: “Parties shall prohibit or shall not permit the export of hazardous or other wastes to the Parties which have prohibited the import of such waste when notified pursuant to subparagraph (a)”.
4. Article 4, paragraphs 2 (a) - (e) and 2 (g), and paragraph 8, set out key provisions pertaining to ESM, waste minimization, reduction of transboundary movement, and waste disposal practices aimed at mitigating adverse effects on human health and the environment:

Paragraphs 2 (a) - (e) and 2 (g): “Each Party shall take appropriate measures to:

(a) Ensure that the generation of hazardous wastes and other wastes within it is reduced to a minimum, taking into account social, technological and economic aspects;

(b) Ensure the availability of adequate disposal facilities, for ESM of hazardous wastes and other wastes, that shall be located, to the extent possible, within it, whatever the place of their disposal;

(c) Ensure that persons involved in the management of hazardous wastes or other wastes within it take such steps as are necessary to prevent pollution due to hazardous wastes and other wastes arising from such management and, if such pollution occurs, to minimize the consequences thereof for human health and the environment;

(d) Ensure that the transboundary movement of hazardous wastes and other wastes is reduced to the minimum consistent with the environmentally sound and efficient management of such wastes, and is conducted in a manner which will protect human health and the environment against the adverse effects which may result from such movement;

(e) Not allow the export of hazardous wastes or other wastes to a State or group of States belonging to an economic and/or political integration organization that are Parties, particularly developing countries, which have prohibited by their legislation all imports, or if it has reason to believe that the wastes in question will not be managed in an environmentally sound manner, according to criteria to be decided on by the Parties at their first meeting; and

(g) Prevent the import of hazardous wastes and other wastes if it has reason to believe that the wastes in question will not be managed in an environmentally sound manner.”

Paragraph 8: “Each Party shall require that hazardous wastes or other wastes, to be exported, are managed in an environmentally sound manner in the State of import or elsewhere”.

1. The Ban Amendment entered into force on 5 December 2019, and it provides that Parties listed in Annex VII to the Convention (members of the European Union, OECD and Liechtenstein) shall prohibit transboundary movements to States not listed in Annex VII of hazardous wastes which are destined for operations according to Annex IV.A and hazardous wastes under Article 1, paragraph 1(a) which are destined for operations according to Annex IV.B[[13]](#footnote-18).

2. Mercury-related provisions

1. Article 1 (“Scope of the Convention”) defines the types of waste that are covered by the Basel Convention. Paragraph 1 (a) sets out a two-step process for determining whether a “waste” is a “hazardous waste” covered by the Convention; first, the waste must belong to one of the categories listed in Annex I to the Convention (“Categories of wastes to be controlled”), and second, it must possess at least one of the characteristics listed in Annex III to the Convention (“List of hazardous characteristics”).
2. Annex I wastes are presumed to exhibit one or more of the hazardous characteristics listed in Annex III. These may include H6.1 “Poisonous (acute)”, H8 “Corrosive”, H11 “Toxic (delayed or chronic)”, or H12 “Ecotoxic”, unless, through “national tests”, they can be shown not to exhibit such characteristics. National tests may be useful for identifying a particular hazardous characteristic listed in Annex III until such time as the hazardous characteristic is fully defined. Guidance papers[[14]](#footnote-19) for some Annex III hazardous characteristics have been drafted under the Convention.
3. List A of Annex VIII describes wastes that are “characterized as hazardous under Article 1, paragraph 1 (a) of this Convention” although “their designation on this Annex does not preclude the use of Annex III [hazard characteristics] to demonstrate that a waste is not hazardous”. List B of Annex IX includes wastes that “will not be wastes covered by Article 1, paragraph 1 (a), of this Convention unless they contain Annex I material to an extent causing them to exhibit an Annex III characteristic”.
4. As stated in Article 1, paragraph 1 (b), “Wastes that are not covered under paragraph (a) but are defined as, or are considered to be, hazardous wastes by the domestic legislation of the Party of export, import or transit” are also subject to the Basel Convention.
5. As stated in Article 1, paragraph 2, “Wastes that belong to any category contained in Annex II that are subject to transboundary movement shall be “other wastes” for the purposes of this Convention”.
6. Mercury wastes listed in annexes I, II and VIII to the Convention are listed in table 1 below.

**Table 1:** Mercury wastes listed in annexes I, II and VIII to the Basel Convention (emphasis added)

|  |  |
| --- | --- |
| Entries with direct reference to mercury | |
| Y29 | Wastes having as constituents:  ***Mercury***, ***mercury compounds*** |
| A1010 | Metal wastes and waste consisting of alloys of any of the following:  …   * ***Mercury***   …  but excluding such wastes specifically listed on list B. |
| A1030 | Wastes having as constituents or contaminants any of the following:  …   * ***Mercury; mercury compounds***   … |
| A1180 | Waste electrical and electronic assemblies or scrap[[15]](#footnote-21) containing components such as accumulators and other batteries included on list A, ***mercury-switches***, glass from cathode-ray tubes and other activated glass and PCB-capacitors, or contaminated with Annex I constituents (e.g. cadmium, ***mercury***, lead, polychlorinated biphenyl) to an extent that they possess any of the characteristics contained in Annex III (note the related entry on list B B1110)[[16]](#footnote-23) |
| **Other entries related to wastes which may contain or be contaminated with mercury** | |
| Y1 | Clinical wastes from medical care in hospitals, medical centers and clinics |
| Y2 | Wastes from the production and preparation of pharmaceutical products |
| Y3 | Waste pharmaceuticals, drugs and medicines |
| Y4 | Wastes from the production, formulation and use of biocides and phytopharmaceuticals |
| Y11 | Waste tarry residues arising from refining, distillation and any pyrolytic treatment |
| Y12 | Wastes from production, formulation and use of inks, dyes, pigments, paints, lacquers, varnish |
| Y46 | Wastes collected from households |
| Y47 | Residues arising from the incineration of household wastes |
| Y48 | Plastic waste, including mixtures of such waste, with the exception of the following[[17]](#footnote-24) |
| A1170 | Unsorted waste batteries excluding mixtures of only list B batteries. Waste batteries not specified on list B containing Annex I constituents to an extent to render them hazardous |
| A2030 | Waste catalysts but excluding such wastes specified on list B |
| A2060 | Coal-fired power plant fly-ash containing Annex I substances in concentrations sufficient to exhibit Annex III characteristics (note the related entry on list B B2050) |
| A3170 | Wastes arising from the production of aliphatic halogenated hydrocarbons (such as chloromethane, dichloro-ethane, vinyl chloride, vinylidene chloride, allyl chloride and epichlorhydrin) |
| A3210 | Plastic waste, including mixtures of such waste, containing or contaminated with Annex I constituents, to an extent that it exhibits an Annex III characteristic (note the related entries Y48 in Annex II and on list B B3011) |
| A4010 | Wastes from the production, preparation and use of pharmaceutical products but excluding such wastes specified on list B |
| A4020 | Clinical and related wastes; that is wastes arising from medical, nursing, dental, veterinary, or similar practices, and wastes generated in hospitals or other facilities during the investigation or treatment of patients, or research projects |
| A4030 | Wastes from the production, formulation and use of biocides and phytopharmaceuticals, including waste pesticides and herbicides which are off-specification, outdated, or unfit for their originally intended use |
| A4070 | Wastes from the production, formulation and use of inks, dyes, pigments, paints, lacquers, varnish excluding any such waste specified on list B (note the related entry on list B B4010) |
| A4080 | Wastes of an explosive nature (but excluding such wastes specified on list B) |
| A4100 | Wastes from industrial pollution control devices for cleaning of industrial off-gases but excluding such wastes specified on list B |
| A4130 | Waste packages and containers containing Annex I substances in concentrations sufficient to exhibit Annex III hazard characteristics |
| A4140 | Waste consisting of or containing off specification or outdated[[18]](#footnote-26) chemicals corresponding to Annex I categories and exhibiting Annex III hazard characteristics |
| A4160 | Spent activated carbon not included on list B (note the related entry on list B B2060) |

B. International linkages

1. Minamata Convention on Mercury

1. The objective of the Minamata Convention, which entered into force on 16 August 2017, is to protect human health and the environment from anthropogenic emissions and releases of mercury and mercury compounds. To achieve this objective, the Minamata Convention aims, *inter alia*:
   1. To reduce the supply of mercury and control the international trade in mercury;
   2. To reduce the demand for mercury in products, manufacturing processes and artisanal and small-scale gold mining;
   3. To reduce emissions and releases of mercury to air, land and water;
   4. To ensure environmentally sound interim storage of mercury and mercury compounds intended for use;
   5. To ensure ESM of mercury wastes; and to address management of sites contaminated with mercury or mercury compounds;
   6. To promote capacity-building, technical assistance and technology transfer, including through specific financial and other arrangements.

**(a) Waste-related provisions**

1. Article 11 (“Mercury wastes”) of the Minamata Convention sets forth the following waste-related provisions:

“1. The relevant definitions of the Basel Convention on the Control of Transboundary Movements of Hazardous Wastes and Their Disposal shall apply to wastes covered under this Convention for Parties to the Basel Convention. Parties to this Convention that are not Parties to the Basel Convention shall use those definitions as guidance as applied to wastes covered under this Convention.

2. For the purposes of this Convention, mercury wastes means substances or objects:

(a) Consisting of mercury or mercury compounds;

(b) Containing mercury or mercury compounds; or

(c) Contaminated with mercury or mercury compounds,

in a quantity above the relevant thresholds defined by the Conference of the Parties, in collaboration with the relevant bodies of the Basel Convention in a harmonized manner, that are disposed of or are intended to be disposed of or are required to be disposed of by the provisions of national law or this Convention. This definition excludes overburden, waste rock and tailings from mining, except from primary mercury mining, unless they contain mercury or mercury compounds above thresholds defined by the Conference of the Parties.

3. Each Party shall take appropriate measures so that mercury waste is:

(a) Managed in an environmentally sound manner, taking into account the guidelines developed under the Basel Convention and in accordance with requirements that the Conference of the Parties shall adopt in an additional annex in accordance with Article 27. In developing requirements, the Conference of the Parties shall take into account Parties’ waste management regulations and programmes;

(b) Only recovered, recycled, reclaimed or directly re-used for a use allowed to a Party under this Convention or for environmentally sound disposal pursuant to paragraph 3 (a);

(c) For Parties to the Basel Convention, not transported across international boundaries except for the purpose of environmentally sound disposal in conformity with this Article and with that Convention. In circumstances where the Basel Convention does not apply to transport across international boundaries, a Party shall allow such transport only after taking into account relevant international rules, standards, and guidelines.

4. The Conference of the Parties shall seek to cooperate closely with the relevant bodies of the Basel Convention in the review and update, as appropriate, of the guidelines referred to in paragraph 3 (a).

5. Parties are encouraged to cooperate with each other and with relevant intergovernmental organizations and other entities, as appropriate, to develop and maintain global, regional and national capacity for the management of mercury wastes in an environmentally sound manner.”

1. In relation to Article 11(2) of the Convention, the Conference of the Parties to the Minamata Convention adopted at its third meeting the decision MC-3/5 on thresholds of mercury wastes with the following key elements:
2. No threshold needs to be established for mercury wastes falling under paragraph 2 (a) of Article 11, and wastes listed in table 2 below are regarded as such waste;
3. No threshold needs to be established for mercury wastes falling under paragraph 2 (b) of Article 11, and mercury-added products that are disposed of, are intended to be disposed of or are required to be disposed of, including those listed in table 3 below, will be regarded as such mercury waste;
4. The group of technical experts will further substantiate its present recommendation that a totals concentration threshold may be appropriate for mercury wastes falling under subparagraph 2 (c) of Article 11, including a technical analysis of options and the consideration of possible impacts;
5. There is no need at present to develop thresholds for overburden and waste rock from mining other than primary mercury mining, and thresholds for tailings from mining other than primary mercury mining should be established in a two-tiered approach using a totals concentration threshold as an initial screen and a leaching threshold as the second tier, and requests the group of technical experts to do further work to establish the thresholds;
6. The group of technical experts will conduct analysis of whether tailings from artisanal and small-scale gold mining should be subject to a threshold, taking into account the relationship between Articles 11 and 7;
7. The group of technical experts will develop thresholds for mercury waste falling under 2 (c) of Article 11 and recommend thresholds for tailings from industrial-scale non-ferrous metal mining other than primary mercury mining;
8. The group of technical experts will, subject to completion of items (e) and (f) above, review, and possibly recommend a revision of, the lists of mercury waste falling under subparagraphs 2 (a) to (c) of Article 11, set out in tables 2, 3 and 4 below, as appropriate.

List of mercury waste consisting of mercury or mercury compoundsa (subparagraph 2 (a) of article 11) as per Table 1 in the annex to decision MC-3/5

| *Type of waste* | *Waste sourceb* |
| --- | --- |
| Recovered elemental mercuryc | Mining activity:   * Tailings from artisanal and small-scale gold mining   Mercury captured from:   * Non-ferrous metals roasting and smelting processes * Crude oil and natural gas processing   Treatment of:   * Mercury-added products upon becoming waste * Waste contaminated with mercury or mercury compounds * Contaminated environmental media   Treatment of waste from:   * Chlor-alkali,d alcoholates (e.g., sodium or potassium methylate or ethylate), dithionite and ultrapure potassium hydroxide solution production with mercury technology, including decommissioning facilities * Polyurethane, vinyl chloride monomer, acetaldehyde production using a mercury-containing catalyst |
| Elemental mercury | Mercury stockpile (e.g. industries, laboratories, dental offices, educational and research institutions, government institutions, landfills, dumpsites, lighthouses) |
| Mercury (I) chloride and mercury (II) chloride | Primary zinc, lead, copper and gold roasting and smelting processes  Reagent  Calomel electrode for electrochemical measurements  Medicine/pharmaceuticals  Vinyl chloride monomer catalyst – mercury (II) chloride  Stockpiles |
| Mercury (II) oxide (mercuric oxide) | Dry cell batteries, pigment in paints and glass modifiers, fungicide, cosmetics, laboratory reagent, antifouling paints  Stockpiles |
| Mercury (II) sulfate (mercuric sulfate) | Lab reagent, catalyst used for the production of acetaldehyde  Stockpiles |
| Mercury (II) nitrate (mercuric nitrate) | Oxidizing agent, laboratory reagent  Stockpiles |
| Cinnabar concentrate | Primary mercury mining  Stockpiles |
| Mercury sulfide | Pigment  Stabilization of waste mercury for storage and/or disposal  Stockpiles |

a The Convention defines a mercury compound as any substance consisting of atoms of mercury and one or more atoms of other chemical elements that can be separated into different components only by chemical reactions.

b A facility or activity where waste is likely to be generated or accumulated.

c Recovered mercury as described in subparagraph 3 (b) of article 11.

d Recovery can sometimes occur without treatment.

Table 3: Non-exhaustive list of waste containing mercury or mercury compounds (subparagraph 2 (b) of article 11)a as per Table 2 in the annex to decision MC-3/5

| *Type of waste* | *Waste sourceb* |
| --- | --- |
| Non-electronic measuring devices containing mercury (barometers, hygrometers, manometers, thermometers, sphygmomanometers) | Hospitals, clinics, healthcare facilities (human and animal), pharmacies, households, schools, laboratories, universities, industrial facilities, airports, meteorological stations, ship recycling facilities |
| Electrical and electronic switches, contacts, relays and rotating electrical connectors with mercury | Dismantling facilities of waste electrical and electronic equipment (relays, connectors and switches), industrial facilities (attached to boilers), households, offices |
| Fluorescent bulbs, high intensity discharge (HID) bulbs (mercury vapour bulbs, metal halide and high-pressure sodium bulbs), neon/argon lamps | Households, industrial and commercial facilities, automobile facilities, collection points |
| Batteries/accumulators containing mercury | Households, industrial and commercial facilities, collection points |
| Biocides and pesticides containing mercury and their formulations and products | Agricultural, horticultural, industrial and commercial facilities (including stockpiles), laboratories |
| Paints and varnishes containing mercury | Industrial and commercial facilities, households |
| Pharmaceuticals containing mercury for human and veterinary uses, including vaccines | Industrial and healthcare facilities (including stockpiles), livestock industry |
| Cosmetics and related products containing mercury | Industrial facilities (including stockpiles) |
| Dental amalgam | Dental offices, dental schools, crematoria |
| Scientific instrument used for the calibration of medical or scientific devices containing mercury | Laboratories, institutionsc (including stockpiles) |

a Mercury-added products listed in the “type of waste” column of this table are regarded as waste under subparagraph 2 (b) of article 11 when they are disposed of, are intended to be disposed of or are required to be disposed of by the provisions of national law or the Minamata Convention.

b A facility or activity where waste is likely to be generated or accumulated.

c Institutions include public and private ones.

Table 4: Indicative list of waste contaminated with mercury or mercury compounds (subparagraph 2 (c) of article 11)a as per Table 3 in the annex to decision MC-3/5

| *Type of waste* | *Waste sourceb* |
| --- | --- |
| Waste from industrial pollution control devices or cleaning of industrial off-gasesc | Flue gas from sources such as:  Extraction and use of fuels/energy sources  Smelting and roasting processes in the production of non-ferrous metals  Production processes with mercury impurities  Recovery of precious metals from waste electrical and electronic equipment  Coal combustion  Waste incineration and co-incineration  Crematoria |
| Bottom ash | Coal combustion  Biomass fired power and heat generation  Waste incineration |
| Wastewater treatment residues/slurriesd | Treatment of wastewater from:  Extraction and use of fuels/energy  Production of mercury-added products  Manufacturing processes in which mercury or mercury compounds are used  Primary non-ferrous metals production  Production processes with mercury impurities  Recovery of precious metals from waste electrical and electronic equipment  Waste incineration, co-incineration and other thermal treatment  Crematoria  Healthcare facilities  Controlled landfills leachate  Uncontrolled dumping of wastes  Agricultural facilities |
| Sludge | Separator tanks and sedimentary sand tanks for refining of crude oil, natural gas production and processing, drilling, ship cleaning, chemical processes, etc.  Treatment of wastes contaminated with mercury (e.g., chemical precipitation and chemical oxidation) |
| Oil and gas refining catalyst | Refining of crude oil  Processing of natural gas |
| Tailings and extraction process residues | Primary mercury mining  Artisanal and small-scale gold mining |
| Rubbles, debris and soile | Construction/demolition  Remediation of contaminated sites |
| Other waste from manufacturing processes using mercury or mercury compoundsf | Chlor-alkali production with mercury technology  Production of alcoholates (e.g., sodium or potassium methylate or ethylate)  Dithionite and ultrapure potassium hydroxide solution  Vinyl chloride monomer (VCM) production with mercuric chloride (HgCl2) catalyst  Acetaldehyde production with mercury sulphate (HgSO4) catalyst, etc. |
| Other waste from the manufacturing of mercury-added productsg | Manufacturing of mercury-added products |
| Other waste from natural gas cleaningh | Natural gas cleaning |
| Wastes from waste treatment facilitiesi | Waste treatment facilities |

a Wastes listed in this table are regarded as mercury waste when they exceed thresholds. Waste exceeding the established threshold but not listed here would also be considered mercury waste.

b A facility or activity where waste is likely to be generated or accumulated.

c Includes filters and activated carbon.

d Include filters and resins.

e Contaminated soil transported off-site is regarded as waste.

f Mercury cells, mercury recovery units (retort), waste catalysts, decommissioning or demolition waste, personal protective equipment, elements used to contain mercury spills, etc.

g Process residues, demolition waste, etc.

h Scale removed from pipework and pipe cleaning equipment, etc.

i Waste treated to stabilize/solidify mercury in the waste, fluorescent coatings, metal and glass.

**(b)**  **Other provisions relevant to mercury wastes**

1. Paragraphs 27 to 34 below describe other provisions of the Minamata Convention that are relevant to mercury wastes.
2. Article 3 of the Minamata Convention addresses sources of mercury supply and trade. Paragraph 4 provides that “each Party shall only allow primary mercury mining that was being conducted within its territory at the date of entry into force of the Convention for it for a period of up to fifteen years after that date. During this period, mercury from such mining shall only be used in manufacturing of mercury‑added products in accordance with Article 4, in manufacturing processes in accordance with Article 5, or be disposed in accordance with Article 11, using operations which do not lead to recovery, recycling, reclamation, direct re-use or alternative uses”. Paragraph 5 (b) provides that “each Party shall take measures to ensure that, where the Party determines that excess mercury from the decommissioning of chlor-alkali facilities is available, such mercury is disposed of in accordance with the guidelines for environmentally sound management referred to in paragraph 3 (a) of Article 11, using operations that do not lead to recovery, recycling, reclamation, direct re-use or alternative uses”.
3. Article 4 of the Convention addresses mercury-added products. Paragraph 1 provides that “each Party shall not allow, by taking appropriate measures, the manufacture, import or export of mercury-added products listed in Part I of Annex A after the phase-out date specified for those products, except where an exclusion is specified in Annex A or the Party has a registered exemption pursuant to Article 6”. Part I of Annex A provides that by 2020 the manufacture, import or export of the mercury-added products subject to paragraph 1 of Article 4 shall not be allowed (i.e., phase-out date). Part I of Annex A lists batteries; switches and relays; compact fluorescent lamps; linear fluorescent lamps; high-pressure mercury vapour lamps; cold cathode fluorescent lamps and external electrode fluorescent lamps; cosmetics; pesticides, biocides and topical antiseptics; non-electronic measuring devices with some conditions and exceptions. It should be noted that not only those wastes listed in Annex A but also all other mercury-added products are, upon becoming waste, categorized as mercury waste (wastes containing mercury or mercury compounds) under Article 11.



1. Article 4 of the Minamata Convention also addresses measures to phase-down the use of mercury dental amalgam. Paragraph 3 provides that “each Party shall take measures for the mercury-added products listed in Part II of Annex A in accordance with the provisions set out therein”. Part II of Annex A provides that “measures to be taken by a Party to phase down the use of dental amalgam shall take into account the Party’s domestic circumstances and relevant international guidance and shall include two or more” of the measures listed therein. Wastes of dental amalgam and its packaging are categorized as waste containing mercury under Article 11. In addition, in decision MC-3/2, Parties are encouraged to take an additional measure listed in Part II of Annex A.
2. Article 5 of the Minamata Convention addresses the use of mercury and mercury compounds in product manufacturing. Paragraph 2 provides that “each Party shall not allow the use of mercury or mercury compounds in the manufacturing processes listed in Part I of Annex B after the phase-out date specified in that Annex for the individual processes, except where the Party has a registered exemption pursuant to Article 6”. Part I of Annex B lists chlor-alkali production and acetaldehyde production in which mercury or mercury compounds are used as a cathode and a catalyst respectively. In addition, Paragraph 3 provides that “each Party shall take measures to restrict the use of mercury or mercury compounds in the processes listed in Part II of Annex B in accordance with the provisions set out therein”. Part II of Annex B lists vinyl chloride monomer production, sodium or potassium methylate or ethylate, and production of polyurethane using mercury containing catalysts. Measures to reduce or control mercury emissions and releases from manufacturing or production processes in which mercury or mercury compounds are used may result in the capture and generation of residues and substances contaminated with mercury or mercury compounds that should be managed appropriately as waste.
3. Article 7 of the Minamata Convention provides for the reduction, and where feasible elimination, of the use of mercury and mercury compounds in, and the emissions and releases to the environment of mercury from artisanal and small-scale gold mining (ASGM) and processing. Article 7 applies to ASGM and processing in which mercury amalgamation is used to extract gold from ore. Such activities may generate wastes consisting of or contaminated with mercury or mercury compounds such as tailings and extraction process residues.
4. Article 8 of the Minamata Convention addresses emissions of mercury or mercury compounds to the ambient air. Paragraph 3 provides that “[a] Party with relevant sources shall take measures to control emissions”. “Relevant source” means a source falling within one of the categories listed in Annex D to the Convention. Relevant sources listed in Annex D include waste incineration facilities, coal-fired power plants, coal-fired industrial boilers, smelting and roasting processes used in the production of non-ferrous metals, and cement clinker production facilities. Mercury emission control measures and practices may generate solid wastes contaminated with mercury or mercury compounds.
5. Article 9 of the Minamata Convention addresses mercury releases to land and water. Paragraph 3 requires that “Each Party shall, no later than three years after the date of entry into force of the Convention for it and on a regular basis thereafter, identify the relevant point source categories”. Relevant point source categories are those anthropogenic sources of mercury release considered significant by Parties, and which are not addressed in other provisions of the Convention. In decision MC-3/4, the Conference of the Parties of the Minamata Convention provided guidance for implementation of Article 9, noting that: “The obligation to ensure the environmentally sound management of waste set out under the Convention addresses significant releases to land and water;…” and that “While wastewater is addressed under article 9, Parties may additionally control wastewater under article 11 of the Convention;…”.
6. Lastly, Article 12 of the Minamata Convention provides that “each Party shall endeavour to develop appropriate strategies for identifying and assessing sites contaminated by mercury or mercury compounds” and that “the Conference of the Parties shall adopt guidance on managing contaminated sites”. Remediation activities for mercury-contaminated sites may generate mercury waste. This guidance[[19]](#footnote-27) was adopted in decision MC-3/6 and may be updated periodically.

2. UNEP Global Mercury Partnership

1. In decision 25/5, Part III, the UNEP Governing Council requested the Executive Director of UNEP, coordinating as appropriate with governments, intergovernmental organizations, stakeholders and the Global Mercury Partnership, to continue and enhance, as part of the international action on mercury, existing work in a number of areas. The Global Mercury Partnership currently has eight identified priorities for action or “partnership areas”[[20]](#footnote-29). One such partnership area is the mercury waste management partnership area, launched in 2008 with the Ministry of the Environment of Japan as lead. Among other things, the partnership area has identified and grouped country-level projects by waste stream and created a list of experts on mercury waste. Several relevant reports have been developed and are also available on the website including the 2015 *Practical sourcebook on mercury waste storage and disposal*[[21]](#footnote-30), and the 2019 *Catalogue of technologies and services on mercury waste management*[[22]](#footnote-31).

3. Heavy Metals Protocol

1. The objective of the 1998 Protocol on Heavy Metals to the 1979 Convention on Long-Range Transboundary Air Pollution, which was amended in 2012, is to control anthropogenic emissions of heavy metals, including mercury, that are subject to long-range transboundary atmospheric transport and are likely to have significant adverse human health or environmental effects. Parties to the Protocol are required to reduce emissions of target heavy metals below 1990 levels (or an alternative year between 1985 and 1995) by applying best available techniques for stationary sources and imposing emissions limit values for certain stationary sources. Parties are also required to develop and maintain emission inventories for heavy metals covered under the Protocol. Annex VII to the Protocol lists mercury-containing electrical components, measuring devices, fluorescent lamps, dental amalgam, pesticides, paint and batteries for recommended product management measures, including substitution, minimization, labelling, economic incentives, voluntary agreements, and programmes for collection, recycling or disposal.

4. SAICM (to be changed after ICCM5)

1. The Strategic Approach to International Chemicals Management (SAICM) is made up of a ministerial declaration (the “Dubai Declaration on International Chemicals Management”), an overarching policy strategy, and a global plan of action. Mercury is specifically addressed in the SAICM global plan of action, under work area 14, as follows: “Mercury and other chemicals of global concern; chemicals produced or used in high volumes; chemicals subject to wide dispersive uses; and other chemicals of concern at the national level”; specific activities under the work area pertain to risk reduction, the need for further action and the review of scientific information. A quick start programme (QSP) was established under SAICM in 2006 to support initial enabling capacity-building and implementation activities in developing countries, least developed countries, small island developing states and countries with economies in transition (UNEP, 2006a). As of February 2014, seven projects including mercury components such as campaign on minimization of mercury use and inventory development of mercury products, mercury releases, and mining sites have been implemented under the quick start programme (UNEP, 2014).

III. Guidance on environmentally sound management (ESM)

A. General considerations

1. ESM is a broad policy concept that is understood and implemented in various ways by different countries, organizations and other stakeholders. International guidance documents and core performance elements pertaining to ESM of hazardous wastes and other wastes have been developed by the Basel Convention and by the Organisation for Economic Co-operation and Development (OECD) and should help stakeholders to implement the ESM of hazardous wastes and other wastes as well as of mercury wastes as defined by the Minamata Convention.

1. Basel Convention

1. While provisions relevant to ESM of hazardous wastes or other wastes under the Basel Convention are listed in paragraph 15, the present guidelines are intended to provide a more precise definition of ESM when applied to mercury wastes, including by defining appropriate treatment and disposal methods for mercury waste streams that constitute ESM.
2. At its eleventh meeting in 2013, the Conference of the Parties to the Basel Convention adopted a *Framework for the environmentally sound management of hazardous wastes and other wastes*[[23]](#footnote-32). The framework establishes a common understanding of what ESM encompasses and identifies tools and strategies to support and promote the implementation of ESM. It is intended as a practical guide for governments and other stakeholders participating in the management of hazardous wastes and other wastes and constitutes the most comprehensive ESM guidance produced thus far, complementing the various technical guidelines adopted under the Basel Convention.
3. Under the Basel Convention, ESM is the subject of multiple provisions (refer to section II A.1 above) and the following two declarations:

(a) The 1999 Basel Declaration on Environmentally Sound Management, which was adopted by the Conference of the Parties to the Basel Convention at its fifth meeting and calls on Parties to enhance and strengthen their efforts and cooperation to achieve ESM, including through prevention, minimization, recycling, recovery and disposal of hazardous and other wastes subject to the Basel Convention, taking into account social, technological and economic concerns, and through further reduction of transboundary movements of hazardous and other wastes subject to the Convention;

(b) The 2011 Cartagena Declaration on the Prevention, Minimization and Recovery of Hazardous Wastes and Other Wastes, which was adopted by the Conference of the Parties to the Basel Convention at its tenth meeting and reaffirms that the Basel Convention is the primary global legal instrument for guiding the ESM of hazardous wastes and other wastes and their disposal.

1. ESM criteria recommendations for computing equipment have been developed under the Basel Convention Partnership for Action on Computing Equipment (PACE).

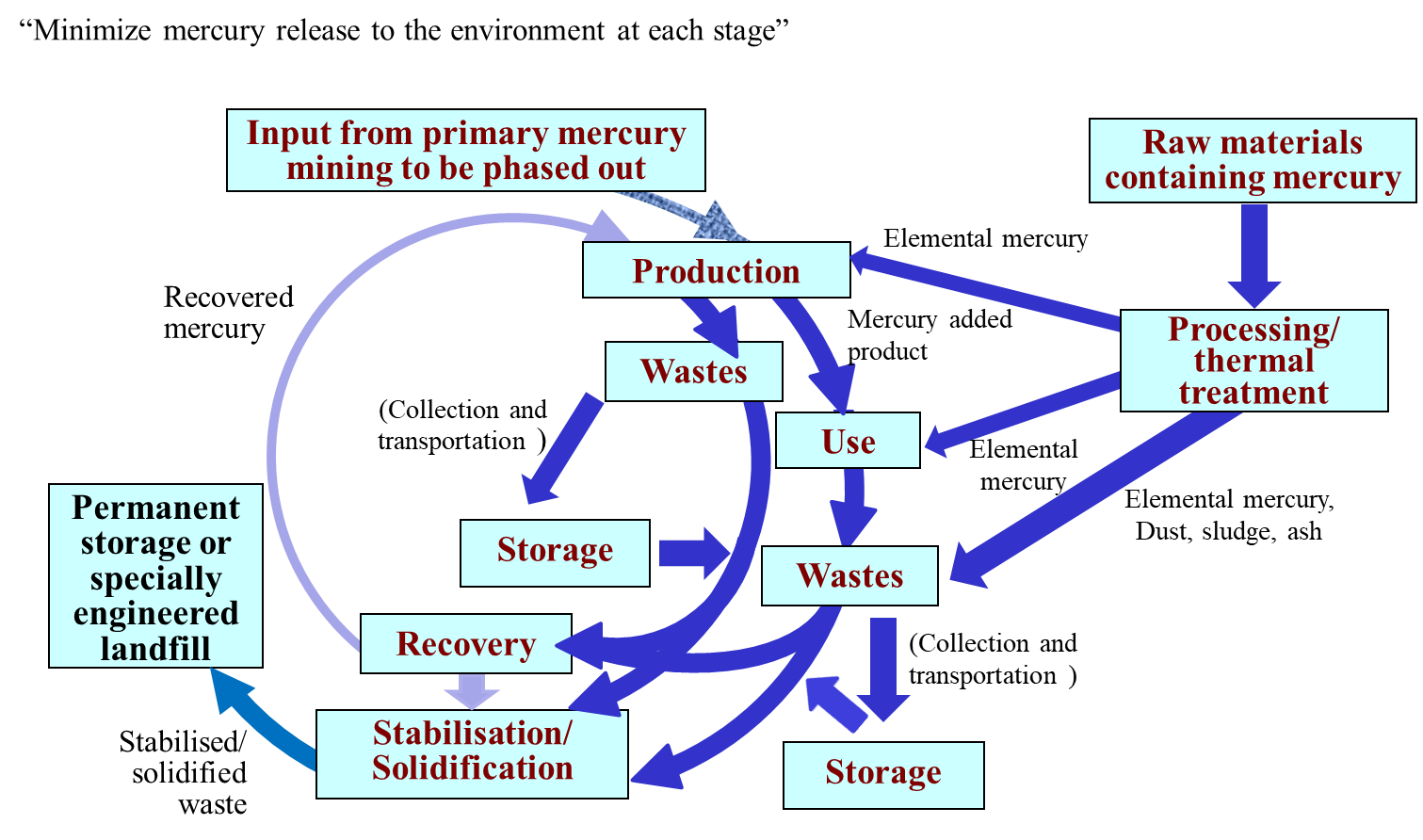
2. Organisation for Economic Co-operation and Development

1. OECD has adopted a recommendation on ESM of waste that covers such items as the core performance elements of ESM guidelines applying to waste recovery facilities, including: elements of performance that precede collection, transport, treatment and storage; and elements subsequent to storage, transport, treatment and disposal of pertinent residues (OECD 2004).
2. Further information may be found in the guidance manual for the implementation of the OECD recommendation on ESM of waste (OECD 2007).

3. Life cycle management of mercury

1. The concept of life cycle management can serve as a useful approach to promote the ESM of mercury wastes. Life cycle management provides a framework for analysing and managing the sustainability performance of goods and services. Global businesses are using it, for instance, to reduce the carbon, raw material and water footprints of their products, improve their social and economic performance, and make value chains more sustainable (UNEP and SETAC, 2009). When a life cycle management approach is applied to mercury, performance should be assessed during the following stages: production of products containing mercury or mercury compounds or of other products using mercury; use of such products; collection and transportation of wastes; and disposal of wastes.
2. In life cycle management of mercury, it is important to give priority to reducing the use of mercury in products and industrial processes, thereby reducing the mercury content of wastes resulting from such products and processes. When using products containing mercury or mercury compounds, special care should be taken not to emit or release mercury into the environment. Mercury wastes should be treated to either recover the mercury in them or to immobilize it in an environmentally sound manner. In those cases where mercury is recovered, the recovered mercury should be disposed of, after stabilization and solidification (S/S), at a permanent storage site or at a specially engineered landfill site. Alternatively, for non-Parties to the Minamata Convention, the recovered mercury may, for example, be used as an input in products for which mercury-free alternatives do not exist or are unavailable, or in cases in which it would take a long time to replace products containing mercury or mercury compounds; such reuse could help to reduce the production of new mercury from primary mining. For Parties to the Minamata Convention, the recovered mercury may be used as an input for products or processes that are still allowed under, and respecting the relevant conditions, of the Minamata Convention. Mercury wastes may be stored pending further treatment or disposal or until export to other countries for disposal is possible (see figure 1 below).

**Figure 1:** Basic concept of mercury management



1. Waste management covers source separation, collection, transportation, storage and disposal (e.g., recovery, solidification, stabilization, permanent storage and disposal in a specially engineered landfill). When a government plans to collect mercury wastes, it also needs to plan for the subsequent waste management steps, such as storage and disposal.

B. Legislative and regulatory framework

1. Parties to the Basel Convention should examine their national controls, standards and procedures to ensure that they are in agreement with their obligations under the Convention, including those pertaining to the transboundary movement and ESM of mercury wastes; Parties to the Minamata Convention should also examine such controls, standards and procedures to ensure that they are in agreement with their waste-related obligations under that Convention.
2. Implementing legislation should give governments the power to enact and enforce specific rules and regulations, conduct inspections and establish penalties for violations. Legislation on hazardous wastes should define hazardous wastes in accordance with the Basel Convention, and legislation on mercury wastes should define mercury wastes in accordance with the Minamata Convention. The legislation could define ESM and require adherence to ESM principles, thus ensuring that countries comply with the provisions on ESM of mercury wastes. The specific components and features of a regulatory framework that would meet the requirements of the Basel Convention and other international agreements are discussed below[[24]](#footnote-34).

1. Registration of waste generators

1. One approach to facilitate ESM of mercury waste involves the establishment, through regulation, of registers of generators of this type of waste. Such registers should include large-scale mercury waste generators such as power plants, industrial establishments (e.g., chlor-alkali plants using mercury cell technology, VCM production facilities using a mercury catalyst or smelting operations), as well as hospitals, medical and dental clinics, research institutes, mercury waste collectors, etc. A register of mercury waste generators would make it possible to determine the origins, types and volumes of various kinds of mercury waste, as well as the quantities of products containing mercury or mercury compounds being used by various waste generators.
2. Regulations on registries of generators of mercury wastes could require waste generators to provide their name, address, the name of the responsible person, their type of business, the amounts and kinds of mercury waste generated, and information on collection schemes applicable to such wastes and how the wastes are to be handed over to collectors and disposed of. Waste generators could be required to transmit and provide regular updates on this information to the authorities (central or local governments). Based on the amounts and kinds of waste obtained through registries, Parties could also develop waste inventory programmes.
3. Mercury waste generators should have a duty to prevent mercury emissions and releases to the environment until the wastes are handed over to collectors or sent to a disposal facility. They should strictly comply with national and local legal requirements regarding the management of mercury wastes and be held liable for remediating or compensating any environmental or health damages that they might cause when handling such wastes, to the degree required by applicable legislation.

2. Reduction and phase-out of mercury use in ASGM, products and industrial processes

1. The reduction and phase-out of mercury use in ASGM, products and industrial processes is one of the most effective ways of reducing emissions and releases of mercury to the environment.
2. According to Article 7 of the Minamata Convention, Parties to that convention that have ASGM and processing within their territories are required to take steps to reduce, and where feasible eliminate, the use of mercury and mercury compounds in, and the emissions and releases to the environment of mercury from, such mining and processing. In addition, a party to the Minamata Convention which determines that ASGM and processing in its territory is more than insignificant is required to develop and implement a national action plan and review its progress every three years.
3. Articles 4 and 5 of the Minamata Convention require Parties to phase-down or phase-out specific mercury-added products and use of mercury in specific industrial processes (see paragraphs 28-30 above).
4. Examples of regional/national provisions on reduction and phase-out of mercury use are shown in boxes 1 to 3.

|  |
| --- |
| **Box 1 Regional Example: European Union**  Directive 2002/95/EC of the European Parliament and of the Council of 27 January 2003on the restriction of the use of certain hazardous substances in electrical and electronic equipment, known as the “RoHS Directive”, is an example of a regulation phasing out the use of mercury in certain products. The RoHS Directive restricts the use of mercury and other substances in electrical and electronic equipment and, while exemptions may be granted in cases where no satisfactory alternatives are available (e.g., some types of lamps containing mercury), most mercury‑containing electrical and electronic equipment has been phased out in the European Union market since the Directive entered into force on 1 July 2006. A revised version of the RoHS Directive entered into force on 21 July 2011 (European Union, 2011a).  Another example from the European Union is Directive 2006/66/EC of the European Parliament and of the Council on batteries and accumulators and waste batteries and accumulators, which prohibits the placing on the market of all batteries, whether or not incorporated into appliances, that contain more than 0.0005 per cent of mercury by weight. Batteries and accumulators lawfully placed on the market prior to the date of application of the prohibitions set out in Article 4 of the Directive may continue to be marketed until stocks run out (European Union, 2006, European Commission, 2014a). |

|  |
| --- |
| **Box 2 Country Example: Canada**  Canada’s *Products Containing Mercury Regulations* prohibit the manufacture and import of products containing mercury or any of its compounds, with some exemptions for essential products for which there are no technically or economically viable alternatives (Government of Canada, 2014). |

|  |
| --- |
| **Box 3 Country Example: Norway**  Norway imposes a general ban on the use of mercury in products to ensure that mercury is not used in products where alternatives exist[[25]](#footnote-36). The regulationprohibits the manufacture, import, export, sale and use of substances or preparations that contain mercury or mercury compounds, and the manufacture, import, export and sale of solid processed products containing mercury or mercury compounds. The regulation therefore is expected to reduce the number of products containing mercury or mercury compounds on the market, in addition to discharges from products that have inadvertently failed to be disposed of as hazardous waste (Government of Norway, 2011). |

3. Transboundary movement requirements

1. Under the Basel Convention, mercury wastes subject to transboundary movement that belong to any category contained in Annex I (see Table 1) are considered hazardous waste unless, through national tests, they can be shown not to exhibit any of the characteristics listed in Annex III (list of hazardous characteristics).
2. If a party to the Convention has national legislation of banning the import of mercury wastes and has reported the information in accordance with Article 4, paragraph 1 (a), other Parties shall prohibit or shall not permit the export of such wastes to that party. In addition, in cases where the State of import has not prohibited the import of mercury wastes, Parties to the Convention shall prohibit or shall not permit the export of such wastes if the State of import does not consent in writing to the specific import.
3. Parties listed in Annex VII to the Convention (members of the European Union, OECD and Liechtenstein), that are bound by the Ban Amendment, shall prohibit transboundary movements to States not listed in Annex VII of hazardous wastes which are destined for operations according to Annex IV.A and hazardous wastes under Article 1.1(a) which are destined to operations according to Annex IV.B[[26]](#footnote-37).
4. The Minamata Convention also includes a provision on transboundary movements of mercury waste in Article 11, paragraph 3 (c) (see paragraph 24 above).
5. Transboundary movements of hazardous wastes and other wastes must be kept to a minimum consistent with their environmentally sound and efficient management and conducted in a manner that protects human health and the environment from any adverse effects that may result from such movements. Transboundary movements of these wastes are permitted only under the following conditions:
   1. If the country of export does not have the technical capacity and the necessary facilities, capacity or suitable disposal sites in order to dispose of the wastes in question in an environmentally sound and efficient manner;
   2. If the wastes in question are required as a raw material for recycling or recovery industries in the country of import; or
   3. If the transboundary movements in question are in accordance with other criteria decided by the Parties.
6. Any transboundary movements of hazardous and other wastes are to be notified in writing to the competent authorities of all countries concerned by the movement (country of export, country of import and, if applicable, country of transit). This notification is to contain the declarations and information requested in the Convention and shall be written in a language acceptable by the State of import. Prior written consent from the importing and the exporting country and, if appropriate, from transit countries, in addition to a confirmation of the existence of a contract specifying ESM of the wastes between the exporter and the owner of the disposal facility are required before any transboundary movements of hazardous and other wastes can take place. Parties are to prohibit the export of hazardous wastes and other wastes if the country of import prohibits the import of such wastes. The Convention also requires that information regarding any consignment be accompanied by a movement document from the point where the transboundary movement commences to the point of disposal.
7. Hazardous wastes and other wastes subject to transboundary movements should be packaged, labelled and transported in conformity with international rules and standards[[27]](#footnote-39).
8. When a transboundary movement of hazardous or other wastes to which the consent of the countries concerned has been given cannot be completed in accordance with the terms of the contract, the country of export must ensure that the wastes in question are taken back into the country of export by the exporter if alternative arrangements cannot be made for their disposal in an environmentally sound manner. This is to be done within 90 days of the importing country’s notification to the country of export and the secretariat, or within another period of time on which the countries involved agree (Article 8). In the case of illegal traffic (as defined in Article 9, paragraph 1), the country of export shall ensure that the wastes in question are returned to the country of export for disposal or are disposed of in accordance with the provisions of the Convention.
9. When required by the country of import or any transit country that is a party to the Convention, transboundary movements of hazardous or other wastes are to be covered by insurance, bonds or other guarantees.
10. No transboundary movements of hazardous wastes or other wastes are permitted between a party and a non-party to the Convention unless a bilateral, multilateral or regional arrangement exists, as required under Article 11 of the Convention. Existing bilateral and multilateral agreements that have been notified to the Secretariat are listed on the Basel Convention webpage[[28]](#footnote-41).

4. Authorization and inspection of disposal facilities

1. Mercury wastes should be disposed of in facilities that practise ESM.
2. Most countries have legislation or sector-specific regulations that require waste disposal facilities to obtain approvals or operating permits to commence their operations. Approvals or operating permits may include specific conditions (e.g., facility design and operating conditions) that must be maintained in order for the approval or permit to remain valid. It may be necessary to add requirements specific to mercury waste to meet the requirements of ESM, to comply with specific requirements of the Basel Convention and to take into account recommendations and guidelines on best available techniques (BAT), such as guidelines on BAT and provisional guidance on best environmental practices produced under the Stockholm Convention, reference documents on BAT produced by the European Union (known as “BREFs”), and guidelines for the chlor-alkali sector (Euro Chlor, 2004). Approvals or operating permits should be reviewed periodically and, if necessary, should be updated in order to improve occupational and environmental safety through the use of improved or new technologies.
3. Disposal facilities should be periodically inspected by independent authorities or technical inspection associations in order to verify their compliance with the requirements set out in their permits. Legislation should enable authorities to conduct extraordinary inspections if there is evidence of non‑compliance with permit requirements by disposal facilities.

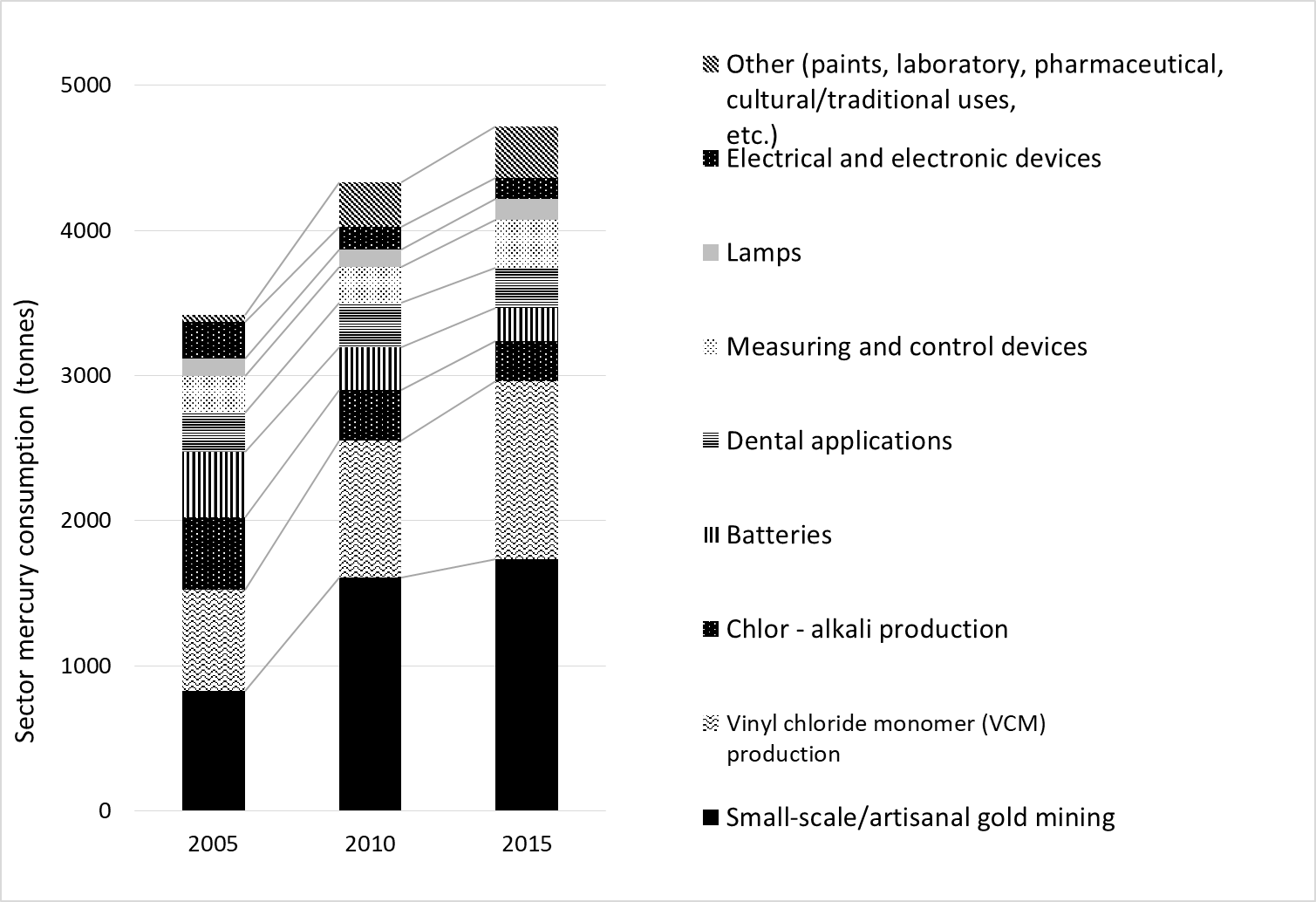
C. Identification and inventory

1. To enable effective action to prevent, minimize and manage mercury wastes, it is important that Parties identify the sources of mercury waste generation and quantify the amount of mercury wastes generated and the mercury concentrations in such wastes.

1. Identification of sources of mercury wastes

1. While the sources of mercury waste generation vary across countries, globally the major sources of mercury waste include industrial processes using mercury or mercury compounds and mercury devices that have become waste. Other major sources of mercury waste are the production, processing and use of natural resources (e.g., non-ferrous metals ore processing, oil and gas processing and coal combustion). Figure 2 shows trends of estimated global mercury use by application from 2005 to 2015. The largest use sector was artisanal and small-scale gold mining (AGSM), followed by vinyl chloride monomer VCM/polyvinyl chloride (PVC) production and   
   chlor-alkali production up to 2010. Mercury is also used in products such as batteries, dental amalgams, measuring devices, lamps, and electrical and electronic devices, although the amounts of mercury used in these categories varies among countries. The total amount of mercury used globally in 2015 was 4,715 tonnes, which was increased from 2005 due to the increase of mercury use in ASGM, VCM production, measuring devices, lamps and other uses. Meanwhile mercury use decreased in chlor-alkali production, batteries and electrical and electronic devices (UNEP, 2017).

**Figure 2:** Trends of estimated global mercury uses (UNEP, 2017)

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1. The sources, categories and examples of mercury wastes are summarized in table 5 below.
2. It should be noted that in some countries some of the industrial sources presented in table 5 (sources 1, 2, 3, 4 and 7, except for production processes using mercury) neither use mercury nor generate mercury wastes. Industrial processes depend on a country’s technological and social conditions, and these will determine whether mercury-free processes are used. It should be noted that when disposal operations (e.g., “8. Waste treatment other than incineration” or “9. Waste deposition/landfilling and wastewater treatment”) are included in an inventory of mercury waste, this may result in an overestimation, as mercury waste generated in or mercury releases from such facilities may result from mercury contained in waste being treated.

**Table 5:** Sources, categories and examples of mercury wastes (UNEP 2002; 2005; 2006b; 2006c, decision MC-3/5).

| **Source** | **Cate­gories\*** | **Examples of waste types generated** | **Remarks** |
| --- | --- | --- | --- |
| 1. **Extraction and use of fuels/energy sources** | | | |
| * 1. Coal combustion in power plants | C | C: Flue gas cleaning residues (fly ash, particulate matter, wastewater / sludge, etc.), bottom ash, sludge, catalyst, process water, contaminated personal protective equipment, equipment wash down waste | * Accumulation in bottom ashes and flue gas cleaning residues |
| * 1. Other coal combustion | C |
| * 1. Extraction and use of other fossil fuels | C |
| * 1. Biomass fired power and heat generation | C |
| * 1. Extraction, refining and use of mineral oil | A/C | A: Mercury captured from crude oil and natural gas processing  C: Flue gas cleaning residues (fly ash, particulate matter, wastewater / sludge, etc.), bottom ash, sludge, catalyst, process water, contaminated personal protective equipment, equipment wash down waste | * Accumulation in bottom ashes and flue gas cleaning residues; * Sludge from separator tanks and sedimentary sand tanks for refining of crude oil; * Sludge from natural gas production and processing; * Catalyst for refining of crude oil and processing of natural gas. |
| * 1. Extraction, refining and use of natural gas | A/C |  |  |
| 1. **Primary (virgin) metal production** | | | |
| * 1. Primary extraction and processing of mercury | A  C | Cinnabar concentrate  Smelting residue, tailings | * Pyrometallurgy of mercury concentrate |
| * 1. Metal (aluminium, copper, gold, lead, manganese, zinc, primary ferrous metal, other non-ferrous metals) initial processing of ores | C | Tailings | * Crushing, grinding, flotation and dewatering of ore to produce concentrate |
| * 1. Metal (aluminium, copper, gold, lead, manganese, zinc, primary ferrous metal, other non-ferrous metals) smelting/roasting processes | A  C | Mercury chloride (I) and mercury chloride (II) recovered from roasting and smelting processes  Smelting/roasting process residues, flue gas cleaning residues, wastewater treatment residues | * Thermal treatment of concentrate * Thermal treatment of concentrate |
| 1. **Production processes with mercury impurities** | | | |
| * 1. Cement production | C | Process residues, flue gas cleaning residues, sludge | * Pyroprocessing of raw materials and fuels with naturally occurring mercury impurities |
| * 1. Pulp and paper production | * Combustion of raw materials with naturally occurring mercury impurities |
| * 1. Lime production and lightweight aggregate kilns | * Calcination of raw materials and fuels with naturally occurring mercury impurities |
| 1. **Intentional use of mercury in industrial production** | | | |
| * 1. Chlor-alkali production with mercury-technology | A  C | Waste electrodes, stockpiles of mercury  Solid waste contaminated with mercury, process residues, soil | * Mercury cell * Mercury recovery units (retort). |
| * 1. Production of alcoholates (e.g. sodium or potassium methylate or ethylate), dithionite and ultrapure potassium hydroxide solution | A  C | Waste electrodes, stockpiles of mercury  Solid waste contaminated with mercury, process residues, soil | * Mercury cell; * Mercury recovery units (retort). |
| * 1. VCM production with mercuric chloride (HgCl2) catalyst | A  C | Waste catalyst, stockpiles of mercury catalyst  Process residues | * Mercury catalyst process |
| * 1. Acetaldehyde production with mercury sulphate (HgSO4) catalyst | A  C | Waste catalysts, stockpiles of mercury catalyst  Wastewater | * Mercury catalyst process |
| * 1. Polyurethane production with mercury catalyst | A  C | Waste catalysts, stockpiles of mercury catalyst  Waste catalysed materials | * Mercury catalyst process |
| * 1. Other production of chemicals and pharmaceuticals with mercury compounds and/or catalysts | A  C | Waste catalysts, stockpiles of mercury catalyst  Process residues, wastewater | * Mercury catalyst process |
| * 1. Production of products referred to in 5. below | A  C | Stockpiles of mercury and mercury compounds not used for production  Process residues, wastewater |  |
| 1. **Products and applications with intentional use of mercury** | | | |
| * 1. Thermometers and other measuring devices with mercury | B1 | Used, obsolete or broken products | * Mercury |
| * 1. Electrical and electronic switches, contacts and relays with mercury |
| * 1. Light sources with mercury | B1 | * Vapour-phase mercury; * Divalent mercury adsorbed on phosphor powder. |
| * 1. Batteries containing mercury | B2 | * Mercury, mercury oxide |
| * 1. Biocides and pesticides | B1 | Stockpiles of obsolete pesticides, soil and solid waste contaminated with mercury | * Mercury compounds (mainly ethylmercury chloride) |
| * 1. Paints | B1  C | Stockpiles of obsolete paints  Solid waste contaminated with mercury, wastewater treatment residues | * Phenylmercuric acetate and similar mercury compounds |
| * 1. Pharmaceuticals for human and veterinary uses | B1  C | Stockpiles of obsolete pharmaceuticals  Medical waste, wastewater treatment residues from agricultural facilities | * Thimerosal; * Mercuric chloride; * Phenyl mercuric nitrate; * Mercurochrome, etc. |
| * 1. Cosmetics and related products | B2 | Stockpiles of cosmetics and related products | * Mercury iodide; * Ammoniated mercury, etc. |
| * 1. Dental amalgam fillings | B2  C | Stockpiles of dental amalgam  Amalgam separator residues and wastewater treatment residues from dental clinics | * Alloys of mercury, silver, copper and tin |
| * 1. Scientific instrument used for the calibration of medical or scientific devices | B1 | Used, obsolete or broken products | * Mercury |
| * 1. Laboratory chemicals and equipment | A  B1/B2  C | Stockpiles of laboratory chemicals  Stockpiles of laboratory equipment  Wastewater treatment residues, laboratory wastes | * Mercury; * Mercury chloride, etc. |
| * 1. Polyurethane elastomers | B2  C | Used or end-of-life products  Defective and excess product waste | * Elastomer waste containing mercury compounds |
| * 1. Sponge gold/gold production from ASGM sources | C | Tailings | * Thermal treatment of gold; * Industrial processing. |
| * 1. Mercury metal use in religious rituals and folklore medicine | A  C | Mercury used for religious rituals  Solid waste | * Mercury |
| * 1. Mercury bearings in lighthouses | A | Waste mercury bearings, stockpiles | * Mercury |
| * 1. Calomel electrode | B1 | Used or obsolete products | * Mercury; * Mercury chloride. |
| * 1. Miscellaneous product uses, mercury metal uses and other sources | B1/B2  C | Stockpiles  Wastewater treatment residues, solid wastes, demolition waste | * Infra-red detection semiconductors with mercury; * Bougie and Cantor tubes; * Educational uses, etc. |
| 1. **Secondary metal production** | | | |
| * 1. Recovery of mercury | A  B1/B2  C | Spillage during recycling processes, stockpiles of recovered mercury  Wastes of products containing mercury or mercury compounds  Extraction process residues , flue gas cleaning residues, wastewater treatment residues, contaminated environmental media | * Treatment of waste from chlor-alkali, alcoholates, dithionite and ultrapure potassium hydroxide solution production with mercury technology including decommissioning facilities; * Treatment of waste from polyurethane, vinyl chloride monomer, acetaldehyde production using a mercury containing catalyst * Recovery from mercury meters used in natural gas pipelines; * Recovery from manometers, thermometers, and other equipment. |
| * 1. Recovery of ferrous metals | C | * Shredding of vehicles containing mercury switches; * Smelting of materials containing mercury. |
| * 1. Recovery of gold from e-waste (printed circuit boards) | A/C | * Mercury used for amalgamation of gold in small-scale, informal recovery process; * Thermal process. |
| * 1. Recovery of other metals, such as copper and aluminium | C | * Other materials or products /components containing mercury or mercury compounds |
| 1. **Waste incineration** | | | |
| 7.1. Incineration of municipal solid waste | C | Flue gas cleaning residues, wastewater treatment residues, bottom ash, fly ash | * Wastes of products containing mercury or mercury compounds and process waste; * Natural mercury impurities in high volume materials (plastics, paper, etc.) and minerals. |
| 7.2. Incineration of hazardous waste |
| 7.3. Incineration of medical waste |
| 7.4. Sewage sludge incineration |
| 1. **Waste treatment other than incineration** | | | |
| * 1. Physico-chemical treatment of mercury wastes | C | Sludge, wastewater treatment residues | * Chemical precipitation and chemical oxidation of wastes contaminated with mercury or mercury compounds |
| * 1. Thermal treatment of mercury wastes | C | Flue gas cleaning residues |  |
| * 1. Other treatment of mercury wastes | C | Wastewater treatment residues, materials contaminated with mercury or mercury compounds | * Crushing and separation of products containing mercury or mercury compounds |
| 1. **Waste deposition/landfilling and wastewater treatment** | | | |
| * 1. Controlled landfills/deposits | C | Wastewater, wastewater treatment residues, solid waste contaminated with mercury | * Results from disposal of products containing mercury or mercury compounds and process waste;   Natural mercury impurities in bulk materials (plastics, tin cans, etc.) and minerals. |
| * 1. Diffuse deposition under some control |
| * 1. Uncontrolled local disposal of industrial production waste |
| * 1. Uncontrolled dumping of general waste |
| * 1. Wastewater system/treatment | Wastewater treatment residues, slurries | * Intentionally used mercury in spent products and process waste; * Mercury as an anthropogenic trace pollutant in bulk materials. |
| 1. **Crematoria and cemeteries** | | | |
| * 1. Crematoria | C | Flue gas cleaning residues, wastewater treatment residues | * Dental amalgam fillings |
| * 1. Cemeteries | Soil contaminated with mercury |
| 1. **Remediation of contaminated sites** | | | |
| * 1. Remediation of contaminated sites | C | Rubble, debris and soil contaminated with mercury or mercury compounds that has been excavated or removed as part of remediation activities. |  |

\*A: Waste consisting of mercury or mercury compounds; B: Wastes containing mercury or mercury compounds; C: Wastes contaminated with mercury or mercury compounds.

1. More detailed information on products containing mercury or mercury compounds (e.g., names and manufacturers of specific products) is available from the following sources:
   1. UNEP, 2008b. *Report on the major mercury-containing products and processes, their substitutes and experience in switching to mercury-free products and processes*;
   2. European Commission, 2008. *Options for reducing mercury use in products and applications, and the fate of mercury already circulating in society*;
   3. UNEP Global Mercury Partnership – Mercury-Containing Products Partnership Area[[29]](#footnote-43);
   4. UNEP, 2016. *Lessons from Countries Phasing Down Dental Amalgam Use*;
   5. Lowell Center for Sustainable Production, 2003. *An Investigation of Alternatives to Mercury-Containing Products*; and
   6. The Interstate Mercury Education and Reduction Clearinghouse (IMERC). Mercury‑Added Products Database[[30]](#footnote-44).

2. Inventories

1. Inventories are an important tool for identifying, quantifying and characterizing wastes. National inventories may be used:
   1. To establish a baseline for quantities of products containing mercury or mercury compounds produced, circulated, traded or in use, commodity mercury, mercury-containing by-products and mercury wastes;
   2. To establish an information registry to assist with safety and regulatory inspections;
   3. To obtain the accurate information needed to draw up plans for lifecycle management of mercury;
   4. To assist in estimating the quantities of mercury and mercury compounds that require storage;
   5. To assist with the preparation of emergency response plans;
   6. To track progress towards reducing and phasing out mercury in products and processes.
2. After identifying the sources and types of mercury wastes, process-specific information and quantities should be used to estimate the amounts of waste from the identified sources for different types of waste in a given country (or area, community, etc.) (UNEP, 2005).
3. It is in some cases very difficult to collect the necessary data to estimate the quantities of mercury waste generated, particularly in developing countries and countries with economies in transition, due to a lack of data, particularly where small-scale facilities are concerned. In cases where actual measurements are not feasible, data collection could be carried out using questionnaire-based surveys.
4. The *Methodological guide for the undertaking of national inventories of hazardous wastes under the Basel Convention* (UNEP, 2015a) could be used when producing inventories of mercury wastes.
5. The *Toolkit for Identification and Quantification of Mercury Releases* (UNEP, 2019b) can also be applied. There are two versions of the toolkit, which correspond to two levels of inventory development (i.e., simplified and comprehensive). The toolkit is intended to help countries to develop national inventories of mercury releases and provides a standardized methodology and accompanying database enabling the development of consistent national and regional mercury inventories. It should be noted that the default factors suggested in this Toolkit are based on a limited data base and as such, they should be considered subject to revisions as the data base grows. Therefore, it may be appropriate to review, and confirm to the extent feasible, main source specific data for local/national conditions, before major decisions are taken on implementation of mitigation initiatives (UNEP, 2019b). The toolkit has been applied in a number of countries (UNEP, 2008c) and GEF-funded projects. Online training is available from the platform MercuryLearn[[31]](#footnote-46) developed by UNEP and UNITAR. The publication by UNIDO (2018), *Waste mercury perspective[[32]](#footnote-47)*, could also be considered.
6. In keeping with a lifecycle management approach, channels or pathways through which mercury in wastes may be released into the environment should also be identified. In view of the potential risks of mercury releases into the environment, the various types of mercury waste should be ranked in order of priority for action. Information about possible measures to mitigate releases should then be collected, especially with regard to sources and types of mercury waste involving large quantities of mercury and presenting higher risk of mercury release into the environment. Measures must then be evaluated in terms of the magnitude of environmental mercury releases that they could help to prevent, their administrative and social costs, the availability of techniques and facilities associated with them, their social acceptability, etc.
7. In some countries, a Pollutant Release and Transfer Registry (PRTR) is used to collect data about specific mercury content in wastes and its transferby each facility (Kuncova et al., 2007). PRTR data are publicly available but often not available for recent years[[33]](#footnote-49).

D. Sampling, analysis and monitoring

1. Sampling, analysis and monitoring are critical components of mercury waste management. Waste sampling, analysis and monitoring should be conducted by trained professionals in accordance with well-designed programmes using internationally accepted or nationally approved methods and should be carried out using the same methods throughout the lives of such programmes. They should also be subjected to rigorous quality assurance and quality control measures. Mistakes in sampling, analysis or monitoring or deviation from standard operational procedures can result in meaningless data or even programme-damaging data. Each party, as appropriate, should therefore develop standards to ensure that training, protocols and laboratory capabilities are in place for sampling, monitoring and analytical methods and that those standards are enforced.
2. Because there are numerous reasons for sampling, analysing and monitoring and because waste comes in so many different physical forms, many different sampling, analysis and monitoring methods are available. Although it is beyond the scope of this document to discuss them specifically, the next three sections consider key elements that should be included in sampling, analysis and monitoring activities. Waste testing priorities should be set based on existing knowledge (or lack thereof) of the mercury content of different types of waste (e.g., testing waste mercury lamps is unlikely to be a high priority, as significant information about their mercury content is readily available).
3. For information on good laboratory practices, the OECD series on good laboratory practice (OECD, various years) may be consulted; on general methodological considerations, the *Guidance for identifying populations at risk from mercury exposure* (UNEP and WHO, 2008)contains helpful information and may be used. .

1. Sampling

1. The overall objective of any sampling activity is to obtain a sample that can be used for a targeted purpose, e.g., site characterization, compliance with regulatory standards or determination of the suitability of proposed treatment or disposal methods. This objective should be identified before sampling is started. It is indispensable that quality requirements for equipment, transportation and traceability be met.
2. Standardized sampling procedures should be established and agreed upon before the start of the sampling campaign (both matrix- and mercury-specific). Elements of these procedures include the following:
   1. The number of samples to be taken, the sampling frequency, the duration of the sampling project and a description of the sampling method to be used (including quality assurance procedures put in place, e.g., use of appropriate sampling containers[[34]](#footnote-53) and field blanks and of chain-of-custody procedures);
   2. Selection of locations or sites at which mercury wastes are generated and time and date of sample-taking (including description and geographic localization);
   3. Identity of person who took the sample and conditions during sampling;
   4. Full description of sample characteristics – labelling;
   5. Preservation of the integrity of samples during transport and storage (before analysis);
   6. Close cooperation between the sampler and the analytical laboratory;
   7. Appropriately trained sampling personnel.
3. Sampling should comply with specific national legislation, where it exists, or with international regulations and standards. In countries where regulations do not exist, qualified staff should be appointed. Sampling procedures include the following:
   1. Development of a standard operational procedure (SOP) for sampling each of the matrices for subsequent mercury analysis;
   2. Application of well-established sampling procedures such as those developed by the International Organization for Standardization (ISO), the European Committee for Standardization (CEN), the United States Environmental Protection Agency (USEPA), the Global Environment Monitoring System (GEMS) and the American Society for Testing and Materials (ASTM);
   3. Establishment of quality assurance and quality control (QA/QC) procedures.
4. All these steps should be followed if sampling programmes are to be successful. Similarly, documentation should be thorough and rigorous.
5. Mercury can occur and be sampled in liquids, solids, gases and biota:
   1. Liquids:
      1. Leachate from dumpsites and landfills;
      2. Liquid collected from spills;
      3. Water (surface water, drinking water and industrial effluents);
   2. Solids:
      1. Stockpiles of products and formulations consisting of, containing or contaminated with mercury or mercury compounds;
      2. Solids from industrial sources and treatment or disposal processes (fly ash, bottom ash, sludge, still bottoms, other residues, clothing, etc.);
      3. Containers, equipment or other packaging materials (rinse or wipe samples), including the tissues or fabric used in the collection of wipe samples;
      4. Soil, sediment, rubble, sewage sludge and compost;
   3. Gases:
      1. Air (indoor) of facilities handling mercury wastes;
      2. Mercury releases to the air from treatment of mercury wastes;
      3. Flue gas from waste incinerators;

(d) Biota:

* + 1. Biological materials (blood, urine and hair, especially those obtained through worker health monitoring);
    2. Plants and animals.

1. In environmental and human monitoring programmes, both biotic and abiotic matrices may be included:
   1. Plant materials and food;
   2. Human hair, urine, nails, breast milk or blood;
   3. Air (ambient, wet or dry deposition or, possibly, snow).

2. Analysis

1. Analysis refers to the extraction, purification, separation, identification, quantification and reporting of mercury concentrations in the matrix of interest. In order to obtain meaningful and acceptable results, the analytical laboratory should have the necessary infrastructure (housing) and proven experience with the matrix and the mercury species (e.g., successful participation in   
   inter-laboratory comparison studies and in external proficiency testing schemes).
2. Accreditation of the laboratory in accordance with ISO 17025 or other standards by an independent body is important. Essential criteria for obtaining high-quality results include:
   1. Specification of the analytical technique used;
   2. Maintenance of analytical equipment;
   3. Validation of all methods used (including in-house methods);
   4. Training of laboratory staff.
3. Mercury analysis is typically performed in a dedicated laboratory. For screening purposes, test kits are available and can be used in the field.
4. For the analysis of mercury, there is no single analytical method available. Methods of analysing the various matrices for mercury, either for total mercury content or speciation of mercury, have been developed by the International Organization for Standardization (ISO), the European Committee for Standardization (CEN) at the international level, and by the Ministry of Ecology and Environment (MEE) of China, the Japan Standards Association and USEPA at the national level. Table 6 lists some examples of methods for analysing mercury in wastes, flue gas, wastewater and soil. Most in-house methods are variations of these. As with all chemical analysis, laboratories should use only validated methods, and performance should be evaluated through QA/QC programmes.

**Table 6:** Chemical analysis of mercury in waste, flue gas, wastewater and soil

| **Matrix or waste medium** | | **Purpose** | **Method** |
| --- | --- | --- | --- |
| Waste | | To determine the mobility of mercury in waste | EN 12457-1 to 4: Characterization of waste - Leaching - Compliance test for leaching of granular waste materials and sludges (CEN, 2002a) |
|  | |  | EN 12920: Characterization of waste - Methodology for the determination of the leaching behaviour of waste under specified conditions (CEN, 2006) |
|  | |  | EN 13656: Soil, treated biowaste, sludge and waste - Digestion with a hydrochloric (HCl), nitric (HNO3) and tetrafluoroboric (HBF4) or hydrofluoric (HF) acid mixture for subsequent determination of elements (CEN, 2020) |
|  | |  | EN 14405: Characterization of waste - Leaching behaviour test - Up-flow percolation test (under certain conditions) (CEN, 2017 ) |
|  | |  | HJ 702-2014. Solid Waste - Determination of Mercury, Arsenic, Selenium, Bismuth, Antimony - Microwave Dissolution / Atomic Fluorescence Spectrometry (MEE China, 2014) |
|  | |  | HJ/T 299-2007. Solid waste - Extraction Procedure for Leaching Toxicity - Sulphuric Acid & Nitric Acid Method (MEE China, 2007a) |
|  | |  | HJ/T 300-2007. Solid waste - Extraction Procedure for Leaching Toxicity - Acetic Acid Buffer Solution Method (MEE China, 2007b) |
|  | |  | HJ 557-2010. Solid waste - Extraction Procedure for Leaching Toxicity - Horizontal Vibration Method (MEE China, 2010) |
|  | |  | GB/T 15555.1-1995. Solid Waste - Determination of Total Mercury - Cold Atomic Absorption Spectrometry (MEE China, 1995) |
|  | |  | Leaching Environmental Assessment Framework (LEAF) How-To Guide (USEPA, 2019) |
|  | |  | EPA Method 1313 - Liquid-Solid Partitioning as a Function of Extract pH Using a Parallel Batch Extraction Procedure[[35]](#footnote-54) (USEPA, 2017a) |
|  | |  | EPA Method 1314 - Liquid-Solid Partitioning as a Function of Liquid-Solid Ratio for Constituents in Solid Materials Using an Up-Flow Percolation Column Procedure31 (USEPA, 2017b) |
|  | |  | EPA Method 1315 - Mass Transfer Rates of Constituents in Monolithic or Compacted Granular Materials Using a  Semi-Dynamic Tank Leaching Procedure31 (USEPA, 2017c) |
|  | |  | EPA Method 1316 - Liquid-Solid Partitioning as a Function of Liquid-Solid Ratio Using a Parallel Batch Extraction Procedure31 (USEPA, 2017d) |
|  | | To determine concentrations of mercury in waste | EN 13657: Characterization of waste - Digestion for subsequent determination of aqua regia soluble portion of elements in waste (CEN, 2002c) |
|  | |  | EN 15309: Characterization of waste and soil - Determination of elemental composition by X-ray fluorescence (CEN, 2007) |
|  | | HJ 702-2014. Solid Waste - Determination of Mercury, Arsenic, Selenium, Bismuth, Antimony - Microwave Dissolution / Atomic Fluorescence Spectrometry (MEE China, 2014) |
|  | | EPA Method 7471B: Mercury in Solid or Semisolid Waste (Manual Cold-Vapor Technique) (USEPA, 2007d) |
|  | | EPA Method 7473: Mercury in Solids and Solutions by Thermal Decomposition, Amalgamation, and Atomic Absorption Spectrophotometry (USEPA, 2007e) |
|  | | EPA Method 7470A: Mercury in Liquid Waste (Manual  Cold-Vapor Technique) (USEPA, 1994) |
| Flue Gas | | | ISO 21741:2020 Stationary source emissions — Sampling and determination of mercury compounds in flue gas using gold amalgamation trap (ISO, 2020) |
| EN 13211: Air quality - Stationary source emissions - Manual method of determination of the concentration of total mercury (CEN, 2001)  \*This method determines the total mercury content  (i.e., metallic/elemental Hg + ionic Hg). |
| EN 14884: Air quality - Stationary source emissions - Determination of total mercury: Automated measuring systems (CEN, 2005) |
| HJ 543-2009. Stationary Source Emission - Determination of Mercury - Cold Atomic Absorption Spectrophotometry (Provisional) (MEE China, 2009) |
| HJ 917-2017. Stationary Source Emission - Determination of Total Gaseous Mercury - Carbon Sorbent Traps / Thermal Cracking Atomic Absorption Spectrophotometric Method (MEE China, 2017) |
| JIS K 0222:1997 Methods for determination of mercury in stack gas (JSA, 1997) |
| EPA Method 0060: Determination of Metals in Stack Emissions (USEPA, 1996) |
|  | For the chemical speciation of mercury | | ASTM D6784 - 16 Standard Test Method for Elemental, Oxidized, Particle-Bound and Total Mercury in Flue Gas Generated from Coal-Fired Stationary Sources (Ontario Hydro Method) (ASTM International, 2016) |
| Wastewater | | | ISO 12846:2012: Water quality – Determination of mercury - Method using atomic absorption spectrometry (AAS) with and without enrichment (ISO, 2012) |
|  | | | ISO 17852: 2006: Water quality – Determination of mercury - Method using atomic fluorescence spectrometry (ISO, 2006) |
| Soil | | | ISO 18400-104:2018 Soil quality - Sampling - Part 104: Strategies (ISO, 2018a) |
|  | | | ISO 18400-202:2018 Soil quality - Sampling - Part 202: Preliminary investigations (ISO, 2018b) |
|  | | | ISO 18400-203:2018 Soil quality - Sampling - Part 203: Investigation of potentially contaminated sites (ISO, 2018c) |

1. In addition, procedures and acceptance criteria for storage, handling and preparation of the sample in the laboratory, e.g., homogenization, should be established.
2. The individual steps in the analytical determination include:
   1. Extraction;
   2. Purification;
   3. Identification by suitable detectors such as inductively coupled plasma (ICP), atomic fluorescence spectroscopy (AFS), amino acid analysers and compact instruments;
   4. Quantification and reporting as required;
   5. Reporting in accordance with regulation(s).

3. Monitoring

1. In Article 10 (“International Cooperation”), paragraph 2 (b), the Basel Convention requires Parties to “cooperate in monitoring the effects of the management of hazardous wastes on human health and the environment”. In Article 19, paragraph 1 (b), the Minamata Convention requires Parties to endeavour to cooperate to develop and improve, taking into account their respective circumstances and capabilities, modelling and geographically representative monitoring of levels of mercury and mercury compounds.
2. Monitoring programmes should be implemented for facilities managing mercury wastes, as they provide an indication of whether a facility is functioning in accordance with its design and environmental regulations. The information obtained through monitoring programmes should be used to ensure that different types of mercury wastes are properly managed, to identify potential issues relating to possible mercury releases or exposure to mercury and to determine whether amendments to the management approach might be appropriate.
3. It should be noted that a number of continuous mercury measurement systems are commercially available for some types of mercury monitoring. Such monitoring may be required under national or local legislation.

E. Waste prevention and minimization

1. The prevention and minimization of mercury wastes are the first and most important steps in the ESM of such wastes. In Article 4, paragraph 2, the Basel Convention calls on Parties to “ensure that the generation of hazardous wastes and other wastes … is reduced to a minimum”. Waste prevention should be the preferred option in any waste management policy so that the need for waste management is reduced, enabling resources for ESM to be used more efficiently. Paragraphs 102 to 133 below provide information on ways to prevent and minimize mercury wastes from important sources of mercury wastes.
2. Article 5 of the Minamata Convention requires Parties to phase out mercury use in chlor-alkali and acetaldehyde production processes and to restrict mercury use in vinyl chloride monomer production, sodium or potassium methylate or ethylate, and production of polyurethane using mercury-containing catalysts (for details, see paragraph 30 of the present guidelines). Article 4 of the Minamata Convention requires Parties to not allow the manufacture of certain mercury-added products and to phase down the use of dental amalgam (for details, see paragraphs 28 and 29 of the present guidelines).

1. Waste prevention and minimization for processes

1. There are many processes where mercury wastes are generated. Artisanal and small-scale gold mining, vinyl-chloride monomer production and chlorine and caustic soda (chlor-alkali) production are major sources of mercury wastes since those processes intentionally use large volume of mercury. Industrial processes where mercury is not intentionally used could also generate mercury wastes when mercury exists in their inputs as an impurity. Such processes include, but are not limited to, coal-fired power plants/industrial boilers, smelting and roasting processes used in the production of non-ferrous metals and crude oil and natural gas production processes. Therefore, this subsection discusses waste prevention and minimization measures with respect to those processes.

(a) Artisanal and small-scale gold mining

1. According to Article 7 of the Minamata Convention, Parties that have ASGM and processing activities within their territories in which mercury amalgamation is used to extract gold from ore shall take steps to reduce, and where feasible eliminate, the use of mercury and mercury compounds in, and the emission and releases of mercury to the environment from, such mining and processing. In decision MC-1/13, the Conference of the Parties to the Minamata Convention adopted a guidance document[[36]](#footnote-55) on the development of national action plans to assist Parties in meeting their obligations under Article 7. In decision MC-3/5, the Conference of the Parties to the Minamata Convention requested the secretariat, in cooperation with the ASGM partnership area, to seek comments from Parties and other stakeholders to improve the guidance on the preparation of national action plans for ASGM regarding management of tailings from such mining.
2. Mercury-free techniques such as gravimetric methods and in combination with non-mercury methodsare available. Specific guides such as the following are available at the website of the Global Mercury Partnership, ASGM area[[37]](#footnote-56):
   1. UNEP, 2018. Illustrated Guide to Mercury Free Artisanal and Small Scale Gold Mining; and
   2. UNEP, 2017. A path to mercury-free artisanal and small scale gold mining.
3. In cases where feasible alternatives are unavailable, interim solutions that lead towards mercury-free techniques should be used. Such solutions could include mercury capture and recycling technologies such as retorts and fume hoods, mercury re-activation, and avoidance of mercury-intensive processing such as whole-ore amalgamation. Details on these interim solutions can be found in the following reference materials:
4. Artisanal Gold Council, 2014. Using Retorts to Reduce Mercury Use, Emissions, and Exposures in Artisanal and Small-Scale Gold Mining: A Practical Guide;

USEPA,

1. 2008. Manual for the Construction of a Mercury Collection System for Use in Gold Shops.
2. Artisanal miners, their families and communities in the vicinity of mining operations should be educated about the risks of exposure to mercury and related health hazards and the environmental impacts of mercury use in ASGM.
3. Once awareness of the environmental and health risks of mercury use in ASGM has increased, training in techniques and systems to prevent waste generation should be provided.

(b) Vinyl chloride monomer (VCM) production

1. VCM production using the acetylene process involves the use of mercuric chloride as a catalyst. According to Article 5, paragraph 3 in connection with Annex B of the Minamata Convention, Parties to that Convention are required to take measures to reduce the use of mercury in VCM production in terms of per unit production by 50 per cent by the year 2020 against 2010 use. Also, such Parties are required to promote measures to reduce the reliance on mercury from primary mining and to take measures to reduce emissions and releases of mercury to the environment. Methods to prevent and minimize mercury waste in VCM production fall into two primary categories: (a) alternative, mercury-free manufacturing methods; and (b) methods to better manage mercury during the production process and to capture mercury environmental releases.
2. In mercury-free VCM manufacturing, VCM is manufactured using a variety of mercury-free methods, most commonly based on the oxychlorination of ethylene (Office of Technology Assessment, 1983). While mercury-free methods are common worldwide, in several countries the acetylene process continues to be used because it is significantly less expensive in locations where coal is cheaper than ethylene (Maxson, 2011). Efforts to develop a mercury-free catalyst for the acetylene process have resulted in the development (by Johnson Matthey) of a mercury-free catalyst for the manufacture of VCM and has been commercialized. It is economically viable and can be used as a direct replacement for catalysts in existing VCM reactors[[38]](#footnote-57).
3. Suggested measures to reduce the generation of wastes contaminated with mercury include better management of mercury; development and application of low-mercury catalysts; implementation of technological reforms to prevent mercuric chloride evaporation; prevention of catalyst poisoning; and delaying carbon deposition to reduce the use of mercury. For further information, see the *Project Report on the Reduction of Mercury Use and Emission in Carbide PVC Production* (Ministry of Environmental Protection of China, 2010) and information in the publication *Global mercury supply, trade and demand* (UNEP, 2017).

(c)  Chlor-alkali production

1. Chlor-alkali production using the mercury cell involves the use of elemental mercury as an electrode. According to Article 5, paragraph 2 in connection with Annex B of the Minamata Convention, Parties to that Convention are required to phase out the use of mercury in chlor-alkali production by the year 2025 except where the Party has a registered exemption. As the mercury cell process is replaced by mercury-free processes in chlor-alkali facilities, mercury emissions and waste generation are eliminated at these facilities. Mercury-free chlor-alkali production employs either diaphragm or membrane processes. Membrane processes are more cost effective than diaphragm processes because they require less electricity (Maxson, 2011). Although the mercury cell process is being phased out, as of 2019 there were still 12 plants using the process in six countries. Solid waste from these chlor-alkali plants amounted to 10,113 kg in 2019 (World Chlorine Council, 2020).
2. Waste contaminated with mercury generated from chlor-alkali plants may include semi-solid sludges from water, brine and caustic treatment, graphite and activated carbon from gas treatment, residues from retorting, and mercury collected in tanks or sumps. In addition to monitoring of possible leakages and good housekeeping, mercury waste generation can be reduced through reduction of mercury evaporation, better control of mercury emissions, recovery of mercury from wastewater and from flue gas and caustic treatment. For further information, the following documents or website should be consulted:
   1. European Commission, 2013a. *Commission implementing decision of 9 December 2013 establishing the best available techniques (BAT) conclusions, under Directive 2010/75/EU of the European Parliament and of the Council on industrial emissions, for the production of chlor-alkali* (2013/732/EU);
   2. European Commission, 2014b. *Integrated Pollution Prevention and Control (IPPC)- Best Available Techniques (BAT) Reference Document for the Production of Chlor-alkali*;
   3. Global Mercury Partnership, Chlor-alkali sector resources[[39]](#footnote-58).
3. When decommissioning mercury cell plants, it is recommended to develop a decommissioning plan in consultation with external contractors as appropriate. The implementation phase of the decommission plan may include setting up a working area, emptying of the cells and handling of metallic mercury, as well as dismantling, demolition and decontamination of equipment and buildings. It is also recommended to use one or a combination of the oxidation and ion exchange, oxidation and precipitation and reduction and adsorption on activated carbon in order to reduce mercury releases during the decommission or conversion of mercury cell plants[[40]](#footnote-59).

(d) Coal-fired power plants and coal-fired industrial boilers

1. Coal washing is a process primarily to reduce ash, particulate matters and sulfur content of coal in order to improve its heating value. It is considered as an economical and effective technique to minimize related environmental problems (Zhao and Luo, 2017). The removal of mineral impurities also reduces operation and maintenance costs and slows the deterioration of the boiler system (UNEP, 2019c). This process can also decrease the mercury content of coal, which leads to the reduction of mercury wastes. Generally, such decrease is greater ~~reduction in~~ for bituminous coals ~~higher~~ than for lower rank coals/lignite since in lower rank coals a greater fraction of mercury is likely to be bound to the organic carbon structure (European Commission, 2017a). Removal of mercury from coal is feasible when conventional coal washing methods are used. The effectiveness of mercury removal during conventional coal washing, however, varies widely depending on the source of the coal and on the nature of the mercury within it (UNEP, 2019c). It should be noted that coal washing may generate slurry waste contaminated with mercury.
2. Coal blending is a cost-effective option to reduce the relative amount of mercury since some coal ranks have more oxidized forms of mercury. Coal blending can increase halide concentrations, thereby increasing the proportion of oxidized mercury which becomes easier to capture mercury by a flue-gas desulphurisation (FGD) scrubber (European Commission, 2017a) since oxidized mercury is soluble to water.

(e) Smelting and roasting processes used in the production of non-ferrous metals

1. Mercury content of non-ferrous metal concentrates significantly varies across ore mining areas. The mercury content of non-ferrous metal concentrates ~~is~~ contributes directly ~~linked~~ to the input of mercury to the smelting and roasting processes and also the resulting emissions and releases. Where possible, priority should be given to use low-mercury concentrates while meeting production specifications in order to minimize the generation of mercury wastes. Blending low-mercury concentrates from different sources is an alternative option to control the mercury input in the processes and to reduce the mercury output to the environment (Zhang et al., 2019).

(f) Crude oil and natural gas processing

1. Mercury is naturally present in natural gas, crude oil and bitumen but the mercury concentration significantly varies depending on reservoirs and geographic areas. Crude oil and natural gas processing facilities generate several types of mercury wastes such as waste elemental mercury, spent catalyst and adsorbent and sludge. One way to reduce mercury wastes is to return spent catalysts and adsorbents to the manufacturers for recycling.

2. Waste prevention and minimization for products containing mercury or mercury compounds

1. Introducing mercury-free alternatives and banning products containing mercury or mercury compounds are important ways to prevent the generation of mercury wastes. Under the Minamata Convention, the manufacture, export and import of specific mercury-added products are prohibited after 2020 (see paragraph 28 above). Article 4 in conjunction with Annex A of the Minamata Convention provides for the phase-out of some mercury-added products (e.g., certain mercury batteries and switches, certain lamps, pesticides, biocides and antiseptics), and phase down of mercury use in dental amalgam.
2. As a transitional measure, setting maximum limits of mercury content in products where mercury-free alternatives are not available in the foreseeable future would help to reduce the generation of mercury wastes from the products containing mercury or mercury compounds sector. Replacement of such products with mercury-free or low-mercury alternatives can be facilitated through green purchasing.
3. Where products containing mercury or mercury compounds are still in use, the establishment of a safe closed system for utilization of mercury is desirable. Mercury contamination of waste streams should be prevented through:
   1. The use of mercury-free products;
   2. Setting maximum limits of mercury content in products;
   3. Setting procurement standards to purchase mercury-free and low-mercury products.
4. Extended producer responsibility (EPR) schemes can be effective instruments to encourage the production of mercury-free or low-mercury products.

(a) Mercury-free products

1. The substitution of mercury in products depends on factors such as efficacy or performance of substitutes, substitute and overall product costs, the environmental and human health impacts of substitutes, technology, government policies and economies of scale. Many kinds of mercury-free alternatives are now available. Detailed information about mercury-free alternatives is available in the following websites and publications:
   1. Information on the uses of mercury and on non-mercury alternatives under “Review of annexes A and B” of the Minamata Convention[[41]](#footnote-60);
   2. “Many products still contain mercury. These alternatives could replace them” on the UNEP website[[42]](#footnote-61);
   3. *Replacement of mercury thermometers and sphygmomanometers in health care: Technical guidance* (WHO, 2011);
   4. “Listings of Alternatives for Mercury Added and Mercury Contaminated Products” on the National Institute of Health, U.S. Department of Health & Human Services website[[43]](#footnote-62).

(b) Setting maximum limits of mercury content in products

1. Article 4 in conjunction with Annex A of the Minamata Convention sets mercury content limits for certain mercury-added products as follows:
   1. Button zinc silver oxide batteries, button zinc air batteries: 2 per cent;
   2. Very high accuracy capacitance and loss measurement bridges and high frequency radio frequency switches and relays in monitoring and control instruments: 20 mg per bridge;
   3. Compact fluorescent lamps (CFLs) for general lighting purposes that are ≤ 30 watts: 5mg per lamp burner,
   4. Linear fluorescent lamps (LFLs) for general lighting purposes: 5 mg per lamp for triband phosphor < 60 watts, 10 mg per lamp for halophosphate phosphor ≤ 40 watts;
   5. Cold cathode fluorescent lamps and external electrode fluorescent lamps for electronic displays: 3.5 mg per lam for short length (≤ 500 mm), 5 mg per lamp for medium length (> 500 mm and ≤ 1,500 mm), 13 mg per lamp for long length (> 1,500 mm);
   6. Cosmetics: 1ppm
2. Mercury content limits should be established for products containing mercury or mercury compounds until such time as those products can be phased out. Such limits can result in less mercury being used per product in the production stage, which in turn could result in less mercury being emitted throughout the entire product lifecycle, including from accidental release or breakage, and reduce the total amount of mercury in wastes requiring mercury-specific management.
3. Mercury content limits for products can be set through legislation (see examples in section III, B, 2 below) or through voluntary industry actions under public environmental/mercury management plans. As stated previously, legal requirements for mercury content limits for batteries and fluorescent lamps have been established in the European Union, and several states of the United States have adopted mercury content limits for batteries. In Japan, mercury content limits for fluorescent lamps have been set by an industry association, and the national government has used those limits as a criterion to select fluorescent lamps under green purchasing policies. In Canada, the Products Containing Mercury Regulations set limits on the amount of mercury that fluorescent and other types of lamps can contain.
4. In order to reduce the amount of mercury in fluorescent lamps, manufacturers have developed various technologies for injecting specific amounts of mercury into individual lamps that correspond to the minimum amounts of mercury required for adequate lamp performance. Examples of methods for injecting precise amounts of mercury in lamps include using mercury amalgam, a mercury alloy pellet, a mercury alloy ring, and a mercury capsule instead of injecting mercury (Ministry of the Environment of Japan, 2010).
5. The use of mercury amalgam dosing may have environmental and performance advantages over the use of mercury throughout the life cycle of compact fluorescent lamps (CFLs) and other types of lamps containing mercury. The strength of accurate mercury dosing methods is that they minimize worker and consumer exposure to, as well as environmental releases of, mercury vapour during manufacturing, transportation, installation, storage and recycling and disposal, particularly when lamps break. In addition, accurate mercury dosing methods enable manufacturers to produce CFLs that contain very low mercury levels (two milligrams or less) while meeting important performance requirements such as high efficiency and long lamp life.

(c) Procurement

1. Procurement programmes for mercury-free products should be encouraged in order to pursue waste prevention and promote the use of mercury-free and low-mercury products. Purchasing practices should, where possible, be aimed at the purchase mercury-free products, except in the few cases where alternatives to products containing mercury or mercury compounds are practically or technologically unavailable, or at the purchase of products whose mercury content is minimized.
2. Larger users of products containing mercury or mercury compounds, such as government institutions and healthcare facilities, can play an important role in stimulating demand for mercury-free products by implementing green procurement programmes. In some cases, financial incentives could be used to encourage green procurement programmes. Some states in the United States, for instance, have subsidized the purchase of mercury-free thermometers.

3. Extended producer responsibility

1. Extended producer responsibility (EPR) is defined as “an environmental policy approach in which a producer’s responsibility for a product is extended to the post-consumer stage of a product’s life cycle”. “Producer”[[44]](#footnote-64) is considered to be the brand owner or importer except in cases such as packaging, and in situations where the brand owner is not clearly identified, as in the case of electronics, in which the manufacturer (and importer) would be considered as the producer (OECD, 2001a). EPR programmes shift the responsibility for the end-of-life management of products from local government authorities and taxpayers to producers, and can create incentives for producers to incorporate environmental considerations into the design of their products and ensure that the cost of environmentally sound treatment and disposal of those products once they have become waste are reflected in product prices. EPR can be implemented through mandatory or voluntary approaches, or a combination of the two (e.g., via negotiated agreements). Take-back collection programmes can be incorporated into EPR programmes (see paragraph 159 below).
2. EPR programmes, depending on how they are designed, can achieve a number of objectives, including: (1) to relieve local governments of the financial and in some cases operational burden of disposing of waste/products/materials; (2) to encourage companies to design products for reuse and recyclability and to reduce both the quantity and hazardousness of materials used; (3) to incorporate waste management costs into product prices; and (4) to promote innovation in recycling technology. EPR therefore promotes a market in which prices reflect the environmental costs of products (OECD 2001a). Detailed descriptions of EPR schemes are available in several OECD publications on the issue[[45]](#footnote-66).
3. When EPR programmes are used, the environmental authorities may develop regulatory frameworks setting out the responsibilities of relevant stakeholders, standards for the management of products and the components that all EPR programmes should have, and encourage participation by relevant parties and the public. The environmental authorities should also monitor the performance of EPR programmes (e.g., amount of wastes collected, amount of mercury recovered and costs accrued for collection, recycling and storage) and make recommendations for improvement as necessary. The responsibility to implement EPR programmes should be shared by all producers of a given product and there should be no “free riders” (i.e., producers who do not have to implement EPR) in such programmes so as to avoid a situation in which certain producers are forced to bear a disproportionate share of the costs of EPR that goes beyond their product market share.
4. Boxes 4 to 6 provide information on regional/national EPR programmes.

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| **Box 4 Regional Example: European Union**  In the European Union, fluorescent lamps, including CFLs, are one of the products subject to the requirements of the *Directive 2012/19/EU of the European Parliament and of the Council of 4 July 2012 on Waste Electrical and Electronic Equipment (WEEE)* (European Union, 2012). The directive requires extended producer responsibility for management of waste electrical and electronic equipment that contain, *inter alia*, mercury.  In addition, batteries are subject to the requirements of the *Directive 2006/66/EC of the European Parliament and of the Council of 6 September 2006 on batteries and accumulators and waste batteries and accumulators and repealing Directive 91/157/EEC* (European Union, 2006). Producers of batteries and accumulators and producers of other products incorporating a battery or accumulator are given responsibility for the waste management of batteries and accumulators that they place on the market. |

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| **Box 5 Country example: Canada**  In Canada, EPR programmes across Canada on mercury containing products, such as lamps and batteries[[46]](#footnote-68) are in operation. For example, in the province of Québec, The RecycFluo is a non-profit program that accepts mercury-containing light bulbs and fluorescent tubes from Québec consumers and businesses, and responsibly recycles them free of charge. It is operated by Product Care Recycling, a non-profit industry association, specializing in product stewardship, on behalf of the manufacturers, distributors and retailers of these products and funded by recycling fees paid by the consumer at the point of sale on new mercury-containing light lamps in Québec[[47]](#footnote-69). |

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| **Box 6 Country Example: Korea**  In the Republic of Korea, fluorescent lamps and batteries are targets of the EPR that mandates producers and importers of products to recycle a certain amount of wastes from products. Ones who fail to follow this obligation are subject to recycling charges. Producers fulfil the responsibility for recycling individually or by joining a PRO (Producer Responsibility Organization) [[48]](#footnote-70). |

F. Handling, separation, collection, packaging, labelling, transportation and storage

1. The procedures for the handling, separation, collection, packaging, labelling, transportation and storage of mercury wastes pending their disposal are similar to those applicable to other kinds of hazardous waste. However, because the physical and chemical properties of mercury make it very mobile in the environment, the ESM of mercury waste requires the use of additional precautions and handling techniques.
2. Specific technical guidance on the most appropriate handling of mercury wastes is provided in this section, but it is imperative that generators also consult and adhere to applicable national and local requirements. For transport and the transboundary movement of hazardous wastes, the following documents should be consulted to determine specific requirements:
   1. UNEP, 2015b. *Manual for the Implementation of the Basel Convention*;
   2. International Maritime Organization (IMO), 2018. *International Maritime Dangerous Goods Code*;
   3. International Civil Aviation Organization (ICAO), 2017. *Technical Instructions for the Safe Transport of Dangerous Goods by Air*;
   4. International Air Transport Association (IATA), 2020. *Dangerous Goods Regulations Manual*; and
   5. United Nations, 2019a. *United Nations Recommendations on the Transport of Dangerous Goods, Model Regulations*.
3. Product-specific guidance for handling, separation, collection, packaging, labelling, transportation and storage of mercury wastes is available in the following reference materials:
   1. UNDP, 2010. *Medical devices: Guidance on the Clean-up, Temporary or Intermediate Storage, and Transport of Mercury Waste from Health Care Facilities*;
   2. WHO, 2010. *Future Use of Materials for Dental Restoration* (Chapter 6, Best management practices (BMP) for amalgam waste); and
   3. Lamp Recycling Outreach Project, undated. *Training Module (1-hour version) for Generators and Handlers of Fluorescent and Mercury-Containing Lamps (and Ballasts)*.
4. Relevant information regarding the hazardous characteristics and risks of mercury wastes should be collected and analysed in order to plan the proper handling of such wastes, for example by consulting and following the instructions given on the chemicals they contain and related safety data sheets. For labelling and packaging, the United Nations Globally Harmonized System of the Classification and Labelling of Chemicals (GHS) should be taken into account, as appropriate.
5. To ensure that releases of mercury or mercury compounds are kept to a minimum, it is important to first raise awareness of those involved in handling, separation, collection, packaging, labelling, transportation and storage (e.g. transporters, recyclers, and treatment operators) about the risks of mercury. Such awareness raising can be achieved through training activities such as seminars providing information on new systems and regulations and opportunities of information exchange, preparation and distribution of leaflets and dissemination of information via the Internet.

1. Handling

1. Those who handle mercury waste should pay particular attention to the prevention of evaporation and spillage of mercury into the environment. Mercury waste should be placed in a gas- and liquid-tight containers that bear a distinctive mark indicating that they contain “toxic” mercury.
2. End users should safely handle and prevent any breakage of or damage to wastes of products containing mercury or mercury compounds such as fluorescent lamps, thermometers and electrical and electronic devices. Wastes of products containing mercury or mercury compounds such as paints and pesticides should be handled safely and should not be discharged into sinks, toilets, storm sewers or other rainfall runoff collection systems. Wastes of products containing mercury or mercury compounds should not be mixed with any other wastes. If such products are accidentally broken or spilled, clean-up procedures should be followed (see subsection III.K.2 below).
3. Those who handle wastes contaminated with mercury should not mix them with other wastes. Wastes contaminated with mercury should be placed in sealed containers to prevent the release of mercury to the environment.

2. Separation

1. Separation of mercury wastes is important because if such wastes are simply disposed of, e.g. as part of municipal solid waste (MSW), without any separation, the mercury content in them could be released into the environment.
2. Industrial mercury wastes should be separated from other wastes generated at industrial facilities in accordance with applicable national laws. Keeping such wastes separate allows for appropriate treatment to either extract the mercury from them or to stabilize the wastes for proper disposal without diluting their mercury content. Dilution of the mercury in the waste by mixing it with other wastes could make treatment less effective, or could inappropriately reduce the mercury concentration to be below the threshold(s) to be established under paragraph 2 of Article 11 of the Minamata Convention, thereby hindering the proper management of such wastes.
3. Depending on national and local legislation, separation can be facilitated through labelling of products that contain mercury. Labelling systems for products containing mercury or mercury compounds should be implemented by producers during the manufacturing stage of such products to facilitate the identification of products that contain mercury and need special handling under collection and recycling programmes[[49]](#footnote-72). Labels may need to comply with national right-to-know regulations, which may require the disclosure of the identity and properties of toxic chemical ingredients in products. Labelling systems for products containing mercury or mercury compounds could also require that labels include instructions on proper use of those products, as well as waste management instructions that encourage recycling and proper disposal.
4. Labelling systems for products containing mercury or mercury compounds could help achieve the following objectives44:
   1. Informing consumers at the point of purchase that such products contain mercury and may require special handling upon becoming waste;
   2. Identifying products containing mercury or mercury compounds at the point of collection so that they can be kept separate from other wastes.
5. Manufacturers can indicate the presence of mercury in products containing mercury or mercury compounds by using the international chemical symbol for mercury, “Hg”, on product labels. For example, products containing mercury or mercury compounds sold in certain states of the United States are required to carry this symbol[[50]](#footnote-75): hg symbol. In the European Union, the chemical symbol “Hg” must be printed on batteries containing mercury under Directive 2006/66/EC (European Union, 2006). The use of a similar symbol on the packaging labels for lamps containing mercury traded internationally could promote global recognition that such lamps contain mercury. Additional information provided in local languages could further serve to explain the meaning of the symbol.
6. Box 7 provides a national example of labelling on products containing mercury or mercury compounds.

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| **Box 7 Country Example: USA**  In the United States, the Lamp Section of the National Electrical Manufacturers Association (NEMA) maintains that a harmonized national or international approach to labelling lamps containing mercury is an essential component of the efficient and economic distribution of energy-efficient lighting[[51]](#footnote-79). On 19 July 2010, the United States Federal Trade Commission promulgated a rule[[52]](#footnote-80) requiring that, starting in 19 July 2011, packaging for CFLs, light emitting diode (LED) lamps and traditional incandescent lamps include new labels to help consumers choose the most efficient lamps for their lighting needs. For lamps containing mercury, both their packaging and the lamps themselves must be labelled as follows[[53]](#footnote-81):  **Figure:** Example of product labelling (fluorescent lamp: left for packaging, right for product)  Contains Mercury  For more on clean up and safe disposal, visit **epa.gov/cfl.**  Mercury disposal  **epa.gov/cfl.** |

1. When products containing mercury or mercury compounds are exported and become waste in importing countries, consumers, users and other stakeholders in those countries may be unable to read or understand product labels written in a foreign language. In such cases, importers, exporters, manufacturers or national agencies in charge of product labelling should make sure that product labels are locally appropriate and/or are written in the relevant local language(s).

3. Collection

1. The following issues should be considered when establishing and implementing collection programmes for mercury wastes, in particular waste of products containing mercury and mercury compounds from households and commercial and institutional entities:
   1. Allow enough time for the operation of collection programmes for the complete collection of all mercury wastes included in the programmes;
   2. Include, to the extent practical, the collection of all mercury wastes in collection programmes.

(a) Collection of waste consisting of mercury or mercury compounds

1. Wastes consisting of mercury or mercury compounds, such as those generated during the closure of mercury cell chlor-alkali facilities, are different from other mercury wastes in terms of the hazards that they may pose if mishandled. Wastes consisting of mercury or mercury compounds may also be generated in higher volumes than other mercury wastes, making their safe collection more difficult. Mercury in bulk form must be carefully packaged in appropriate containers before shipping to designated storage or disposal facilities[[54]](#footnote-83).

(b) Collection of wastes of products containing mercury or mercury compounds

1. Products containing mercury or mercury compounds, after becoming wastes, should be collected separately from other wastes and efforts should be made to prevent as much breakage and contamination as possible. The quantity of waste generated by households on the one hand and by other waste generators such as companies, governments, schools and other organizations on the other hand differs. It is therefore recommended that the waste from the two groups be collected separately.
2. There are six options, discussed below, for collecting wastes of products containing mercury or mercury compounds from households such as those referred to in table 5 above. Batteries containing mercury may be collected together with other types of batteries if they are treated through the same process.

##### (i) Waste collection stations or drop-off depots

1. Wastes of products containing mercury or mercury compounds should be discarded in a specially designed container at waste collection stations or depots in order to avoid the mixing of such wastes with other wastes. Wastes of products containing mercury or mercury compounds should be collected exclusively by collectors authorised by local governments or other appropriate authorities.
2. Containers or boxes for wastes of products containing mercury or mercury compounds should be made available for public use at waste collection stations. Indoor waste collection stations should be ventilated separately whenever possible. Waste of products containing mercury or mercury compounds such as fluorescent lamps, thermometers and batteries should only be placed in coloured, marked waste containers. Designated containers should all be the same colour and/or bear the same logo so as to facilitate public awareness. Breakage of fluorescent lamps and thermometers should be avoided, inter alia through appropriate design and the provision of written information on collection procedures. Different containers should be used for tube bulbs and CFLs. For CFLs, it is important to minimize the “free fall” of lamps by installing soft, cascading baffles or flaps. ~~could a small, open~~ Another option to minimize lamp breakage would be to ask consumers to hand fluorescent lamps over to qualified collection station staff members. In the event of lamp breakage, staff should be informed and apply clean up procedures[[55]](#footnote-85). Waste drop-off stations for lamps could be established outdoors to minimize worker exposure should lamp breakage occur.

##### (ii) Collection at public places or shops

1. Waste of products containing mercury or mercury compounds, particularly used fluorescent lamps, thermostats, batteries and thermometers may be collected by specially designed collection vehicles or at public places such as near or in municipal buildings, electronics stores, shopping malls and other retail outlets, provided that appropriate collection containers are used and that necessary measures to prevent mercury exposures from broken products are taken. Indoor waste collection places should be ventilated separately whenever possible. Separate collection containers or boxes for wastes of products containing mercury or mercury compounds should be designed to accommodate the characteristics of such wastes and to minimize waste breakage. Only containers specifically designed to carry wastes containing mercury or mercury compounds and shown to be capable of containing mercury vapours from broken lamps should be used in any outdoor public collection locations[[56]](#footnote-87). Consumers should be able to take used fluorescent lamps, batteries, thermostats and thermometers to such locations free of charge. Authorized collectors, such as municipal collectors or private sector collectors (e.g., collectors trusted by producers), should collect the wastes in the designated waste collection boxes or containers.
2. Containers or boxes for wastes of products containing mercury or mercury compounds should be monitored to ensure that no other type of waste is deposited in them. They should be labelled and placed in areas where they can be monitored, either inside buildings, in well-ventilated areas, or outside, in covered and protected areas. In any indoor retail settings, they should be placed away from areas of high foot traffic and positioned in a way to prevent them from being hit or bumped. This could involve placing the containers in a corner or surrounding them with a barrier to direct the flow of people and minimise any accidental contact.

##### (iii) Collection at households by collectors

1. The collection of wastes of products containing mercury or mercury compounds from households by authorized collectors can be carried out to deal with certain wastes such as   
   mercury-containing e‑waste. In order to ensure the efficient collection of such wastes by local collectors, special arrangements or legal mechanisms will often be required; for example, government or other agencies and producers of products containing mercury or mercury compounds may need to provide arrangements for the collection of wastes of such products by local collectors.

##### (iv) Collection coordinated by business associations

1. The collection of wastes of products containing mercury or mercury compounds from business and commercial entities could be carried out by business or commercial associations in an efficient manner. In Japan, for example, the Tokyo Medical Association established an ad hoc collection system for unnecessary mercury thermometers and sphygmomanometers that collected several thousands of these devices during a collection period of one month. During the collection period, each member medical institution was encouraged to bring the devices to a designated local association branch office and requested to pay specific fees for transportation and disposal. The Tokyo Medical Association coordinated with local association branches and waste transporters and managers to ensure the efficient collection and disposal of devices collected. Each member medical institution benefited from lower transportation fees because the system created economies of scale and efficient transport arrangements.

##### (v) Take-back collection programmes

1. Take-back programmes can refer to a variety of programmes established to divert spent products from the waste stream for the purpose of recycling, reuse, refurbishing, and in some cases recovery. Take-back programmes are often voluntary initiatives developed by the private sector   
   (e.g., manufacturers and retailers) that offer consumers the opportunity to return used products at their points of purchase or at some other specified facilities. Some take-back programmes offer financial incentives to consumers, others are mandated or operated by governments (e.g., bottle deposits schemes), and some serve to partly finance disposal or recycling activities. Take-back collection programmes generally focus on consumer products that are widely used, such as batteries, switches, thermostats, fluorescent lamps and other products containing mercury or mercury compounds (Honda, 2005).

##### (vi) Pre-paid shipping services

1. Some mercury-containing product recyclers offer a recycle-by-mail concept where consumers or waste generators purchase boxes or containers from the recycler, which includes the cost of delivery. Waste mercury-containing products are placed in the box, and once it is full, is shipped back to the recycler. This pre-paid shipping service is convenient for small quantity generators, and for those in remote locations that otherwise may not have access to other types of collection programs or have limited transportation options. The type of waste accepted via this program may vary by recycler.

(c) Collection of wastes contaminated with mercury or mercury compounds

1. Sewage treatment plants and waste incinerators are generally designed to include equipment for collecting sewage sludge, ash and residues that might contain trace amounts of mercury and other heavy metals. Mercury air pollution control devices in incinerators may increase mercury concentrations in collected fly ash. If mercury concentrations in these wastes exceed the criteria for hazardous waste or mercury wastes, the wastes should be collected separately.

4. Packaging and labelling

1. Mercury wastes being transported from the premises of generators or from public collection points to waste treatment facilities should be properly packaged and labelled. Packaging and labelling for transport are often controlled by national hazardous waste or dangerous goods transportation legislation, which should be consulted first. If such legislation is lacking or does not provide sufficient guidance, care should be taken to use labels that are in line with the United Nations *Globally Harmonized System of Classification and Labelling of Chemicals*. IATA, IMO and UNECE should be consulted. International standards for the proper packaging, labelling and identification of wastes have been developed, including the following reference materials:
   1. United Nations*,* 2019b. *Globally Harmonized System of Classification and Labelling of Chemicals* (revised and improved every two years); and
   2. OECD, 2001b. *Harmonised Integrated Classification System for Human Health and Environmental Hazards of Chemical Substances and Mixtures*.

5. Transportation

1. Mercury wastes should be transported in a way to avoid accidental spills; they should also be tracked during transport until they have reached their final destination. Prior to transportation, contingency plans should be prepared in order to minimize environmental impacts associated with spills, fires and other potential emergencies. The shipping paper should include an emergency response telephone number and a certificate that the shipment is in compliance with the regulations. In addition, the shipper should mark the containers with appropriate signs, including the specified label, the proper shipping name and the UN number. For mercury, the proper shipping name is “Mercury,” and the UN number is “UN 2809” (QSC, 2003). During transportation, mercury wastes should be identified, packaged and transported in accordance with the following: (a) *United Nations Recommendations on the Transport of Dangerous Goods: Model Regulations* (Orange Book) (United Nations, 2019a); (b) *International Maritime Dangerous Goods Code* (IMO, 2018); (c) *Technical Instructions for the Safe Transport of Dangerous Goods by Air* (ICAO, 2017); and (d) *Dangerous Goods Regulations Manual* (IATA, 2020).
2. Companies transporting mercury wastes within their own countries should, if appropriate, be certified as carriers of hazardous materials and wastes, and their personnel should, if appropriate, be qualified and certified as handlers of hazardous materials and wastes in accordance with applicable national and local requirements. Transporters of mercury wastes should have an insurance that covers the expenses caused by personal, material and environmental damages from any accidents that may occur during the transportation of mercury wastes. Transporters should manage mercury wastes in a way that prevents breakage, environmental releases and exposure to moisture.
3. Transports of waste consisting of mercury should be accompanied by information that demonstrates the mercury’s level of purity and identifies any contaminants. Upon arrival, the transport vehicle should be visually inspected for any obvious leaks, spills, droplets or other pools of free elemental mercury and all suspected mercury sources should be documented and reported to management. The shipment should be accepted as compliant or rejected as non-compliant on the basis of the inspection; a written report including all the relevant information should be kept by the facility (QSC, 2003). When rejecting the transport, the facility operator should ensure that the required actions under the contingency plan have been taken without causing further spread of leakage outside of the facility. Guidance on the safe transportation of hazardous materials can be obtained from IATA, IMO, UNECE and ICAO (see paragraph 135 above).

6. Storage

(a) Storage of wastes containing mercury by waste generators pending collection

1. Wastes of products containing mercury or mercury compounds should be stored temporarily at the premises of waste generators before they are collected for disposal. Wastes of products containing mercury or mercury compounds should be stored safely and kept apart from other wastes until they are brought to waste collection stations or facilities or picked up by collection programmes or contractors. Bulk waste should be stored in such a manner as to minimize mercury releases to the environment, including, if feasible, through the use of closed containers, impermeable concrete pads with runoff controls, or waterproof tarp covers. Waste should be stored by generators for a limited period of time, as allowed by national standards or regulations, and should in any case be sent off-site for appropriate disposal as soon as is practical.
2. Household wastes containing mercury or mercury compounds, mainly fluorescent and other lamps, batteries and thermometers, should be appropriately packaged (e.g., using product packaging or boxes that fit the shape of the wastes) and temporarily stored. Any mercury devices that are broken during handling should be cleaned up and all clean-up materials should be stored outdoors until they are collected for further management[[57]](#footnote-89). Liquid wastes containing mercury or mercury compounds such as paints and pesticides should be kept in their original containers, whose lids should be tightly closed. Containers and packages holding wastes of products containing mercury or mercury compounds should not be placed together with other wastes; containers and packages should be marked and stored in a dry and secure place, such as a warehouse or other space that is not usually frequented by people.
3. In addition to following the guidance contained in paragraphs 166 and 167 above, large-scale users of products containing mercury or mercury compounds such as governments, businesses and schools will also need plans for storing large amounts of such wastes. When original boxes or packages are not available, containers that are specially designed to store wastes of products containing mercury or mercury compounds (e.g., fluorescent lamp containers) should be purchased. Containers or boxes for storing such wastes should be marked and dated and stored in a dry place. The use of an area or room separated from work areas or areas not open to the public for storing such wastes is recommended. Such areas should not share building ventilation systems with work or public areas and should have their own ventilation systems or be vented directly to the outdoors. Guidance developed by UNDP for mercury wastes generated by healthcare facilities[[58]](#footnote-91) provides detailed advice in this regard and may be applicable to many commercial facilities that generate waste mercury devices as well as local governments and other entities that need to store collected wastes of products containing mercury or mercury compounds for a longer period.

(b) Storage of mercury wastes pending disposal operations

1. During the storage of mercury wastes at disposal facilities, the potential for mercury releases to the environment should be minimized.

(i) Technical and operational considerations for storage facilities

1. In terms of siting and design, in principle storage facilities should not be built in sensitive locations whenever possible. Sensitive locations may include floodplains, wetlands, groundwater, earthquake zones, Karst terrain, unstable terrain and areas with unfavourable weather conditions and incompatible land uses in order to avoid any risk of mercury releases and possible human and environmental exposure to mercury. However, such location limitations may not apply in cases where technical design and legal requirements govern the ESM of storage facilities. Storage areas should be designed to ensure the security of the facilities and to prevent unnecessary chemical and physical reactions to mercury. The floors of storage facilities should be covered with mercury-resistant materials to prevent seepage or penetration of mercury from accidental leaks and spills. Storage facilities should have fire alarm and fire suppression systems. To reduce the risks of fire, facilities should be constructed of non-combustible materials, which should be used as well for pallets, storage racks and other interior furnishings. To further minimize the risks of fire, it is suggested that battery-powered electric forklifts be used to transport the mercury or mercury compounds inside the storage facility. The temperature in storage areas should be maintained as low as feasibly possible, preferably at a constant temperature of 21 degrees Celsius. Storage areas for mercury wastes should be clearly marked with warning signs (FAO, 1985; USEPA, 1997b; UNEP, 2022a; U.S. Department of Energy, 2009).
2. In terms of operation, storage facilities should be kept locked to avoid theft or unauthorized access. Access to mercury wastes should be restricted to those with adequate training in, among other things, the identification of types of mercury wastes, mercury-specific hazards and the handling of such wastes. It is recommended that storage buildings for suitable types of mercury wastes not be used to store other liquid wastes or materials. A full inventory of the wastes kept at the storage site should be created and updated as waste is added or removed. Regular inspections of storage areas should be undertaken, focusing particularly on damage, leaks, spills and deterioration. Clean-up and decontamination should be carried out speedily, but not without alerting the authorities concerned, and in compliance with national laws and regulations. (FAO, 1985; USEPA, 1997b).
3. In terms of the safety of facilities, site-specific procedures should be developed to implement safety requirements identified for the storage of mercury wastes. A workable emergency plan, preferably with multiple procedures, should be in place and implemented immediately in the event of accidental spillage and other emergencies. The protection of human life and the environment is of paramount importance. In the event of an emergency, there should be a responsible person who can authorize modifications to safety procedures when necessary in order to allow emergency response personnel to act. Adequate security siting and access to the area should be ensured (Environmental Management Bureau, Republic of the Philippines, 1997; UNEP, 2022a; U.S. Department of Energy, 2009).

(ii) Special considerations for storage of wastes consisting of mercury or mercury compounds

1. Containers of wastes consisting of mercury or mercury compounds should be structurally sound and make possible the environmentally sound storing of such wastes. Containers should meet the following requirements: (1) they should not be damaged from any materials previously stored in them or have contained materials that could adversely react with mercury; (2) their structural integrity should be intact; (3) they should not be corroded; and (4) they should have a protective coating (paint) to prevent against corrosion (if not made from stainless steel). Appropriate materials for mercury containers include carbon (minimum ASTM A36) and stainless steel (AISI 304 or 316L) (European Union, 2010), which do not react with mercury at ambient temperatures. There are two main types of internationally approved mercury storage and transport containers: 34.5 kg flasks and one-metric-ton containers (QSC, 2003). The design type of the container should pass the drop test and the leak proof tests as described in chapters 6.1.5.3 and 6.1.5.4 of the United Nations Recommendations on the Transport of Dangerous Goods: Manual of Tests and Criteria (European Union, 2011b). For transporting smaller quantities of mercury, other sizes (e.g. 1-8 kg) and types (e.g., polyethylene, glass) of containers are often used (QSC, 2003); however, the level of protection such containers provide should be taken into account. When storing mercury in containers, it is important to leave some “head space” to allow for thermal expansion of the mercury. In European Union guidance, the maximum filling ratio of a container is 80 per cent by volume, and the head space is therefore at least 20 per cent (European Union, 2011b). Other jurisdictions specify a maximum filling ratio of 85 per cent, for a 15 per cent head space. No protective coating is required for the inner surface of such containers as long as the mercury to be stored in them meets purity requirements and no water is present inside the container. Protective coatings (e.g., epoxy paint and electroplating) should be applied to all exterior carbon steel surfaces in a manner that does not leave any steel exposed. Coatings should be applied in a manner that minimizes paint blistering, peeling and cracking. Labels should include information on the content (description) and risks, the names of the supplier/generator of the mercury waste, the origin of the waste, the container number, net weight and gross weight, the date when the mercury was injected and a “corrosives” label indicating that the container contains corrosive materials, should be affixed to each container (U.S. Department of Energy, 2009). In addition, the label should show that the container meets specific technical requirements regarding tightness, pressure stability, shock resistance, behaviour when exposed to heat, etc.
2. In the storage facilities, containers for wastes consisting of mercury or mercury compounds should be stored upright on pallets off the ground. Aisles in storage areas should be wide enough to allow the passage of inspection teams, loading machinery and emergency equipment. The floors of storage facilities should be coated with an epoxy coating and be light in colour to allow the detection of mercury droplets. The floor and floor coating should be inspected frequently to ensure that the floor has no cracks, and the coating is intact. The floor should not be penetrated by any drains or plumbing, and sloped floors and open flow gutters with rounded-down edges (to avoid mercury trapping under gutter covers) could be used to facilitate the collection of spills. The liquid containment volume should have enough margin from the maximum liquid volume, taking into account the space taken up by items stored in the containment area. If two or more containers are stored at a containment area, the volume of the largest container should be used as the maximum liquid volume. When choosing the materials from which to construct the walls of storage facilities, materials that do not readily absorb mercury vapour should be selected. It is important to include redundant systems to prevent mercury releases in the event of an unexpected occurrence. Areas used to store mercury or mercury compounds should have their own ventilation systems or be vented directly to the outdoors. Ideally, ventilation systems for handling areas should be filtered or include pollution control devices to capture any mercury vapor or dust. When mercury is transferred between containers, negative pressure environments are recommended in order to avoid mercury emissions escaping from buildings (U.S. Department of Energy, 2009; Euro Chlor, 2004; UNEP, 2018).
3. When storing wastes consisting of mercury or mercury compounds, the mercury content of such wastes should be as pure as possible in order to avoid chemical reactions and the degradation of containers. A mercury content greater than 99.9 weight per cent is recommended. For purification techniques, see subsection III.G.1 (b) (v) below.

(iii) Special considerations for storage of wastes contaminated with mercury or mercury compounds

1. Liquid wastes in containers should be placed in containment trays or curved, leak-proof areas. The liquid containment volume should be at least 125 per cent of the maximum liquid waste volume, taking into account the space taken up by stored items in the containment area.
2. Solid wastes should be stored in sealed containers such as barrels or pails, steel waste containers or in specially constructed containers that do not release mercury vapour.

G. Environmentally sound disposal

1. The following disposal operations, as provided for in Annex IV, parts A and B, of the Basel Convention, should be permitted for the purpose of ESM of mercury wastes[[59]](#footnote-93):

D5 Specially engineered landfill;

D9 Physico-chemical treatment;

D12 Permanent storage;

D13 Blending or mixing[[60]](#footnote-95) prior to submission to operations D5, D9, D12, D14 or D15;

D14 Repackaging prior to submission to operations D5, D9, D12, D13 or D15;

D15 Storage pending operations D5, D9, D12, D13 or D14;

R4 Recycling/reclamation of metals and metal compounds;

R5 Recycling/reclamation of other inorganic materials;

R7 Recovery of components used for pollution abatement;

R8 Recovery of components from catalysts;

R12 Exchange of wastes[[61]](#footnote-97) for submission to operations R4, R5, R8 or R13;

R13 Accumulation of material intended for operations R4, R5, R8 or R12.

1. In addition, the stowage of suitable mercury waste in underground facilities may also be permitted whereby waste is utilized in underground facilities for mining safety purposes taking advantage of the respective structural properties of the waste, except for waste consisting of mercury (metallic mercury) that underwent stabilization and solidification (S/S)[[62]](#footnote-99). Box 8 provides information on national legislation related to this issue.

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| **Box 8 Country Example: Germany**  In Germany, for example, stowing of suitable waste in underground facilities for mining safety purposes is regulated by the *Ordinance on Underground Waste Stowage* (see: <http://www.bmu.de/fileadmin/bmu-import/files/pdfs/allgemein/application/pdf/underground_waste_stowage.pdf>), whose requirements are equivalent to those found in the European Union’s *Council Directive 1999/31/EC of 26 April 1999 on the landfill of waste* and include special licensing procedures and supervision. |

1. Wastes of products containing mercury or mercury compounds should be treated to remove or recover mercury in accordance with the guidance provided in section 1 below. This treatment will produce wastes consisting of mercury or mercury compounds and wastes contaminated with mercury or mercury compounds.
2. Should one of the mercury recovery processes described in subsection III.G.1 below be carried out and the mercury subsequently sent to a specially engineered landfill or permanent storage   
   (i.e., operations D5 or D12), such mercury recovery would fall under operations D13 and D9   
   (i.e., blending or mixing, and physico-chemical treatment, respectively). On the other hand, soil washing and acid extraction described under subsection III.G.2(ii) below can lead, under specific conditions, to subsequent “R” operations listed in paragraph 178 above. In this case, the operation would also constitute an “R” operation. These two conclusions may not apply to all countries.

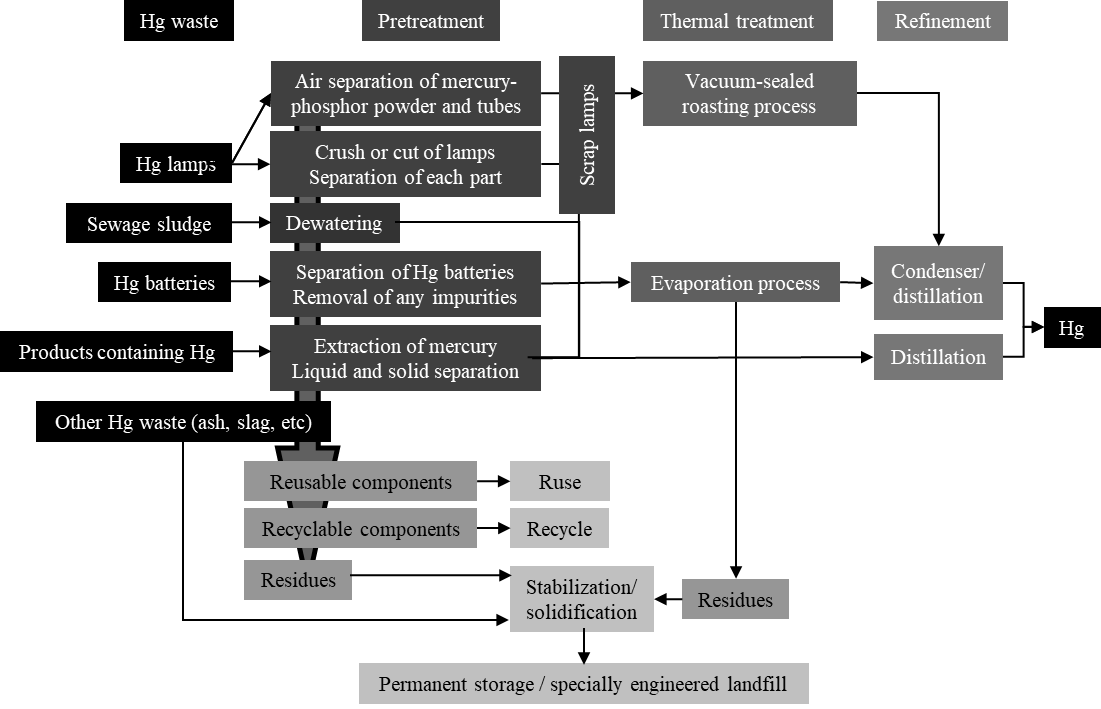
1. Mercury recovery processes

1. Mercury recovery from solid waste generally comprises: (1) pre-treatment; (2) thermal treatment; and (3) purification, as shown in figure 3 below. In order to minimize mercury emissions from the mercury recovery process, facilities should employ closed systems. The entire process should take place under reduced pressure in order to prevent leakage of mercury vapour into processing areas (Tanel, Reyes-Osorno, and Tansel, 1998). The small amount of exhausted air that is emitted in the recovery process should pass through a series of particulate filters and a carbon bed that absorbs the mercury before the air is exhausted into the environment. For example, with certain measures for mechanical treatment of waste electrical and electronic equipment, it is possible to reduce emission levels for channelled mercury emissions to air to 2-7 μg/Nm3 (European Commission, 2018a).
2. Examples for mercury wastes whose recovery can generate mercury emissions are wastes of products containing mercury or mercury compounds that easily release mercury into the environment when broken and wastes contaminated with high concentrations of mercury. The former includes measuring devices containing mercury (thermometers, sphygmomanometers and manometers) and mercury switches and relays, and could also include lamps containing mercury. Mercury waste can also include wastewater treatment sludges from wet scrubbers in non-ferrous metal smelters, depending on their mercury content. Box 9 provides information on national legislation for the treatment of mercury-bearing hazardous wastes.

|  |
| --- |
| **Box 9 Country Example: USA**  In the United States, standards require mercury recovery for reuse for mercury-bearing hazardous wastes with a mercury content greater than or equal to 260 mg/kg. Waste containing less than 260 mg/kg can be land disposed following treatment to control mercury leaching (see U.S. Code of Federal Regulations (CFR), Title 40 (Protection of the Environment), Section 268.40: Applicability of treatment standards). |

1. The Basel Convention *Technical guidelines on the environmentally sound recycling/reclamation of metals and metal compounds (R4)* focus mainly on the environmentally sound recycling and reclamation of metals and metal compounds, including mercury, that are listed in Annex I to the Convention (“Categories of wastes to be controlled”). It is possible to recycle mercury wastes, particularly wastes consisting of mercury or mercury compounds, in special facilities that have advanced mercury-specific recycling technologies. It should be noted that appropriate procedures should be employed in such recycling to prevent any releases of mercury into the environment. In addition, recycled mercury may be sold on the international commodities market, where it can be   
   re-used. The recycling of mercury[[63]](#footnote-101) will usually be determined by the degree of allowable use, demand, and a commercial evaluation as to whether the mercury can be profitably reused.

**Figure 3:** Flow of mercury recovery from solid waste (Nomura Kohsan Co. Ltd., 2007)



1. Mercury in wastewater should generally be captured through chemical oxidation, chemical precipitation, or adsorption, followed by various treatment processes. Mercury exists in wastewater due to accidental or intentional discharging of mercury from dental amalgams, or industrial processes using mercury or mercury compounds. Mercury may be found in wastewater from wet-type air pollution control devices and leachate from landfills and dumping sites where wastes containing mercury such as mercury thermometers were disposed of or dumped. To the extent possible, mercury in wastewater should not be released into the aquatic environment, where mercury is methylated and converted into methylmercury, which then bioaccumulates and biomagnifies in the food chain. With appropriate pre-treatment in certain processes such as adsorption, it is possible to reduce discharges of mercury to a receiving water body for certain solid wastes to 0,5-5 μg/l expressed as Hg (for metals and metalloids) in daily average or average over the water discharge duration, and for water-based liquid wastes to 1-10 μg/l (European Commission, 2018a). Mercury content from the influent should be considered when establishing wastewater treatment procedures.
2. Pre-treatment prior to recycling/reclamation of mercury (operation R4) falls under operation R12 (see paragraph 178 above), while roasting in a furnace (hereinafter referred to as “roasting"), purification, chemical oxidation/precipitation and adsorption fall under operation R4.

(a) Pre-treatment (exchange of wastes for submission to operations R4 or R13)

1. Before undergoing thermal treatment, wastes that contain or are contaminated with mercury are treated in order to increase the efficiency of the thermal treatment; pre-treatment processes include, *inter alia*, removal of materials other than those containing mercury by crushing and air separation, dewatering of sludge, and removal of impurities. Examples of mercury waste-specific   
   pre-treatment operations are summarized in table 7 below.
2. For the reduction of mercury emissions from waste pre-treatment, the following document provides additional technical information: European Commission, 2018b. *Best Available Techniques (BAT) Reference Document Waste Treatment*.

**Table 7:** Examples of pre-treatment operations by type of waste

| **Waste Type** | **Pre-treatment** |
| --- | --- |
| *Fluorescent lamps* | *Mechanical crushing*  Waste lamps containing mercury should be processed in a machine that crushes and separates the lamps into three categories: glass, end-caps and  mercury-phosphor powder mixtures. This is accomplished by injecting the lamps into a sealed crushing and sieving chamber. Upon completion, the chamber automatically removes the end products to prevent cross-contamination.  End-caps and glass should be removed and sent for reuse in manufacturing. However, the metal pins of the end caps should be removed and treated separately as their mercury content may be considerable. Mercury-phosphor powder may be disposed of or further processed to separate the mercury from the phosphor (Nomura Kohsan Co. Ltd., 2007).  Lamp glass from the crushed lamps can contain significant amounts of mercury and should be treated thermally or in other ways to remove the mercury before sending the glass for recovery (Jang, Hong, and Park, 2005) or disposal. If the glass is sent to a melting unit for re-melting as part of the recovery process, the unit should have air pollution controls specifically directed at capturing released mercury (e.g., activated carbon injection).  A high-performance exhaust air system should prevent the emission of any mercury vapours or dust during the entire process. |
|  | *Air separation*  The aluminium end caps of fluorescent lamps (straight, circular and compact tubes) are cut by hydrogen burners. Air is then blown into the cut lamps from the bottom of the lamps to remove mercury-phosphor powder adsorbed on lamp glass (Jang, Hong, and Park, 2005). Mercury-phosphor powder is collected in a precipitator, and glass parts are crushed and washed with acid; mercury-phosphor powder adsorbed on the glass is completely removed. In addition, the end-caps are crushed, and aluminium, iron and plastics are magnetically separated for recycling (Kobelco Eco-Solutions Co. Ltd., 2001; Ogaki, 2004). |
| *Batteries containing mercury* | ***Removal of impurities***  In order to recycle mercury from batteries containing mercury, the batteries should be collected separately and stored in suitable containers before treatment and recycling. If batteries containing mercury are collected together with other types of batteries or with waste electrical and electronic equipment (WEEE), the batteries containing mercury should be separated from other types of batteries. However, when a facility can treat batteries containing mercury, other types of batteries, and WEEE together without emitting mercury to outside of the facility, such separation may not be necessary. Before roasting treatment, impurities mixed with and adsorbed onto batteries containing mercury should be removed, preferably through a mechanical process. In addition, mechanical screening of the size of the batteries is necessary for an effective roasting process. (Nomura Kohsan Co. Ltd., 2007). |
| *Sewage sludge* | ***Dewatering***  Sewage sludge has a high water content of over 95 per cent. As a result, sludge contaminated with mercury needs to be dewatered to about 20 to 35 per cent solids before any thermal treatment. After dewatering, sewage sludge with high mercury content should be treated through a roasting process (Nomura Kohsan Co. Ltd., 2007; USEPA, 1997a). Extracted water may need to be managed as mercury waste, depending on the mercury content. |
| *Wastes of products containing mercury* | ***Extraction***  Wastes of products containing mercury such as thermometers and barometers should be collected without any product breakage, to the degree feasible. After collection of such wastes, mercury in the products should be extracted, and so the extracted mercury distilled for purification under reduced pressure. |
| *Wastes containing mercury attached to devices* | ***Dismantling***  Wastes containing mercury, such as electric switches and relays, are usually attached to electric devices. Therefore, such wastes should be removed from such devices without breaking the devices’ outer glass.  Computer monitors and televisions with flat screen liquid crystal displays (LCD) contain one or more small lamps for illumination, usually located along the outside edge of screens. While current technologies use light emitting diodes (LED) in these small lamps, old LCD screens contain fluorescent lamps. These lamps will often break during handling and mechanized processing and release mercury vapour, so they should be carefully removed by hand[[64]](#footnote-103) and should not be treated through mechanized processing such as shredding, unless shredding machines have the necessary pollution control equipment to manage such operations and are licensed and permitted for the processing of fluorescent lamps, such as at mercury treatment facilities. For further information, see section 7.3 of the Basel Convention Partnership for Action on Computing Equipment (PACE) *Guideline on environmentally sound material recovery and recycling of end-of-life computing equipment* (UNEP, 2013). Further information on the presence of mercury in LCD backlights is available in *Demonstration of flat panel display recycling technologies* (Waste and Resources Action Programme, 2010). |

(b) Recycling/reclamation of mercury or mercury compounds (R4)

(i) Thermal treatment

1. Thermal treatment facilities for wastes containing or contaminated with mercury or mercury compounds such as sewage sludge, contaminated soils and other wastes from contaminated sites should be equipped with mercury vapour collection technology for the recovery of mercury (ITRC, 1998; Chang and Yen, 2006).
2. Thermal desorption is a process that uses either indirect or direct heat exchange to heat primarily organic contaminants to a high enough temperature to volatilize and separate them from a contaminated solid matrix and then either collect or destroy them. Thermal desorption using indirect heat exchange is the recommended option for mercury and its compounds. Air, combustion gas or an inert gas is used as a transfer medium for vaporized components. Thermal desorption systems involve physical separation processes that transfer contaminants from one phase to another. The systems have two major components, a desorber and an off-gas treatment/collection system[[65]](#footnote-105).
3. There are several evaporation processes for the treatment of mercury wastes, including rotary kiln distillation and vacuum thermal processing.
4. Rotary kiln distillation serves to remove and recover mercury in wastes such as mineral industrial slurries, slurries from the movement of natural gas, activated carbon, catalysts, button cells and contaminated soil by means of evaporation and recycling of the resulting mercury‑free products (e.g., glass, iron and non-ferrous metals, zeolites).
5. In the rotary kiln distillation process, input wastes, such as activated carbon, catalysts, soils, batteries and conditioned sludge, are transported by a scraper chain conveyer directly into the feeding screw of the rotary kiln. The wastes are continuously treated at temperature of up to 800 °C. As a result, mercury is transformed into a gas phase and can be removed as exhaust vapor from the rotary kiln. The treatment is carried out at under pressure in order to prevent dust emissions. The exhaust gas from the rotary kiln is transported by a cyclone, which separates dust particles, into the post-combustion chamber. After combustion, the remaining exhaust gas reaches a three-stage treatment, in which the exhaust gas is cooled down to achieve a condensation of mercury and water, passed through an electro filter to achieve the deposition of aerosols, and then fed to an active carbon filter system for final cleaning[[66]](#footnote-107).
6. Pre-treated waste, such as mercury-phosphor powder in fluorescent lamps, crushed lamp glass, cleaned batteries containing mercury, dewatered sewage sludge and screened soil, may be treated in roasting/retorting facilities equipped with mercury vapour collection technology through which mercury can be recovered. However, it should be noted that volatile metals, including mercury, and organic substances (including persistent organic pollutants), are emitted during roasting and other thermal treatments and are transferred from the input waste to both the resulting flue gas and fly ash. Therefore, flue gas treatment devices should be used to capture volatilized pollutants and to prevent the pollutants from being emitted to the environment (see subsection III.H.1 below).
7. Waste contaminated with mercury or mercury compounds can be treated in a continuous roasting and reactivation kiln. In the furnace the material is cleaned at temperatures of up to 850 °C. At this temperature volatile compounds that are absorbed, adsorbed or chemisorbed on the material are thermally destroyed and/or vaporized; this includes mercury and sulphur. It is important to note that at 850 °C the very stable mercury sulphide (HgS or cinnabar) is broken down into mercury and sulphur. The system is therefore especially qualified for the treatment of adsorbents (e.g., S-impregnated activated carbon, mercury guards) that fix the mercury in the form of HgS on the adsorbent. Activated carbon is reactivated by means of temperature and steam releasing contaminants trapped in the pores and opening those pores up to increase the surface area of the carbon. This process maximises the efficiency of the activated carbon as a filter material for re-use. Activated carbon can be re-charged in this manner multiple times before disposal. Spent catalysts such as Hg guards are decontaminated so that a subsequent Cu or Zn recovery in a smelter is possible. Pollutants, such as mercury, remaining after the furnace travel with the exhaust gas into the waste gas treatment system. Here they first pass a post combustion chamber operating at temperatures between 1,000 and 1,200 °C, where organic compounds are incinerated. After the post combustion chamber, the waste gas is quenched to 40 °C and finally cleaned in a system consisting of 4 scrubbers (acid, basic, lime milk, peroxide), a bag filter and a final activated carbon filter. Mercury is transferred into the slurry of the first 2 scrubbers. This slurry is removed in batches and transferred to the mercury distillation plant for recovery of the mercury in a separate process for further treatment[[67]](#footnote-108).
8. Vacuum thermal processing enables the treatment of thermometers, batteries, especially button cells, dental amalgam, electrical switches and rectifiers, fluorescent powder, exhaust tubes, crushed glass, soil, sludge, primary mercury mining residues and catalyst materials, among others. The vacuum thermal process generally includes the following stages:
   1. Heating the input waste in a special kiln or in a charging operation at temperatures of between 340 °C and 650 °C and pressures of a few millibars so that the mercury contained in the waste evaporates;
   2. Applying thermal post-treatment to mercury-containing vapour at temperatures ranging from 800 °C to 1,000 °C, where, for example, organic components can be destroyed;
   3. Collecting and cooling of mercury‑containing vapour;
   4. Using distillation to generate pure liquid mercury.
9. The residue that remains at the end of the vacuum thermal processing is essentially mercury‑free and is recycled or otherwise disposed of depending on its composition[[68]](#footnote-110).

(ii) Chemical oxidation

1. Chemical oxidation of mercury and organomercury compounds in waste is carried out to convert the mercury into mercury salts and to destroy the organomercury compounds. Chemical oxidation is an effective method for treating liquid and aqueous waste, such as slurry and primary mercury or ASGM tailings, containing or contaminated with mercury. Oxidizing reagents used in the chemical oxidation process include sodium hypochlorite, ozone, hydrogen peroxide, chlorine dioxide and free chlorine (gas). Chemical oxidation can be conducted as a continuous or batch process in mixing tanks or plug flow reactors. Mercury halide compounds formed in the oxidation process are separated from the waste matrix, treated and sent for subsequent treatment such as acid leaching and precipitation (USEPA, 2007a).

(iii) Chemical precipitation

1. Precipitation uses chemicals to transform dissolved contaminants into insoluble solids that may precipitate or be removed by flocculation or filtration. In co-precipitation, the target contaminant may be in a dissolved, colloidal or suspended form. Dissolved contaminants do not precipitate but are adsorbed onto other species that are then precipitated. Colloidal or suspended contaminants become enmeshed with other precipitated species or are removed through processes such as coagulation and flocculation. Processes to remove mercury from wastewater can include a combination of precipitation and co-precipitation. The precipitated/co-precipitated solid is then removed from the liquid phase through clarification or filtration. More detailed information can be found in the report entitled *Treatment technologies for mercury in soil, waste, and water* (USEPA, 2007b).

(iv) Adsorption treatment

1. Adsorption materials hold mercury on their surfaces through various types of chemical forces such as hydrogen bonds, dipole-dipole interactions and van der Waals forces. Adsorption capacity is affected by surface area, pore size distribution and surface chemistry. Liquid wastes are passed through adsorption materials that are usually packed into a column, through which mercury or mercury compounds are adsorbed. The column should be regenerated or replaced with new media when adsorption sites become filled (USEPA, 2007b). The resulting spent adsorbent is mercury waste.
2. Examples of adsorption materials include activated carbon and zeolite. Activated carbon is a carbonic material that has many fine interconnected openings and typically has a wooden (coconut shells and sawdust), oil or coal base. It can be classified, based on its shape, into powdery activated carbon and granular activated carbon. Many activated carbon products are commercially available, offering the specific advantages of their individual materials. Mercury and other heavy metals, as well as organic substances, adsorb on activated carbon (Bansal, 2005). Zeolites are naturally occurring silicate minerals that can also be produced synthetically. Zeolites, in particular clinoptilolite, have a strong affinity for heavy metal ions where the adsorption mechanism is an ion exchange (Chojnacki et al., 2004). Ion exchange resins have proven to be useful in removing mercury from aqueous streams, particularly at mercury concentrations in the order of 1 µg/L to 10 µg/L. Ion exchange applications usually serve to treat mercuric salts, such as mercuric chlorides, that are found in wastewaters. The ion exchange process involves suspending a medium, which can be either a synthetic resin or mineral, into a solution where suspended metal ions are exchanged onto the medium. The anion exchange resin can be regenerated with strong acid solutions, but this is difficult since the mercury salts are not highly ionized and are not readily cleaned from the resin. The resin would therefore have to be disposed of. In addition, organic mercury compounds do not ionize, so they are not easily removed by using conventional ion exchange. If a selective resin is used, the adsorption process is usually irreversible and the resin should be disposed of as hazardous waste in a disposal facility not leading to recovery (Amuda, 2010).
3. Chelating resin is an ion exchange resin that has been developed as a functional polymer that selectively catches and removes ions, including various metal ions, from solutions. It is made of a polymer-based three-dimensional mesh construction, with a functional group that chelate‑combines metal ions. The most commonly used chelating resin is polystyrene, followed by phenolic plastic and epoxy resin. Chelating resins are used to treat plating wastewater to remove mercury and other heavy metals remaining after neutralization and coagulating sedimentation, or to collect metal ions by adsorption from wastewater whose metal-ion concentration is relatively low. Chelating resin of the mercury adsorption type can effectively remove mercury in wastewater (Chiarle and Ratto, 2000).

(v) Distillation of mercury – purification

1. Following waste treatment, collected mercury is purified by successive distillation (USEPA, 2000). High purity mercury is produced by distillation through many steps, with each step enabling a higher purity grade to be achieved. High purity mercury is required for many uses of mercury, or if the mercury is to be stored for a number of years, since high purity will help to prevent chemical reactions between the container and impurities.

2. Operations not leading to recovery of mercury or mercury compounds

1. Before mercury wastes undergo final disposal in accordance with operations D5 and D12, they should be treated so that they meet the acceptance criteria of disposal facilities (see subsection III.G.2 (b) and (c) below). Wastes consisting of mercury or mercury compounds should be stabilized and solidified before final disposal and final disposal should be carried out in accordance with national and local laws and regulations. It is noted that a number of methods for final disposal are now available, e.g., the methods referred to in paragraphs 214 and 219. Treatment operations prior to D5 and D12 operations fall under operation D9 (see subsection III.G.2 (a) below).

(a) Physico-chemical treatment (D9)

**(i) Stabilisation and solidification (S/S)**

1. Stabilisation processes include chemical reactions that may change the hazardous characteristics of waste by reducing the mobility and sometimes the toxicity of the waste constituents. Solidification processes only change the physical state of waste (e.g., converting a liquid into a solid) through the use of additives without changing the chemical properties of the waste (European Commission, 2000). The combination of these processes is referred to as S/S, or sometimes ~~referred to~~ as “conversion” ~~that~~, which means the chemical transformation of the physical state of mercury from a liquid state to mercury sulphide or a comparable chemical compound that is equally or more stable and equally or less soluble in water and that presents no greater environmental or health hazard than mercury sulphide (European Union, 2017).
2. S/S is applied, for example, to waste consisting of mercury or mercury compounds and mercury-contaminated waste such as soil, sludge, ash and liquid. S/S reduces the mobility of contaminants in the waste media by physically binding them within a stabilized mass or by inducing chemical reactions that may reduce their solubility or volatility, or both, thereby reducing their mobility (USEPA, 2007b). In the United States, only low-concentration mercury wastes (i.e., those with a mercury content lower than 260 mg/kg) can be stabilized, after which they may be landfilled.
3. Stabilization induces chemical reactions between a stabilizing agent and waste contaminants to reduce the mobility of the contaminants, and solidification involves physically binding or enclosing contaminants within a stabilized mass. Solidification is used to encapsulate or absorb waste, and forms a solid material when free liquids other than mercury are present in the waste. Waste can be encapsulated in two ways: microencapsulation and macroencapsulation. Microencapsulation is the process of mixing the waste with an encasing material before solidification occurs. Macroencapsulation refers to the process of pouring an encasing material over and around a waste mass, thus enclosing it in a solid block (USEPA, 2007b).
4. Generally speaking, the solidification process involves mixing soil or waste with binders such as Portland cement, sulphur polymer cement (SPC), sulphide or phosphate binders, cement kiln dust, polyester resins or polysiloxane compounds to create a slurry, a paste, or another kind of semi-liquid substance, which is then cured into a solid form over a period of time (USEPA, 2007b).
5. There are two main chemical approaches to the S/S process that can be applied to mercury wastes (Hagemann, 2009):
   1. Chemical conversion to mercury sulphide;
   2. Amalgamation (formation of a solid alloy with suitable metals).
6. In both approaches, the risk of mercury volatility and leachability can be reduced to an acceptable level if the conversion rate to mercury sulphide (percentage of reacted mercury) reaches near 100 per cent. If a sufficiently high rate is not achieved, the probability of mercury volatility and leachability remains high, as is the case with amalgams (Mattus, 1999).
7. While appropriate technologies for S/S of wastes consisting of mercury can reduce releases of mercury to the environment, the long-term effectiveness of these technologies has not been sufficiently studied. Therefore, the collection and analysis of information and data on such effectiveness are necessary.
8. When wastes consisting of mercury or mercury compounds that have gone through the S/S process are disposed of, special consideration should be given to the combination of S/S and methods for the final disposal of such wastes. Mercury sulphide is thermally decomposed at fire temperatures and can be oxidised by atmospheric oxygen at approximately 250-300 °C to gaseous mercury and sulphur dioxide. Therefore, measures to prevent fire incidents and release of gaseous mercury should be considered at the stage of planning disposal facilities.

#### Stabilization and solidification as mercury sulphide

1. One of the most important and well investigated mercury S/S approaches is to produce mercury sulphide (HgS), which is much less soluble and has lower volatility than most mercury compounds and is therefore less mobile in the environment. Mercury is mixed with elemental sulphur or other sulphur-containing substances to form HgS. The production of HgS can result in two different types of HgS, alpha-HgS (cinnabar) and beta-HgS (meta-cinnabar). Pure alpha-HgS is intensively red and has a slightly lower water solubility compared to pure beta-HgS, which is black in colour. HgS is a powder with a density of 2.5-3 g/cm³.
2. HgS can be produced by blending mercury and sulphur under ambient conditions for a certain period of time, until HgS is produced. To start the reaction process, a certain activation energy is required and may be provided by forcefully mixing the blend. Among other factors, high shear rates and temperatures during the process support the production of alpha-HgS, while a longer process time favours the creation of beta-HgS. Excessively long milling in the presence of oxygen can lead to the production of mercury (II) oxide (HgO). Because HgO has higher water solubility than HgS, its creation should be avoided through milling under inert atmospheric conditions or through the addition of an antioxidant (e.g., sodium sulphide). Since the reaction between mercury and sulphur is exothermic, an inert atmosphere also contributes to making the operation safe. The conversion process is robust and relatively simple to carry out, but strict controls are needed to prevent volatile losses of mercury during conversion. In addition, treatment residue should be tested to ensure the completeness of the conversion to mercury sulphide.
3. Mercury sulphide can also be formed by creating a reaction between mercury and sulphur in the vapour phase. Reaction of mercury with sulphur in a sealed vessel at elevated temperature and pressure can also form the alpha cinnabar form of mercury sulphide (United States Patent: US 7691361 B1, April 10, 2010). This process is commercially available in the U.S.[[69]](#footnote-111), although U.S. waste disposal regulations currently prohibit landfilling of most high concentration mercury waste.
4. HgS is very insoluble in water and has very low volatility. Although HgS is chemically very stable and non-reactive, exposure to ambient environmental conditions will result in its conversion to other mercury compounds over time. The isolation of HgS from the environment through encapsulation and disposal in a specially engineered landfill or permanent underground storage may therefore be necessary to ensure that it does not convert into other mercury compounds. In addition, the fact that dissolved organic matter and high chloride concentrations in leachate increase mercury releases from HgS (Waples et al., 2005; Science Applications International Corporation, 2002) suggest that mercury converted into HgS should be disposed of in ways that ensure the waste has no contact with water or other types of waste, especially those containing organic matter and chloride. Moreover, since micro-organisms inhabiting acid mine drainage systems, which are dominated by Fe-oxidizing and S-oxidizing bacteria, in the microcosm with metacinnabar (beta-HgS) increase dissolved Hg concentrations (Jew et al., 2014), exclusion or at least inhibition of the influence of such microorganisms in specially engineered landfills and permanent underground storage may be required for disposal of mercury treated as HgS.
5. As a fine powdery material, the handling of HgS is subject to specific requirements, including solidification, in order to avoid, for example, the risk of dust releases. The solidification process leads to an increase in volume by approximately 300 per cent and in weight by approximately 16 per cent, based on molecular weight compared to mercury. For solidification of mercury sulphide, materials with low alkali content should be used as a recent study indicates that mercury release from mercury sulphide increases when pH value of eluate exceeds 10 (Mizutani, Kadotani, and Kanjo, 2010).
6. Large-scale S/S processes for waste consisting of mercury that use sulphur compounds to form HgS are also available[[70]](#footnote-113):
   1. One process takes place in a vacuum mixer operated in inert vacuum atmosphere which ensures good process control and safe operation. The mixer is operated batch-wise. A dust filter and an activated carbon filter prevent releases from the facility. The reaction between mercury and sulphur takes place at a stoichiometric ratio. The end product consists of red mercury sulphide. The end product is thermodynamically stable up to 350 °C[[71]](#footnote-114); and
   2. Another process consists in a wet chemical reaction. The active sulphur of a polysulfide solution reacts with the metallic mercury to become HgS. The benefits of using a wet chemical process over a dry powder process are that it mitigates the chances of HgO (mercuric oxide) forming without the need for an inert gas environment and without the explosion risks associated with processing dry powders. The wet chemistry reaction is conducted at a low temperature (60-70 °C). The chemistry should be carefully controlled to stop unwanted toxic compounds forming. The conversion rate of Hg to HgS is greater than 99.999 per cent with a remaining metallic mercury concentration in the output of less than100 mg/kg[[72]](#footnote-115).

#### Sulphur polymer stabilization and solidification (SPSS)[[73]](#footnote-118), [[74]](#footnote-119)

1. The sulphur polymer stabilization and solidification (SPSS)[[75]](#footnote-121) process involves sulphur stabilization followed by solidification, with the advantage of a lower chance of mercury vaporization and leaching because the final product is monolithic with a low surface area. The process involves two steps: in the first step, mercury is stabilized with sulphur to form beta-HgS (meta-cinnabar dust) (López et al., 2010; López-Delgado et al., 2012); in the second step, the beta-HgS is incorporated and microencapsulated in a polymeric sulphur matrix at 135°C to obtain a fluid that is cooled to room temperature, in moulds, to obtain solid blocks (monoliths). The second step is the creation of a barrier to prevent mercury releases to the environment, thereby minimizing the possibility of HgS converting into other forms of mercury by reducing its contact with the environment. The SPSS process has low energy consumption, entails low mercury emissions, requires no water, has no effluents and generates no wastes other than HgS. Controls should be in place to prevent and emissions of mercury and to ensure safe conditions for workers and for the environment, including engineering controls to prevent possible fires and explosions.
2. A relatively high Hg load of the monolith (about 70 per cent) can be achieved with this process. The process is robust and relatively simple to implement and its final product is very insoluble in water, has a high resistance to corrosive environment, is resistant to freeze-thaw cycles and has a high mechanical strength. The SPSS technology can be directly applied to waste consisting of mercury with different degrees of purity, without the need for previous distillation, and to a wide range of mercury-containing wastes without previous treatment of such wastes (López et al., 2010, López-Delgado et al., 2012, López et al., 2015). All the final stabilized-microencapsulated products (from metallic mercury, zinc waste, aluminum waste and fluorescent lamp dust) are compact solids that have a stability and resistance similar to that of concrete. The SPSS process therefore ensures the complete immobilization of mercury, rendering it impermeable and giving it extremely low porosity, thereby minimizing the risk of mercury releases to the environment. The final products are hard monolithic blocks, the shape of which can be adapted to facilitate transport.
3. Monolith samples (of 40 x 40 x 160 mm) have been tested for leaching in accordance with European standards TS 14405 (CEN, 2017) and EN-12457-4 (CEN, 2002a), with tests involving a dynamic flow percolation assay on encapsulate monoliths and a leaching by agitation assay on granular material obtained through monolith crushing. All mercury concentrations found in the leachates, for a liquid/solid ratio of 10 l/kg, were <0.01 mg/kg, so the monoliths met the European Union criterion for the acceptance of waste into landfills for inert waste (<0.01 mg/kg, as per Council Decision 2003/33/EC, establishing criteria and procedures for the acceptance of waste at landfills (European Union, 2003).
4. Another example of a similar technology is solidification of beta-HgS with modified sulphur. The first step is to form beta-HgS by mixing mercury with a purity of 99.9 per cent or higher with sulphur powder, and the second step involves solidifying the beta-HgS with modified sulphur by mixing the two substances for one hour and subsequently heating the mix to 130°C for one hour. Results of a Japanese leaching test (JLT-13) of solidified mercury sulphide show that the product’s leaching rate ranges from 0.0009 to 0.0018 mg/L, which is below the elution test standard (0.005 mg/L) (Committee on consideration of environmentally sound management of mercury waste, 2014).

**Stabilization and solidification with sulphur microcements[[76]](#footnote-123)**

1. The treatment of mercury wastes with sulphur microcements is another S/S technology. Application of the technology results in a solid matrix that ensures the confinement of mercury because of its precipitation in the form of very insoluble oxides, hydroxides and sulfides. The technology is commercially available and has been tested in wastes with low levels of mercury contamination (Hg ≤ 2 per cent by weight).
2. Once the contaminated material to be treated has been characterized, the amount and type of microcement suitable for the application is determined. Microcements must have the following characteristics so that an adequate level of stabilization and microencapsulation of the mercury contained in contaminated materials can be achieved:

(a) They must be inorganic and all particles are less than a certain size (a few microns);

(b) They must contain mercury stabilising components, such as alkali sulphides;

(c) They must have very high mechanical properties, so that volatilization and leaching of mercury can be avoided;

(d) They should have a blast furnace slag percentage of over 60 per cent, a Portland clinker component C3A content lower than 3 per cent and an alkali content below 0.6 per cent.

1. The process involves mixing of mercury-contaminated waste with the selected sulphur microcement and with water; the mixture is then discharged into the desired mould and matured over a 24 - 48 hour period in watertight and leak-protected areas. The final product can take different forms; the ones with fewer exposed surfaces, such as large cubic blocks, are recommended for the most contaminated waste.
2. The sulphur microcement treatment technology has been tested on, among others, mercury waste obtained from the dredging of contaminated sludge from the Flix dam in the province of Tarragona in Spain. The final products ensure a high level of strength and durability and enable a safe handling and transport. The leaching values following tests according to standard EN 12457-4 (CEN, 2002a) with a water/solid ratio of 10/1 are below 0,003 mg/kg, well below the EU acceptance criteria for landfills for inert solid wastes (<0.01 mg/kg, as per Decision 2003/33/EC). It is an inert product, and has high strength and durability enabling a safe mechanical handling and transport.

#### Amalgamation

1. Amalgamation is the dissolution and solidification of mercury in other metals such as copper, nickel, zinc and tin, resulting in a solid, non-volatile product. It is a subset of S/S process. Two generic processes are used for amalgamating mercury in wastes: aqueous and non‑aqueous replacement. The aqueous process involves mixing a finely divided base metal such as zinc or copper into a wastewater that contains dissolved mercury salts; the base metal reduces mercuric and mercurous salts to mercury, which dissolves in the metal to form a solid mercury-based metal alloy called amalgam. The non-aqueous process involves mixing finely divided metal powders into waste mercury, forming a solidified amalgam. The aqueous replacement process is applicable to both mercury salts and mercury, while the non-aqueous process is applicable only tomercury. However, mercury in the resultant amalgam is susceptible to volatilization and leaching. Therefore, amalgamation is typically used in combination with an encapsulation technology, although based on concerns regarding possible volatilization and leaching, it should not be considered as a first option for treating wastes consisting of mercury (USEPA, 2007b).

#### (ii) Soil washing and acid extraction

1. Soil washing is an *ex situ* treatment of soil and sediment contaminated with mercury. It is a water-based process that uses a combination of physical particle size separation and aqueous-based chemical separation to reduce contaminant concentrations in soil. The process is based on the concept that most contaminants tend to bind to the finer soil particles (clay and silt) rather than the larger particles (sand and gravel). Physical methods can be used to separate the relatively clean larger particles from the finer particles because the finer particles are attached to larger particles through physical processes (compaction and adhesion). This process thus concentrates the contamination bound to the finer particles for further treatment. Acid extraction is also an *ex situ* technology that uses an extracting chemical such as hydrochloric acid or sulphuric acid to extract contaminants from a solid matrix by dissolving them in the acid. The metal contaminants are recovered from the acid leaching solution using techniques such as aqueous-phase electrolysis. More detailed information can be found in *Treatment technologies for mercury in soil, waste, and water* (USEPA, 2007b).

(b) Disposal in specially engineered landfills (D5)

1. Wastes consisting of mercury or mercury compounds that result from S/S that meet the acceptance criteria for specially engineered landfills defined by national or local regulations may be disposed of in specially engineered landfills. Additional measures should be taken at such landfills to minimize releases and the methylation of mercury, for instance through the prevention of rainwater and groundwater inflows, the prohibition of mixing of different kinds of waste at the landfill site, the maintenance of records of waste amounts and areas landfilled, the collection of leachate and the   
   long-term monitoring of releases of mercury and methylmercury from landfill sites into, e.g., air and groundwater (UNEP, 2022b).
2. When wastes consisting of mercury or mercury compounds that result from S/S are disposed of in a specially engineered landfill, special consideration should be given to thermal decomposability of such wastes. The final cover of a landfill may be permeable to air in the long term. Mercury sulphide can then come into contact with atmospheric oxygen and become oxidised to mercury and sulphate. The formation of methylmercury may occur under certain geochemical conditions. Both gaseous mercury and methylmercury can leave the landfill via the gas circuit (landfill gas) (German Federal Environment Agency, 2014).
3. In addition to fire prevention, methods to stabilize and solidify wastes consisting of mercury or mercury compounds as well as the structure and final cover of the landfill accepting them should respond to mechanism referred to in the preceding paragraphs to ensure mercury releases from the disposed wastes are be minimized.
4. Box 10 provides an example of disposal of wastes consisting of mercury in Japan.

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| **Box 10 Country Example: Japan**  As an option for the disposal of wastes consisting of mercury, Japan has identified specific treatments and types of landfills that can be combined to dispose of such wastes, namely: 1) stabilization of mercury as mercury sulphide (HgS), followed by solidification (e.g., through the use of modified sulphur as described in paragraph 222) and disposal in a leachate-controlled type of landfill for industrial wastes, with additional measures taken to minimize releases and methylation of mercury, for instance through the prevention of rainwater and groundwater inflow, the prohibition of mixing of different types of waste at the landfill, and the maintenance of records of waste amounts and areas landfilled; and 2) stabilization of mercury as HgS followed by solidification and disposal in an isolated type of landfill for hazardous industrial wastes (see figure below). Further specifications on these combinations will be determined based on further experiments and studies (Ministry of the Environment of Japan, 2015).  **Figure:** Landfill for hazardous industrial wastes (isolated type) (Ministry of the Environment of Japan, 2007a) |

1. It should be noted that wastes of products containing mercury or mercury compounds may be disposed in engineered landfills for household wastes despite best efforts to keep them separate. For information on methods to reduce mercury releases from these landfills, see section H.2 below.
2. Boxes 11 to 14 provide examples of such criteria in the European Union, China, Japan, and USA. In addition, disposal of certain mercury wastes in landfills is banned in some countries.

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| **Box 11 Regional Example: European Union**  Under European Union legislation, only wastes with leaching limit values of 0.2 and 2 mg Hg/kg dry substance at a liquid‑solid ratio of 10 l/kg can be accepted into landfills for non-hazardous and hazardous wastes, respectively (European Union, 2003). |

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| **Box 12 Country Example: China**  Under Chinese legislation, only hazardous wastes with total mercury leaching value equal to or less than 0.12 mg/l (Chinese Leaching Test Method: HJ/T 299) can be accepted into landfills for hazardous wastes (MEE China, 2019), and non-hazardous wastes with total mercury leaching value equal to or less than 0.05 mg/l (Chinese Leaching Test Method: HJ 557) can be accepted into non-hazardous industrial solid waste storage and landfill facilities (GB 18599) (MEE China, 2020). |

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| **Box 13 Country Example: Japan**  Under Japanese legislation, treated wastes with mercury concentration equal to or less than 0.005 mg/l (Leaching Test Method: Japanese Standardized Leaching Test No. 13 (JLT‑13) (Ministry of the Environment Notification No. 13) can be accepted into landfills for industrial wastes (leachate-controlled type), and those with mercury concentration in excess of 0.005 mg/L should be disposed of at landfills for hazardous industrial wastes (isolated type) (see figure in Box 10) (Ministry of the Environment of Japan, 2007b). |

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| **Box 14 Country Example: USA**  Under United States mercury waste treatment regulations, only low concentration mercury wastes can be treated and landfilled (high concentration waste must be retorted for mercury recovery). Treated mercury waste must leach less than 0.025 mg/l mercury (by TCLP testing) to be accepted for landfill disposal (USEPA, undated). |

1. For further information on specially engineered landfills, see Basel Convention *Technical guidelines on the environmentally sound disposal of hazardous wastes and other wastes in specially engineered landfill* (D5) (UNEP, 2022).

(c) Disposal in permanent storage (underground facilities) (D12)

1. After having been solidified or stabilized, where appropriate, mercury wastes that meet the acceptance criteria for permanent storage (disposal operation D12) may be permanently stored in special containers in designated areas in underground storage facilities such as salt mines. Box 14 provides an example of national provisions for permanent storage.

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| **Box 14 Regional Example: European Union**  The *Regulation (EU) 2017/852 of the European Parliament and of the Council of 17 May 2017 on mercury, and repealing Regulation (EC) No 1102/2008* (European Union, 2017) provides that mercury waste that underwent conversion and, if applicable, solidification shall only be permanently disposed of in the following permanent storage facilities licensed for disposal of hazardous waste:  (a) salt mines that are adapted for the permanent storage of mercury waste that underwent conversion, or deep underground hard rock formations providing a level of safety and confinement equivalent to or higher than that of such salt mines; or  (b) above-ground facilities dedicated to and equipped for the permanent storage of mercury waste that underwent conversion and solidification and that provide a level of safety and confinement equivalent to or higher than that of the facilities referred to in point (a). |

1. The technology for underground storage is based on mining engineering, which uses technologies and methodologies to excavate mining areas and construct mining chambers[[77]](#footnote-126). Decommissioned mines could be used for the permanent storage of solidified and stabilized waste once they have been evaluated and specifically adapted for that purpose.
2. In addition, the principles of and experience obtained with underground disposal of radioactive waste can be applied to the underground storage of mercury wastes. While the excavation of deep underground repositories can be achieved using standard mining or civil engineering technologies, such operations can only be conducted when locations are accessible (e.g., below surface or nearshore locations), rock units are reasonably stable and there is no major groundwater flow, and the excavation depth ranges between 250 m and 1,000 m. At a depth greater than 1,000 m, excavations become increasingly technically difficult and correspondingly expensive (World Nuclear Association, 2020).
3. The following publications, among others, contain detailed information on permanent underground storage for mercury wastes:
   1. European Union, 2003. “Safety Assessment for Acceptance of Waste in Underground Storage”,Appendix A to *Council Decision 2003/33/EC of 19 December 2002 establishing criteria and procedures for the acceptance of waste at landfills pursuant to Article 16 of and Annex II to Directive 1999/31/EC*;
   2. BiPRO, 2010. *Requirements for Facilities and Acceptance Criteria for the Disposal of Metallic Mercury*;
   3. International Atomic Energy Agency (IAEA), 2009. *Geological Disposal of Radioactive Waste: Technological Implications for Retrievability*;
   4. World Nuclear Association, 2020. *Storage and Disposal of Radioactive Waste*.
4. Permanent storage in facilities located underground in geohydrologically isolated salt mines and hard rock formations is an option for separating hazardous wastes from the biosphere for geological periods of time. A site-specific risk assessment in accordance with pertinent national legislation, such as the provisions regarding safety assessment for acceptance of waste in underground storage contained in Appendix A to the Annex to Council Decision 2003/33/EC (European Union, 2003), which establishes criteria and procedures for the acceptance of waste at landfills, should be performed for every planned underground storage facility.
5. Wastes should be disposed of in a manner that (a) excludes any undesirable reaction between different types of wastes and between stored wastes and storage linings; and (b) prevents the release and transport of hazardous substances. Operational permits should define the types of wastes that should be generally excluded. Wastes should be isolated through a combination of engineered barriers and natural barriers (rock, salt, clay), also known as a “multi-layer” approach to waste disposal. Facilities should be checked or monitored periodically to assure that the containment remains secure and stable. This is often termed a multi-barrier concept, with the waste packaging, the engineered repository and the geology all providing barriers to prevent any mercury leakage from reaching humans and the environment (BiPRO, 2010; European Union, 2003; IAEA, 2009; World Nuclear Association, 2020).
6. Specific factors that can affect the behaviour of mercury in the host rocks and geological formations used for permanent storage, such as the layout of storage facilities, types of containments used, storage location and conditions, monitoring, site access conditions, storage closure strategy, sealing and backfilling and depth of storage facilities, need to be considered separately from the properties of the wastes to be stored and the storage system to be used. Potential host rocks for permanent storage of mercury wastes include salt rock and hard rock formations (igneous rocks such as granite, metamorphic rocks, and gneiss sedimentary rocks such as limestone or sandstone) (BiPRO, 2010; European Union, 2003; IAEA, 2009; World Nuclear Association, 2020).
7. The following issues should be borne in mind when selecting a permanent underground storage site for the disposal of mercury wastes:
   1. Caverns or tunnels used for storage should be completely separated from active mining areas and areas that may be reopened for mining;
   2. Caverns or tunnels should be located in geological formations that are well below zones where groundwater is present or in formations that are completely isolated from water-bearing zones by impermeable rock or clay layers;
   3. Caverns and tunnels should be located in geological formations that are extremely stable and not in areas subject to earthquakes.
8. In order to guarantee the complete inclusion of wastes in permanent storage facilities, the disposal mine and any area around it that might be affected (geomechanically or geochemically) by disposal operations should be surrounded by a host rock (known as an “isolating rock zone”) of sufficient thickness and homogeneity, with suitable properties and at a suitable depth (see figure 4). As a basic principle, a long-term risk assessment should enable parties to prove that the construction and operational and post-operational phases of underground disposal facilities do not lead to any degradation of the environment. Consequently, appropriate models must be used to analyse and assess all technical barriers (e.g., waste forms, backfilling, sealing measures), the behaviour of the host and surrounding rock, overburden rock formations and the sequence of possible events in the overall system.

**Figure 4:** Concept of complete inclusion (schematic diagram) (courtesy: GRS)



1. If the host rock under consideration shows any deficiencies (e.g., insufficient homogeneity or thickness), a multi-barrier system can compensate for the missing or inadequate barrier properties of the rock. In general, a multi-barrier system of this kind is composed of one or several additional barrier components (see table 8 and figure 5) that can help to achieve the ultimate goal, i.e., to durably isolate stored wastes from the biosphere.
2. A long-term safety assessment (see above) should be conducted to ascertain the need for a multi-barrier system, and the mode of action of barrier components, within the disposal system. By way of example, the geological formation(s) overlaying a disposal mine (“overburden”) may be effective in:
   1. Protecting the underlying host rock from any impairments of its properties; and/or
   2. Providing additional retention capacities for contaminants that might be released from the disposal mine under certain circumstances.

**Table 8:** Possible components of a multi-barrier system and examples for their modes of action

|  |  |
| --- | --- |
| **Barrier component** | **Example for mode of action** |
| Waste content | Reduction of the total amount of contaminants to be disposed of |
| Waste specification | Treatment of waste in order to get a less soluble contaminant |
| Waste canister | Bridging of a limited time period until natural barriers become effective |
| Backfill measures | Backfilling of void mine spaces to improve geomechanical stability and/or to provide special geochemical conditions |
| Sealing measures | Shaft sealing should provide the same properties where the natural barrier(s) is(are) disturbed by mine-access |
| Host rock | Complete inclusion of contaminants (in ideal cases) |
| Overburden | Provision of an additional natural (geological) barrier, e.g., through overlaying a clay layer with sufficient thickness and suitable properties |

**Figure 5:** Main components of a multi-barrier system and their layout within the system (schematic diagram) (courtesy: GRS)

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1. In general, an underground disposal system including all the criteria, requirements and final layout described above should be designed according to waste-specific and site-specific criteria, taking into consideration all relevant regulations (e.g., European Union, 2003). To give readers a rough idea of the depth and thickness of different types of host rocks suitable for underground storage, table 9 lists typical acceptable dimensions, based on past experience and current plans.

**Table 9:** Typical values of vertical thickness of host rock body and potential disposal depth (Grundfelt et al., 2005)



H. Reduction of mercury releases from thermal treatment and landfilling of waste

1. Reduction of mercury releases from thermal treatment of waste

1. Mercury waste may be combusted with MSW because mercury waste may be contained in MSW despite best efforts to keep them separate. By reason of its low boiling point, almost all the mercury in the waste may be transferred to combustion gas and, to a lesser extent, bottom ash. Most of the mercury in the combustion gas fed into waste combustion units is elemental mercury and transforms to divalent mercury after passing through the units, and part of the divalent mercury transfers to fly ash. The divalent mercury is assumed to be mercuric chloride; consequently, flue gas treatment devices that can effectively remove mercuric chloride and mercury should be selected. In addition, waste that potentially contains or is contaminated with mercury, such as poorly segregated waste from healthcare facilities, should not be incinerated in incinerators that lack flue gas treatment devices (Arai et al., 1997). Emission and effluent standards for mercury should be set and mercury levels of treated flue gas and wastewater should be monitored to ensure that mercury releases into the environment are kept to a minimum. Such practices should also be applied in other thermal waste treatment processes such as vacuum-sealed roasting facilities.
2. Primary techniques for reducing the inclusion of mercury in the waste stream include the following (European Commission, 2019):
   1. Efficient removal of products containing mercury or mercury compounds from the waste stream (e.g., through separate collection of certain types of batteries and dental amalgam (using amalgam separators) before mercury wastes are co-mingled with other wastes or wastewaters;
   2. Notification of waste producers of the need to segregate mercury;
   3. Identification and/or restriction of receipt of potential mercury wastes;
   4. When mercury wastes have been knowingly received, control of feeding of such wastes into abatement systems in order to avoid overloading system capacities.
3. Secondary techniques for preventing mercury releases to the air from the waste stream include treatment of flue gas. Box 15 provides an example of emission limit values for waste incineration plants.

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| --- |
| **Box 15 Regional Example: European Union**  The European Commission Implementing Decision 2019/2010 establishing BAT conclusions for waste incineration (European Commission, 2019) sets BAT-associated emission levels for discharges of wastewater from the cleaning of flue (waste) gases and air emission limit values for waste incineration plants. Regarding the former, unfiltered samples must not contain more than 10µg/L mercury and its compounds, expressed as mercury (Hg); as for the latter, air emissions must not exceed 20µg/Nm3 mercury and its compounds, expressed as mercury (Hg) for continuous monitoring (daily average or average over the sampling period), or 10µg/Nm3 for a long-term sampling period. |

1. The selection of a process to control mercury flue gas emissions depends on the chlorine and SO2 concentrations (Vosteen, Kanefke and Köser, 2006) in the flue gas. When the chlorine concentration is high, the mercury contained in resultant crude flue gas may tend to be in the oxidized form depending on the SO2 concentration (which consumes Cl2), in which case it can be captured in the flue gas cleaning system. The addition of reagents for a more efficient removal of mercury can be necessary and provides a means for removing mercury from the process. It should be noted that, in the incineration of sewage sludge, mercury emissions will consist mostly of elemental mercury, due to the lower chlorine content of such sludge compared to that of municipal or hazardous waste. Consequently, special attention should be paid to capturing these emissions. Elemental mercury can be removed by transforming it into oxidized mercury; this can be done by adding oxidants to the mercury and then depositing the resultant mix in a scrubber or directly on sulphur doped activated carbon, hearth furnace coke, or zeolites depending on the flue gas cleaning system (dry, wet or semi-dry). The removal of heavy metals, including mercury, from wet scrubber systems can be achieved through flocculation, a process in which metal hydroxides are formed under the influence of flocculating agents (poly-electrolytes) and FeCl3. For the removal of mercury, complex-builders and sulphides (e.g., Na2S and Tri-Mercaptan) are added.
2. The removal of mercury in oxidised form from flue gas can be achieved through adsorption on activated carbon reagents in an entrained flow system in which activated carbon is injected into the gas flow and filtered from the gas flow with the use of bag filters. Activated carbon has demonstrated to have a high adsorption efficiency for mercury in oxidised form, as well as for dioxins and furans (PCDDs/PCDFs). Different types of activated carbon have different adsorption efficiencies, presumably due to the nature of various carbon particles and the influence exerted on the particles by manufacturing processes (European Commission, 2019a). Almost all emission relevant flue-gas components, in particular residual contents of hydrochloric acid, hydrofluoric acid, sulphur oxides and heavy metals (including mercury in oxidised form), can effectively be deposited in static bed filters of grained hearth furnace coke (HFC), a fine coke of 1.25 mm to 5 mm. The depositing effect of HFC is essentially based on mechanisms of adsorption and filtration. In general, incinerators should be equipped with flue gas treatment devices to capture NOx, SO2, particulate matter, mercury vapour and particulate-bound mercury as a co‑benefit. Powdered activated carbon injection is one of the advanced technologies used for mercury removal in incinerators and coal-fired power plants. Mercury adsorbed on activated carbon can be stabilized or solidified for disposal (see subsection III.G.2 (a) above).
3. For the reduction of mercury emissions from waste incineration, the following documents provide additional technical information:
   1. UNEP, 2022. Technical Guidelines on the environmentally sound incineration of hazardous wastes and other wastes as covered by disposal operations D10 and R1;
   2. Minamata Convention, 2011. *Guidance on Best Available Techniques and Best Environmental Practices* (Chapter VI);
   3. UNECE,1998. *Heavy Metals Protocol* to the *Long-range Transboundary Air Pollution* and 2013 *Guidance document on best available techniques for controlling emissions of heavy metals and their compounds from the source categories listed in annex II*;
   4. UNEP, 2010. *Study on mercury sources and emissions and analysis of cost and effectiveness of control measures:* *“UNEP Paragraph 29 study”* (doc. UNEP(DTIE)/Hg/INC.2/4);
   5. European Commission, 2019a. *Reference Document on the Best Available Techniques for Waste Incineration*;
   6. European Commission, 2019b. *Commission Implementing Decision (EU) 2019/2010 of 12 November 2019 establishing the best available techniques (BAT) conclusions, under Directive 2010/75/EU of the European Parliament and of the Council, for waste incineration*;
4. For the reduction of mercury emissions from waste co-incineration, the following documents provide additional technical information:
   1. European Commission, 2017a. *Best Available Techniques (BAT) Reference Document for Large Combustion Plants*;
   2. European Commission, 2017b. *Commission Implementing Decision (EU) 2017/1442 of 31 July 2017 establishing the best available techniques (BAT) conclusions, under Directive 2010/75/EU of the European Parliament and of the Council, for large combustion plants*;
   3. European Commission, 2013b. *Commission Implementing Decision of 26 March 2013 establishing the best available techniques (BAT) conclusions under Directive 2010/75/EU of the European Parliament and of the Council on industrial emissions for the production of cement, lime and magnesium oxide.*
5. When a wet scrubber is used as a flue gas treatment method, it is essential that wastewater from the wet scrubber be treated.

2. Reduction of mercury releases from engineered landfills for household wastes

1. Mercury waste may be deposited in engineered landfills for household waste despite best efforts to keep them separate. In such cases, there are three types of pathways through which mercury can be released into the environment: the working face of the landfill, leachate, and landfill gas. The most important sites of mercury emissions are landfill working faces and methane vents (Lindberg and Price, 1999).
2. Landfill cover should be applied every day to reduce the direct release of mercury from wastes that have been newly added to landfills (Lindberg and Price, 1999). Landfill fires can also result in increased mercury releases. For prompt application of soil cover in case of landfill fires, soil cover materials and machines used for applying soil cover for the purpose of extinguishing fires (e.g., dump truck, dozer shovel) should be readily available.
3. It is reported that mercury releases through leachate are fairly minimal compared to releases through landfill gas depending on the management practices of the landfill (Yanase, Hirato, and Matsufuji, 2009; Takahashi et al., 2004; Lindberg et al., 2001). Mercury transferred to leachate can be removed through leachate collection and treatment, as is the case for wastewater from wet scrubbers of waste incinerators.
4. A landfill gas capture system should be installed at the site to capture mercury vapour and if applicable methylmercury and thereby prevent their release into the atmosphere if appropriate.

I. Management of contaminated sites

1. Mercury-contaminated sites are widespread around the world. There is a range of possible sources of site contamination, including mercury storage, manufacturing of products containing mercury or mercury compounds, use of mercury in manufacturing processes, ASGM activities using mercury or primary ore rich in mercury in which the mercury is mobilized, primary mercury mining and abandoned, historical mines not managed in accordance with modern practices, point sources of emissions and releases, and waste treatment and disposal (UNEP, 2019d).
2. For guidance on the management of contaminated sites, see the *Guidance on the management of contaminated sites* adopted by the Conference of the Parties to the Minamata Convention in decision MC-3/6 (UNEP, 2019d). It provides guidance on managing sites contaminated with mercury or mercury compounds, from site identification and detailed site investigation to the decision process for site management and, where appropriate, remediation.
3. When decommissioning mercury cell plants, a plan for the remediation of contaminated   
   chlor-alkali sites should be developed (European Commission, 2014b).

J. Health and safety

1. Employers should ensure that the health and safety of every employee are protected while they are at work. Every employer should obtain and maintain insurance, under an approved policy from an authorized insurer, providing a sufficient level of coverage in case of liability (compensation) for bodily illness or injury sustained by employees arising out of and in the course of their employment, in accordance with national law. Health and safety plans should be in place at all facilities that handle mercury wastes to ensure the protection of everyone in and around such facilities. Such plans should be developed for each facility by trained health and safety professionals with experience in managing health risks associated with mercury.
2. An exposure assessment should if appropriate be undertaken for all employees who are directly exposed to mercury or mercury compounds and appropriate monitoring and industrial hygiene practices adopted. Colorimetric badges and/or personal monitoring equipment (vapour sampling devices) are needed for this kind of comprehensive exposure assessment and monitoring programme. Pre-employment physical examinations should be carried out to establish a baseline for determining an individual’s background mercury level and to ensure that the employee has normal body chemistry for mercury removal. Personnel may have other considerations that should be handled on a   
   case-specific basis. Medical monitoring programmes should also include periodic physical exams (e.g., every one to three years), regular blood tests and regular urinalysis. Consideration should be given to offering alternative jobs to workers who become pregnant or who are breastfeeding.
3. The protection of workers who are engaged in the management of mercury wastes and the general public can be achieved through the following ways:
   1. By allowing access to facilities to authorized personnel only;
   2. By ensuring that occupational exposure limits for hazardous substances are not exceeded by making sure that all personnel use appropriate protective equipment;
   3. By ensuring appropriate ventilation of facilities to minimize risk from exposure to volatile substances or substances that can become airborne;
   4. By ensuring facility compliance with all national and regional laws on workplace health and safety.
4. Guideline values for mercury concentrations in drinking water and ambient air have established by WHO are 0.006 mg/L for inorganic mercury and 1 μg/m3 for inorganic mercury vapour (WHO, 2006; WHO Regional Office for Europe, 2000). Governments are encouraged to monitor air and water in order to protect human health, especially near sites where mercury waste management activities take place. Some countries have established permissible levels of mercury in the working environment (e.g., 0.025mg/m3 Hg for inorganic mercury, excluding mercury sulphide, and 0.01mg/m3 Hg for alkylmercury compounds in Japan); waste management operations should be conducted so as to satisfy requirements regarding permissible levels of mercury in the working environment, and facilities where such operations are conducted should be designed and operated so as to minimize mercury releases to the environment as far as is technically possible.
5. Special attention should be paid to sites where products containing mercury or mercury compounds are handled. Within the waste stream, mercury emissions from products containing mercury or mercury compounds can lead to exposures that raise health concerns and contribute to environmental releases at multiple points. Waste collectors, truck drivers and workers at transfer stations can be exposed to brief peaks of mercury vapour when handling wastes of products containing mercury or mercury compounds. Waste management employees at the “working face” of a landfill – the active area where waste is dumped, spread, compacted and buried – can be repeatedly exposed to mercury vapour. Those in the informal waste sector involved in scavenging landfills for reclaimable items can be chronically exposed. Venting points for methane gas generated by decaying organic wastes are additional sources of mercury release and exposure.
6. Disposal facilities, especially where mercury recovery processes are conducted, also present a high risk of mercury exposure to employees as well as a risk to people living and working near such facilities. Major activities presenting high risk of exposure include the crushing of fluorescent lamps, the extraction of mercury from products containing mercury such as thermometers and barometers, the thermal treatment of wastes containing or contaminated with mercury, and S/S of wastes consisting of mercury or mercury compounds.
7. Employee training in effective ESM and workplace health and safety should be provided to, among other things, ensure employee safety against mercury exposure and accidental injury when managing waste.
8. The basic knowledge that employees need includes:
   1. The definition of mercury wastes and the chemical properties and adverse effects of mercury;
   2. How to identify mercury wastes and to segregate such wastes from other types of wastes;
   3. Occupational safety standards relevant to mercury and how to safeguard their health against mercury exposure;
   4. How to use personal protective equipment, such as body coverings, eye and face protectors, gloves and respiratory protectors;
   5. Proper labelling and storage requirements, container compatibility and dating requirements, and closed-container requirements;
   6. How to safely handle mercury wastes, particularly used products containing mercury such as thermometers and barometers, using the equipment available at the facility in which they work;
   7. How to use engineering controls to minimize exposure;
   8. How to respond in an emergency if mercury in waste is accidentally spilled.
9. It is important to have worker insurance and employer liability insurance in order to be better prepared for accidents or injuries sustained by workers in the facility, as appropriate under national law.
10. A mercury awareness-raising package developed by UNEP (UNEP, 2008d) is recommended for use in employee training. All training materials should be translated into local languages and made accessible to employees.

K. Emergency response

1. Emergency response plan

1. Emergency response plans should be in place at each stage of the mercury waste processing chain (e.g., generation, storage, transport, treatment or recovery, and disposal). While emergency response plans can vary depending on the activities carried out at each stage of waste management and the physical and social conditions of each management site, the principal elements of an emergency response plan include the identification of potential hazards; compliance with legislation governing emergency response plans; specification of actions to be taken in emergency situations, including mitigation measures, personnel training plans, communication targets (e.g., fire services, police, neighbouring communities, local governments, etc.) and methods to be used in case of emergency; and specification of the method and frequency of testing of emergency response equipment.
2. When an emergency occurs, the first step is to examine the site. The person in charge should approach the site cautiously from upwind, secure the scene and identify any hazards. Placards, container labels, shipping documents, material safety data sheets, car identification charts and/or knowledgeable persons on the scene are valuable information sources. The need for site evacuation, the availability of human resources and equipment, and possible immediate actions should then be assessed. In order to ensure public safety, an emergency response agency call should be made and, as an immediate precautionary measure, spill and leak areas should be isolated for at least 50 meters in all directions. In case of fire, a suitable extinguishing agent should be used, and the use of water should be avoided. For further information, the *Emergency Response Guidebook* (U.S. Department of Transportation et al., 2012) can be consulted.

2. Special consideration for spillage of mercury or mercury compounds

1. Spillage of mercury or mercury compounds occurs primarily when wastes of products containing mercury or mercury compounds are broken. Most cases of spillage seem to involve glass thermometers, which are easily broken. Although the mercury in each glass thermometer is about   
   0.5-3 g and does not usually lead to serious health problems, all mercury spills should be considered hazardous and should be cleaned up with caution. Indoor areas should be adequately ventilated. If a person experiences discomfort following a mercury spill, a medical doctor and/or the environmental health authorities should be contacted immediately.
2. If the spill is small and simple (e.g., it occurs on a non-porous area such as linoleum or hardwood flooring, or on a porous item that can be thrown away, such as a small rug or mat), it can be cleaned up by an individual. If the spill is large or complex (e.g., it occurs on a rug that cannot be discarded, on upholstery, or in cracks or crevices), it may be necessary to hire a trained professional to contain it or clean it up. Large spills involving more than the amount of mercury found in a typical household product should be reported to the local environmental health authorities. If there is any uncertainty as to whether a spill should be classified as “large”, the local environmental health authorities should be contacted as a precautionary measure. Under certain circumstances, it may be advisable to obtain the assistance of qualified personnel for professional clean-up or air monitoring, regardless of spill size (Environment Canada, 2016).
3. Spills of mercury in the course of commercial activities and in households have the potential to expose workers and the general public to hazardous mercury vapours. In addition, the spills are both costly to clean up and disruptive. Clean-up procedures for small mercury spills are found in the USEPA’s website (USEPA, 2007c).
4. Critical to determining what type of response is appropriate for any mercury spill is evaluating the size of the spill and the extent of mercury dispersal and whether the necessary clean-up resources and expertise are available. Professional help should be sought in the following cases:
   1. The amount of mercury could be more than 2 tablespoons (30 millilitres): larger spills should be reported to the authorities for oversight and follow-up;
   2. The spill area is undetermined: if the spill was not witnessed or the extent of the spill is hard to determine, there could be small amounts of mercury that are hard to detect and must be cleaned up;
   3. The spill area contains surfaces that are porous or semi-porous: surfaces such as carpet and acoustic tiles can absorb the spilled mercury and make clean-up practically impossible;
   4. The spill occurs near a drain, fan, ventilation system or other conduit: mercury and mercury vapours can quickly move away from the spill site and contaminate other areas without being easily detected.
5. Scattering of spilled mercury (e.g., with the use of water jets) should be avoided as much as possible because it significantly increases the mercury evaporation rate (Euro Chlor, 2004).

L. Awareness and participation

1. Prior to a facility in which mercury wastes are disposed of beginning operations, there should be established procedures to communicate with local authorities regarding mercury releases. Facility operators should host community awareness forums to address questions concerning facility siting, operations, and emergency response plans, as well as other concerns the public may raise.
2. Public awareness and participation play key roles in implementing ESM of mercury wastes. Public participation is a core principle of the 1999 Basel Declaration on Environmentally Sound Management and many other international agreements. It is essential that the public and all stakeholders have a chance to participate in the development of legislation, policy, programmes and other decision-making processes related to mercury.
3. Articles 6, 7, 8 and 9 of the 1998 United Nations Economic Commission for Europe (UNECE) Convention on Access to Information, Public Participation in Decision-making and Access to Justice in Environmental Matters (Aarhus Convention) require specific action pertaining to public participation in specific government activities, the development of plans, policies and programmes, and the development of legislation, and call for access to justice for the public with regard to the environment.
4. When initiating activities such as the collection and recycling of mercury waste, it is advisable that participation and cooperation from the consumers who generate mercury-containing waste be obtained. Continuous awareness-raising is key to the successful collection and recycling of mercury waste. Encouraging public involvement in the design of a collection and recycling system for mercury waste by providing the public with information about potential problems caused by the environmentally unsound management of such waste would help to increase consumer awareness about the risks of mercury and mercury waste.
5. Public awareness and sensitization campaigns for local communities and citizens are important elements in promoting public participation in the ESM of mercury wastes. In order to raise the awareness of citizens, the authorities concerned, e.g., local governments, need to initiate various awareness-raising and sensitization campaigns to prompt citizens to take an interest in protecting themselves and others against the adverse effects of mercury on human health and the environment. It is important to involve community-based organizations in such campaigns because they have a close relationship with residents and other stakeholders in their communities (Honda, 2005).
6. Programmes for public awareness and public participation should generally be developed around a waste management situation at the national, local or community level. Table 10 shows examples of programmes for public awareness and participation. Such programmes have four elements: publications, environmental education programmes, public relations activities, and risk communication, to which citizens should have easy access in public places (Honda, 2005).

**Table 10:** Programmes for public awareness and public participation

|  | **Contents** | **Expected results** |
| --- | --- | --- |
| **Publications** | * Booklets, pamphlets, brochures, magazines, posters, websites, etc., in various languages and dialects to explain mercury issues in simple terms * Guidebooks on how to dispose of mercury waste | * Knowledge sources * Explanation of how people can handle products containing mercury or mercury compounds and dispose of mercury waste |
| **Environmental Education Programmes** | * Voluntary seminars * Community gatherings * Linkages with other health workshops * Demonstrations of take-back programmes * Scientific studies * Tours to facilities, etc. * eLearning | * Raising knowledge * Sharing common issues * Opportunities to discuss environmental issues directly |
| **Activities** | * Take-back programmes * Mercury-free product campaigns * Waste minimization campaigns * Community gatherings * House-to-house visits | * Implementation of environmental activities among all partners * Environmental appeal for citizens * *One-on-one* communications |
| **Risk Communication** | * Mercury exposure in general living environments * Safe level of mercury exposure * Mercury pollution levels * PRTR * Fish consumption advisories * Rice consumption advisories * Response to mercury spills from products containing mercury or mercury compounds | * Proper understanding of safe- and risk levels of mercury exposure, in appropriate circumstances * Avoidance of overreaction |

1. As part of environmental education programmes, publications should provide basic knowledge of mercury properties, mercury toxicology, the adverse effects of mercury on human health and the environment, and mercury waste-related issues, including how to manage and avoid possible exposure to mercury from such waste. Publications should be translated into locally relevant languages and dialects to ensure the information is communicated efficiently to the target population.
2. The components of an environmental education programme on mercury wastes are as follows (Honda, 2005):
   1. Awareness-raising and sensitization regarding the environment and environmental challenges;
   2. Building of knowledge and understanding of the environment and environmental challenges;
   3. Development of attitudes of concern for the environment and a motivation to improve or maintain environmental quality;
   4. Development of skills to identify and help resolve environmental challenges;
   5. Participation in activities that lead to the resolution of environmental challenges.
3. The partners of programmes on public participation can be summarized as follows (Honda, 2005):
   1. Government officials and staff in governments who work on environmental issues;
   2. People who are interested in environmental problems and have the ability to quickly understand and disseminate information to others:
      1. Children and students;
      2. Teachers and professors;
   3. Leaders and representatives from local communities and groups and others working in the field of the environment at the local or community level:
      1. People who work in non-governmental organizations;
      2. People who work in small and medium-sized enterprises;
      3. Local producers, collectors and recyclers, disposal facility owners, operators, and workers who handle mercury waste;
   4. People who are, or reside, in proximity to waste management or mercury contaminated sites;
   5. Local organizations;
   6. City residents;
   7. Enterprises.
4. To ensure that mercury releases from collection, transportation and disposal of waste are kept to a minimum, it is important to raise the awareness of the parties concerned (e.g., transporters, recyclers, and treatment operators) about the risks of mercury. This can be achieved through awareness-raising activities, such as seminars, that can provide information about new systems and regulations and opportunities for information exchange; the preparation and distribution of leaflets; and the dissemination of information via the Internet.

Annex to the technical guidelines

Bibliography

Amin-Zaki, L. et al., 1978. “Methylmercury Poisoning in Iraqi Children: Clinical Observations over Two Years”, *British Medical Journal*, vol. 11, pp. 613-616. Available at: https://www.ncbi.nlm.nih.gov/pmc/articles/PMC1603391/.

Amuda, O.S. et al., 2010. “Wastewater Treatment Process”, in Wang, L.K., Hung, Y.T. and Shammas, N.K., eds., *Handbook of Industrial and Hazardous Wastes Treatment, Volume 2*. CRC Press, New York, USA, p. 926.

Arai, Norio et al., (ed.) 1997. *Products of Incineration and Their Control Technology* [in Japanese].

Artisanal Gold Council, 2014. *Using Retorts to Reduce Mercury Use, Emissions, and Exposures in Artisanal and Small-Scale Gold Mining: A Practical Guide*. Available at: https://wedocs.unep.org/bitstream/handle/20.500.11822/31365/Guide.pdf?sequence=1&isAllowed=y.

Asano, S. et al., 2000. “Acute Inorganic Mercury Vapour Inhalation Poisoning”, *Pathology International*, vol. 50, pp. 169-174.

ASTM International, 2016. *ASTM D6784 - 16 Standard Test Method for Elemental, Oxidized, Particle-Bound and Total Mercury in Flue Gas Generated from Coal-Fired Stationary Sources* (Ontario Hydro Method). Available at: https://www.astm.org/Standards/D6784.htm.

Bakir, F. et al., 1973. “Methylmercury Poisoning in Iraq”, *Science*, vol. 181, pp. 230-241.

Bansal, R.C. and Goyal, M., 2005. *Activated Carbon Adsorption of Mercury. In: Activated Carbon Adsorption*, CRC Press, New York, pp. 326-334.

BiPRO, 2010. *Requirements for Facilities and Acceptance Criteria for the Disposal of Metallic Mercury*. Available at: http://ec.europa.eu/environment/chemicals/mercury/pdf/bipro\_study20100416.pdf.

Bull, S., 2006. Inorganic Mercury/Elemental Mercury. Available at: http://www.hpa.org.uk/chemicals/compendium/Mercury/PDF/mercury\_general\_information.pdf.

Butler, M., 1997. “Lessons from Thor Chemicals: the Links between Health, Safety and Environmental Protection”, in *The Bottom Line: Industry and the Environment in South Africa,* L. Bethlehem, Goldblatt, M. Cape Town, South Africa, University of Cape Town Press, pp. 194-213.

Canadian Centre for Occupational Health and Safety, undated. *OHS Fact Sheets: Mercury*. Available at: http://www.ccohs.ca/oshanswers/chemicals/chem\_profiles/mercury.html.

CEN, 2001. *EN 13211: Air quality - Stationary source emissions - Manual method of determination of the concentration of total mercury*. Available at: https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:13552,6245&cs=17016185F3097F69DF15410E3BF888194.

CEN, 2002a. *EN 12457-1 to 4: Characterization of waste - Leaching - Compliance test for leaching of granular waste materials and sludges*. Available at: https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:14486,2046877&cs=13604EB5109EFF1F1C575A037AD9EED7A, https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:14487,2046877&cs=16C9030D02A7B4C4C2F5C397836FC4E63, https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:14488,2046877&cs=1BACC3204C9E378F637D117CAD6640072, https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:14489,2046877&cs=15ABE8C36B9C1F3F6AC6FACF0CD6B07F8.

CEN, 2002c. *EN 13657: Characterization of waste - Digestion for subsequent determination of aqua regia soluble portion of elements in waste*. Available at: https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:14478,2046877&cs=1B6C2FCEBC23091B4121AB9598EB936F4.

CEN, 2005. *EN 14884: Air quality - Stationary source emissions - Determination of total mercury: Automated measuring systems.* Available at: https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:22225,6245&cs=1D0594D4952ACBFFCE7DD69CB326A1A4D.

CEN, 2006. *EN 12920: Characterization of waste - Methodology for the determination of the leaching behaviour of waste under specified conditions*. Available at: https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:30332,2046877&cs=170A07AF9B89260D08F2D2786DACC460C.

CEN, 2007. *EN 15309: Characterization of waste and soil - Determination of elemental composition by X-ray fluorescence.* Available at: <https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP_PROJECT,FSP_ORG_ID:21521,2046877&cs=17A780A8D8A3C9CBF29743B8970943E0F>.

CEN, 2017. *BS EN 14405: Characterization of waste - Leaching behaviour test - Up-flow percolation test*. Available at: <https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP_PROJECT,FSP_ORG_ID:39353,2046877&cs=1059433662EBF10E5DBF062E454F6B4EA>.

CEN, 2020. *EN 13656: Soil, treated biowaste, sludge and waste - Digestion with a hydrochloric (HCl), nitric (HNO3) and tetrafluoroboric (HBF4) or hydrofluoric (HF) acid mixture for subsequent determination of elements*. Available at: https://standards.cen.eu/dyn/www/f?p=204:110:0::::FSP\_PROJECT,FSP\_ORG\_ID:68059,2046877&cs=1D505D01FAE23D03768F8E5947DDC370B.

Chang, T. C. and J. H., Yen, 2006. “On-site mercury-contaminated soils remediation by using thermal desorption technology”, *Journal of Hazardous Materials*, vol. 128(2-3), 208-217.

Chiarle, S. and Ratto, M., 2000. “Mercury Removal from Water by Ion Exchange Resins Adsorption”, *Water Research*, vol. 34, pp. 2971-2978.

Chojnacki, A. et al., 2004. “The application of natural zeolites for mercury removal: from laboratory tests to industrial scale”, *Minerals Engineering*, vol. 17, pp. 933-937.

Committee on consideration of environmentally sound management of mercury waste, working group on mercury recovery and disposal, 2014. *Report on consideration of environmentally sound management of mercury wastes* [in Japanese], p. 67, Reference document No. 3-1. Available at: http://www.env.go.jp/council/03recycle/y039-01b/ref3.pdf.

Damluji, S. F. and Tikriti, S., 1972. “Mercury Poisoning from Wheat”, *British Medical Journal*, vol. 25, p. 804.

Environment Canada, 2016. Cleaning Up Small Mercury Spills. Available at: https://ec.gc.ca/mercure-mercury/default.asp?lang=En&n=D2B2AD47-1.

Environmental Management Bureau, Republic of the Philippines, 1997. *DENR Administrative Order No. 38, Chemical Control Order for Mercury and Mercury Compounds*. Available at: <http://www.environment.gov.au/epbc/notices/assessments/2010/5477/dao-97-38.pdf>.

Euro Chlor, 2004. *Code of Practice, Mercury Housekeeping, Environmental Protection 11, 5th edition*. Available at: https://wedocs.unep.org/bitstream/handle/20.500.11822/13795/ENV\_Prot\_11\_Edition\_5.pdf?sequence=1&isAllowed=y.

European Commission, 2008. *Options for reducing mercury use in products and applications and the fate of mercury already circulating in society*. Available at: http://ec.europa.eu/environment/chemicals/mercury/pdf/EU\_Mercury\_Study2008.pdf.

European Commission, 2013a. *Commission implementing decision of 9 December 2013 establishing the best available techniques (BAT) conclusions, under Directive 2010/75/EU of the European Parliament and of the Council on industrial emissions, for the production of chlor-alkali (2013/732/EU)*. Available at: <https://eur-lex.europa.eu/eli/dec_impl/2013/732/oj>. Translations to Arabic, Chinese and Russian are available at <https://eippcb.jrc.ec.europa.eu/translation/index.html>.

European Commission, 2013b. *Commission Implementing Decision of 26 March 2013 establishing the best available techniques (BAT) conclusions under Directive 2010/75/EU of the European Parliament and of the Council on industrial emissions for the production of cement, lime and magnesium oxide.* Available at: [http://data.europa.eu/eli/dec\_impl/2013/163/oj](http://data.europa.eu/eli/dec_impl/2013/163/oj" \o "Gives access to this document through its ELI URI.).

European Commission, 2014a. *Report on the availability of mercury-free button cells for hearing aids, in accordance with Article 4.4 of Directive 2006/66/EC of the European Parliament and of the Council on batteries and accumulators and waste batteries and accumulators and repealing Directive 91/157/EEC*. Available at: <https://ec.europa.eu/transparency/regdoc/rep/1/2014/EN/1-2014-632-EN-F1-1.Pdf>.

European Commission, 2014b. *Integrated Pollution Prevention and Control (IPPC) - Best Available Techniques (BAT) Reference Document for the Production of Chlor-alkali*. Available at: <https://eippcb.jrc.ec.europa.eu/sites/default/files/2019-11/CAK_BREF_102014.pdf>.

European Commission, 2017a. *Best Available Techniques (BAT) Reference Document for Large Combustion Plants, Integrated Pollution Prevention and Control (IPPC)*. Available at: <https://eippcb.jrc.ec.europa.eu/sites/default/files/2019-11/JRC_107769_LCPBref_2017.pdf>.

European Commission, 2017b. *Commission Implementing Decision (EU) 2017/1442 of 31 July 2017 establishing the best available techniques (BAT) conclusions, under Directive 2010/75/EU of the European Parliament and of the Council, for large combustion plants*. Available at: [http://data.europa.eu/eli/dec\_impl/2017/1442/oj](http://data.europa.eu/eli/dec_impl/2017/1442/oj" \o "Gives access to this document through its ELI URI.).

European Commission, 2018a. *Commission Implementing Decision (EU) 2018/1147 of 10 August 2018 establishing best available techniques (BAT) conclusions for waste treatment, under Directive 2010/75/EU of the European Parliament and of the Council*. Available at: [http://data.europa.eu/eli/dec\_impl/2018/1147/oj](http://data.europa.eu/eli/dec_impl/2018/1147/oj" \o "Gives access to this document through its ELI URI.).

European Commission, 2018b. *Best Available Techniques (BAT) Reference Document Waste Treatment*. Available at: <https://eippcb.jrc.ec.europa.eu/sites/default/files/2019-11/JRC113018_WT_Bref.pdf>.

European Commission, 2019a. *Reference Document on the Best Available Techniques for Waste Incineration*. Available at: <http://eippcb.jrc.ec.europa.eu/reference/waste-incineration>.

European Commission, 2019b. *Commission Implementing Decision (EU) 2019/2010 of 12 November 2019 establishing the best available techniques (BAT) conclusions, under Directive 2010/75/EU of the European Parliament and of the Council, for waste incineration*. Available at: [http://data.europa.eu/eli/dec\_impl/2019/2010/oj](https://urldefense.com/v3/__https:/smex-ctp.trendmicro.com:443/wis/clicktime/v1/query?url=http*3a*2f*2fdata.europa.eu*2feli*2fdec*5fimpl*2f2019*2f2010*2foj&umid=70e28e1b-ed67-4b92-b784-7c00e9edfe9a&auth=af930d9d75d0cf4b98837d729862ac50a7415b97-b2d0cc3421fc6873582c3fd317c07643cd2e8d50__;JSUlJSUlJSUl!!DOxrgLBm!VT8gVlFuALwnjFqObmxOBbpE0HHZCRvAeV_Wreu5dYgbwzy7jR3Lr0PyJ_RWue_vKvzGjAzT_e6IFqBPjnDytA$" \o "Gives access to this document through its ELI URI.)

European Union, 2003. "Safety Assessment for Acceptance of Waste in Underground Storage", *Appendix A to Council Decision 2003/33/EC of 19 December 2002 establishing criteria and procedures for the acceptance of waste at landfills pursuant to Article 16 of and Annex II to Directive 1999/31/EC*.

European Union, 2006. *Directive 2006/66/EC of the European Parliament and of the Council of 6 September 2006 on batteries and accumulators and waste batteries and accumulators and repealing Directive 91/157/EEC.* Available at: <http://data.europa.eu/eli/dir/2006/66/2018-07-04>.

European Union, 2010. Former *Regulation (EC) No. 1102/2008 of the European Parliament and of the Council of 22 October 2008 on the banning of exports of metallic mercury and certain mercury compounds and mixtures and the safe storage of metallic mercury*. Available at <https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=celex%3A32008R1102>.

European Union, 2011a. *Directive 2011/65/EU of the European Parliament and of the Council of 8 June 2011 on the restriction of the use of certain hazardous substances in electrical and electronic equipment*. Official Journal of the European Union, L 174, 1.7.2011, p. 88. Available at: [http://data.europa.eu/eli/dir/2011/65/2020-05-01](http://data.europa.eu/eli/dir/2011/65/2020-05-01" \o "Gives access to this document through its ELI URI.)

European Union, 2011b. *Council Directive 2011/97/EU of 5 December 2011 amending Directive 1999/31/EC as regards specific criteria for the storage of metallic mercury considered as waste*. Official Journal of the European Union, L 328, 10/12/2011, p. 49. Available at:

http://eurlex.europa.eu/LexUriServ/LexUriServ.do?uri=OJ:L:2011:328:0049:0052:EN:PDF.

European Union, 2012. *Directive 2012/19/EU of the European Parliament and of the Council of 4 July 2012 on Waste Electrical and Electronic Equipment* (WEEE)   [Official Journal of the European Union, L 197, 24.7.2012, p. 38. Available at: http://data.europa.eu/eli/dir/2012/19/2018-07-04](https://exri.sharepoint.com/sites/msteams_d954f0/Shared%20Documents/Q202002水銀廃棄物ESM/Official%20Journal%20of%20the%20European%20Union,%20L%20197,%2024.7.2012,%20p.%2038.%20Available%20at:%20http:/data.europa.eu/eli/dir/2012/19/2018-07-04).

European Union, 2017. *Regulation (EU) 2017/852 of the European Parliament and of the Council of 17 May 2017 on mercury, and repealing Regulation (EC) No 1102/2008*. Available at:

<https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX:32017R0852>

FAO, 1985. *Guidelines for the Packaging and Storage of Pesticides*. Available at: http://www.bvsde.paho.org/bvstox/i/fulltext/fao12/fao12.pdf.

Gay, D.D., Cox, R.D. and Reinhardt, J.W., 1979. “Chewing Releases Mercury from Fillings”, *Lancet,* vol. 1, pp. 985-986.

German Federal Environment Agency, 2014. *Behaviour of mercury and mercury compounds at the underground disposal in salt formations and their potential mobilisation by saline solutions*. Available at: https://www.umweltbundesamt.de/sites/default/files/medien/378/publikationen/texte\_07\_2014\_behaviour\_of\_mercury\_and\_mercury\_compounds\_at\_the\_underground\_disposal\_in\_salt\_formations-summary.pdf.

Glenz, T. G., Brosseau, L.M. and Hoffbeck, R.W., 2009. “Preventing Mercury Vapor Release from Broken Fluorescent Lamps during Shipping”, *Journal of the Air and Waste Management Association*, vol. 59, pp. 266-272.

Government of Canada, 2014. *Products Containing Mercury Regulations* (in Norwegian). Available at: https://laws-lois.justice.gc.ca/eng/regulations/SOR-2014-254/index.html.

Government of Norway, 2011. *Product Regulations.* Available at: https://lovdata.no/dokument/SF/forskrift/2004-06-01-922/KAPITTEL\_2#KAPITTEL\_2.

GroundWork, 2005. *Advising and Monitoring the Clean-up and Disposal of Mercury Waste in Kwazulu-Natal, South Africa*. Available at: http://www.zeromercury.org/phocadownload/Whats\_on\_in\_the\_regions/groundWork\_Phase\_one\_Final\_Report\_1006\_WebVs.pdf.

Grundfelt, B. et al., 2005. “Importance of the multi-barrier concept for the final disposal of radioactive waste” [in German], Kemakta Konsult AB, Bericht, Stockholm. Available at: http://www.bfs.de/de/endlager/publika/AG\_3\_Konzeptgrund\_Mehrbarrierenkonzept1.pdf.

Hagemann, S., 2009. “Technologies for the stabilization of elemental mercury and mercury-containing wastes”, Gesellschaft für Anlagen-und Reaktorsicherheit (GRS). *GRS Report 252*.

Honda S., 2005. “Study on the Environmentally Sound Management of Hazardous Wastes and Other Wastes in the Asia”, postdoctoral dissertation, Tsinghua University, Beijing, China.

Honda, S. et al., 2006. “Current Mercury Level in Cambodia - with Issue on Waste Management”, *NIMD Forum 2006 II: Current Issues on Mercury Pollution in the Asia-Pacific Region*, Minamata City, Japan, pp. 91-102. Available at: http://nimd.env.go.jp/english/kenkyu/nimd\_forum/nimd\_forum\_2006\_II.pdf.

Hylander, L.D. and Meili, M., 2005. “The Rise and Fall of Mercury: Converting a Resource to Refuse after 500 Years of Mining and Pollution”, *Critical Reviews in Environmental Science and Technology*, vol. 35, pp. 1-36.

IAEA, 2009. *Geological Disposal of Radioactive Waste: Technological Implications for Retrievability*. Available at: http://www-pub.iaea.org/MTCD/publications/PDF/Pub1378\_web.pdf.

IATA, 2020. *Dangerous Goods Regulations Manual* (61st edition). Available at: https://www.labeline.com/product/iata-dgr-61st/.

ICAO, 2017. *Technical Instructions for the Safe Transport of Dangerous Goods by Air* (2017-2018 edition). Available at:

https://www.icao.int/safety/dangerousgoods/pages/technical-instructions.aspx.

ILO, 2014. *Mercuric Oxide.* Available at: http://www.ilo.org/dyn/icsc/showcard.display?p\_card\_id=0981&p\_version=2&p\_lang=en.

ILO, 2019. *Mercurous Chloride*. Available at: http://www.ilo.org/dyn/icsc/showcard.display?p\_lang=en&p\_card\_id=0984.

IMO, 2018. *International Maritime Dangerous Goods Code* (2018 edition). Available at: https://www.imo.org/en/publications/Pages/IMDG%20Code.aspx.

Interstate Technology and Regulatory Cooperation Work Group (ITRC), 1998. *Technical Guidelines for On-site Thermal Desorption of Solid Media and Low Level Mixed Waste Contaminated with Mercury and/or Hazardous Chlorinated Organics.* Available at: http://www.itrcweb.org/GuidanceDocuments/td-3.pdf.

ISO, 2006. *ISO 17852: 2006: Water quality – Determination of mercury - Method using atomic fluorescence spectrometry*. Available at: https://www.iso.org/standard/38502.html.

ISO, 2012. *ISO 12846:2012: Water quality – Determination of mercury - Method using atomic absorption spectrometry (AAS) with and without enrichment*. Available at: https://www.iso.org/standard/51964.html.

ISO, 2018a. *ISO 18400-104:2018 Soil quality - Sampling - Part 104: Strategies*. Available at: https://www.iso.org/standard/65223.html.

ISO, 2018b. *ISO 18400-202:2018 Soil quality - Sampling - Part 202: Preliminary investigations*. Available at: <https://www.iso.org/standard/65225.html>.

ISO, 2018c. *ISO 18400-203:2018 Soil quality - Sampling - Part 203: Investigation of potentially contaminated sites*. Available at: https://www.iso.org/standard/65226.html.

ISO, 2020. ISO 21741:2020 Stationary source emissions — Sampling and determination of mercury compounds in flue gas using gold amalgamation trap. Available at: https://www.iso.org/standard/71532.html.

Jang, M., Hong, S. M. and Park, J. K., 2005. “Characterization and Recovery of Mercury from Spent Fluorescent Lamps”, *Waste Management*, vol. 25, pp. 5-14.

Japan Public Health Association, 2001. *Preventive Measures against Environmental Mercury Pollution and Its Health Effects*, Japan Public Health Association, Tokyo, Japan. Available at <http://nimd.env.go.jp/english/kenkyu/docs/manual.pdf>.Jew, AD et al., 2014. “Microbially enhanced dissolution of HgS in an acid mine drainage system in the California Coast Range”, *Geobiology*, vol. 12 No. 1, pp. 20-33.

JSA, 1997. *JIS K 0222:1997 Methods for determination of mercury in stack gas*. Available at: https://webdesk.jsa.or.jp/books/W11M0090/index/?bunsyo\_id=JIS+K+0222%3A1997.

Kanai, Y. and Endou, H. 2003. “Functional Properties of Multispecific Amino Acid Transporters and Their Implications to Transpoter-Mediated Toxicity”, *Journal of Toxicological Sciences*, vol. 28, pp. 1-17. Available at: https://www.jstage.jst.go.jp/article/jts/28/1/28\_1\_1/\_pdf.

Kerper, L.E., Ballatori, N. and Clarkson, T.W., 1992. “Methylmercury Transport Across the Blood-Brain Barrier by an Amino Acid Carrier”, *American Journal of Physiology - Regulatory, Integrative and Comparative Physiology*, vol. 262, pp. 761-765.

Kobelco Eco-Solutions Co. Ltd., 2001. “Recycling System for Fluorescent Lamps” [in Japanese], p.45.

Kuncova, H., Petrlik, J. and Stavkova, M., 2007. “Chlorine Production – a Large Source of Mercury Releases (The Czech Republic Case Study)”, prepared by Arnika Association, Prague. Available at: http://english.arnika.org/files/documents/Mercury\_CZ.pdf.

Lambrecht, B., 1989. “Zulus Get Exported Poison - US Mercury Waste Pollutes Drinking Water in S. Africa”, *St Louis Post-Dispatch*, p.26.

Lamp Recycling Outreach Project, undated. “Training Module (1-hour version) for Generators and Handlers Of Fluorescent and Mercury-Containing Lamps (and Ballasts)”. Available at: <http://www.almr.org/1hourtrainingmodule.pdf>.

Lindberg, S.E. and Price, J. L., 1999. “Airborne Emissions of Mercury from Municipal Landfill Operations: A Short-Term Measurement Study in Florida”, *Journal of the Air & Waste Management Association*, vol., 49, pp. 520-532.

Lindberg, S. E. et al., 2001. “Methylated mercury species in municipal waste landfill gas sampled in Florida, USA”, *Atmospheric Environment*, vol. 35 No. 23, pp. 4011-4015.

López, F.A. et al., 2010. “Formation of metacinnabar by milling of liquid mercury and elemental sulfur for long term mercury storage”, *Science of the Total Environment*, vol. 408 No. 20, pp. 4341-4345.

López, F.A. et al., 2015. “Mercury leaching from hazardous industrial wastes stabilized by sulfur polymer encapsulation”, *Waste Management*, vol. 35, pp. 301-306.

López-Delgado, A. et al., 2012. “A microencapsulation process of liquid mercury by sulfur polymer stabilization/solidification technology. Part I: Characterization of materials”, *Revista de Metalurgia*, vol. 48 No. 1, pp. 45-57.

Lowell Center for Sustainable Production, 2003. *An Investigation of Alternatives to Mercury Containing Products*. Available at: https://www.unenvironment.org/resources/report/investigation-alternatives-mercury-containing-products-0.

Maine Department of Environmental Protection, 2008. *Maine Compact Fluorescent Lamp Study*. Available at: http://www.maine.gov/dep/rwm/homeowner/cflreport.htm.

Mattus, C. H., 1999. *Measurements of mercury released from amalgams and sulfide compounds*. Available at: https://www.osti.gov/servlets/purl/5899.

Maxson, P., 2011. Personal communication.

MEE China, 1995. *GB/T 15555.1-1995. Solid Waste - Determination of Total Mercury - Cold Atomic Absorption Spectrometry*. Available at: http://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/jcffbz/199601/t19960101\_82016.shtml.

MEE China, 2007a. *HJ/T 299-2007. Solid waste - Extraction Procedure for Leaching Toxicity - Sulphuric Acid & Nitric Acid Method*. Available at: http://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/jcffbz/200704/t20070418\_102859.shtml.

MEE China, 2007b. *HJ/T 300-2007. Solid waste - Extraction Procedure for Leaching Toxicity - Acetic Acid Buffer Solution* Method. Available at: http://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/jcffbz/200704/t20070418\_102860.shtml.

MEE China, 2009. *HJ 543-2009. Stationary Source Emission - Determination of Mercury - Cold Atomic Absorption Spectrophotometry* (Provisional). Available at: http://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/jcffbz/201001/t20100112\_184167.shtml.

MEE China, 2010. *HJ 557-2010. Solid waste - Extraction Procedure for Leaching Toxicity - Horizontal Vibration Method*. Available at: http://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/jcffbz/201002/t20100209\_185623.shtml.

MEE China, 2014. *HJ 702-2014. Solid Waste - Determination of Mercury, Arsenic, Selenium, Bismuth, Antimony - Microwave Dissolution / Atomic Fluorescence Spectrometry*. Available at: http://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/jcffbz/201409/t20140912\_288936.shtml.

MEE China, 2017. *HJ 917-2017. Stationary Source Emission - Determination of Total Gaseous Mercury - Carbon Sorbent Traps / Thermal Cracking Atomic Absorption Spectrophotometric Method*. Available at: http://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/jcffbz/201801/t20180108\_429285.shtml.

MEE China, 2019. *GB 18598 Standard for pollution control on the security landfill site for hazardous wastes*. Available at: https://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/gthw/wxfwjbffbz/201910/t20191012\_737241.shtml.

MEE China, 2020. *GB 18599 Standard for pollution control on the non-hazardous industrial solid waste storage and landfill*. Available at: http://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/gthw/gtfwwrkzbz/202012/t20201218\_813927.shtml.

Minamata Convention, 2011. *Guidance on Best Available Techniques and Best Environmental Practices* (Chapter VI). Available at: <http://mercuryconvention.org/Portals/11/documents/publications/BAT_BEP_E_interractif.pdf>.

Ministry of the Environment of Japan, 1997. *Our Intensive Efforts to Overcome the Tragic History of Minamata Disease*.

Ministry of the Environment of Japan, 2002. *Minamata Disease - The History and Measures*. Available at: http://www.env.go.jp/en/chemi/hs/minamata2002/index.html.

Ministry of the Environment of Japan, 2007a. *Guidebook for Waste Management - Case Study of Promoting 3Rs in Japan*, JICA Seminar on Waste Management in Japan, Yokohama International Center.

Ministry of the Environment of Japan, 2007b. *Waste Disposal and Recycling Measures*. Available at: http://www.env.go.jp/en/recycle/manage/waste.html.

Ministry of the Environment of Japan, 2010. *Lessons from Minamata Disease and Mercury Management in Japan*. Available at: http://www.env.go.jp/chemi/tmms/pr-m/mat01/en\_full.pdf.

Ministry of the Environment of Japan, 2015. *Japan’s policy on the environmentally sound management of mercury wastes (summary)* (recommended by the Central Environment Council in February 2015). Available at: http://www.env.go.jp/en/recycle/wm/150413jpmw.pdf.

Ministry of Environmental Protection of China, 2010. *Project Report on the Reduction of Mercury Use and Emission in Carbide PVC Production*. Available at: http://www.unep.org/chemicalsandwaste/Portals/9/Mercury/VCM%20Production/Phase%20I%20Final%20Report%20-%20PVC%20Project%20Report%20for%20China.pdf.

Mizutani, S., Kadotani, K. and Kanjo, Y., 2010. “Adsorption behavior of mercuric compounds on soils under different pH condition” [in Japanese], *Environmental Engineering Research*, Vol. 47, pp. 267-272.

Mottet, N.K., Shaw, C.M. and Burbacher, T.M., 1985. “Health Risks from Increases in Methylmercury Exposure”, *Environmental Health Perspectives*, vol. 63, pp. 133-140. Available at: http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=1568483.

National Institute for Minamata Disease (NIMD), 1999. “Investigation into Suspected Mercury Contamination at Sihanoukville, Cambodia”, *Proceedings of NIMD Forum ’99*. Available at: http://www.nimd.go.jp/english/kenkyu/nimd\_forum/nimd\_forum\_1999.pdf.

Nomura Kohsan Co. Ltd., 2007. Treatment of Mercury-containing Wastes at Itomuka Plant of Nomurakohsan Co., Ltd. Tokyo, Japan.

OECD, 2001a. *Extended Producer Responsibility: A Guidance Manual for Governments.*

OECD, 2001b. *Harmonised Integrated Classification System for Human Health and Environmental. Hazards of Chemical Substances and Mixtures*. Available at: http://www.oecd.org/officialdocuments/publicdisplaydocumentpdf/?doclanguage=en&cote=env/jm/mono(2001)6.

OECD, 2004. *Recommendation of the Council on the Environmentally Sound Management of Waste*. Available at: http://acts.oecd.org/Instruments/ShowInstrumentView.aspx?InstrumentID=51.

OECD, 2007. *Guidance Manual on Environmentally Sound Management of Waste*. Available at: <http://www.oecd.org/dataoecd/23/31/39559085.pdf>.

Office of Technology Assessment, 1983. “Case Examples of Process Modification - Appendix 5A”, in *Technologies and Management Strategies for Hazardous Waste Control*, The Office of Technology Assessment, Darby, USA, Diane Publishing, pp. 213-217.

Ogaki, Y., Yamada, Y. and Nomura, M., 2004. “Recycling Technology of JFE Group for Recycle Oriented Society” [in Japanese], *JFE GIHO*, vol. 6, pp. 37-43. Available at: http://www.jfe-steel.co.jp/research/giho/006/pdf/006-07.pdf.

Oikawa, K. et al., 1983. “Respiratory Tract Retention of Inhaled Air Pollutants, Report 1: Mercury Absorption by Inhaling Through the Nose and Expiring Through the Mouth at Various Concentrations”, *Chemosphere*, vol. 11, 943-951.

Oliveira, R.B. et al., 1998. “Methylmercury Intoxication and Histochemical Demonstration of NADPH-Diaphorase Activity in the Striate Cortex of Adult Cats”, *Brazilian Journal of Medical and Biological Research*, vol. 31, pp. 1157-1161.

Ozonoff, D.M., 2006. “Methylmercury”. Available at: http://www.ijc.org/rel/pdf/health\_effects\_spring2006.pdf.

Parker, J. L. and Bloom, N.S., 2005. “Preservation and storage techniques for low-level mercury speciation”, *Science of the Total Environment,* vol. 337, pp. 253-263.

QSC, 2003. Mercury Stewardship Best Management Practices, October (available on request).

Richardson, G.M. and Allan, M., 1996. “A Monte Carlo Assessment of Mercury Exposure and Risks from Dental Amalgam”, *Human and Ecological Risk Assessment*, vol. 2, pp. 709-761.

Richardson, G.M., 2003. “Inhalation of Mercury-Contaminated Particulate Matter by Dentists: An Overlooked Occupational Risk”, *Human and Ecological Risk Assessment*, vol. 9, pp. 1519-1531.

Sakamoto, M. et al., 2005. “Difference in Methylmercury Exposure to Fetus and Breast-Feeding Offspring”, *Korean Journal of Environmental Health*, vol. 31, pp. 179-186.

Sakamoto, M. et al., 2006. “Maternal and Fetal Mercury and n-3 Polyunsaturated Fatty Acid as a Risk and Benefit of Fish Consumption to Fetus”, *Environmental Health and Preventive Medicine*, 11(4):171-6. Available at: https://www.ncbi.nlm.nih.gov/pubmed/1529819

Sanborn, J.R. and Brodberg, R.K., 2006. “Evaluation of Bioaccumulation Factors and Translators for Methylmercury”. Available at: https://semspub.epa.gov/work/01/466770.pdf.

School of Natural Resources and Environment, University of Michigan, 2000. “Environmental Justice Case Study - Thor Chemicals and Mercury Exposure in Cato-Ridge, Kwazulu-Natal, South Africa”. Available at: http://www.umich.edu/~snre492/Jones/thorchem.htm.

Science Applications International Corporation, 2002. “Technical Background Document: Mercury Wastes Evaluation of Treatment of Bulk Elemental Mercury Final Report”. Available at: http://www.regulations.gov/#!documentDetail;D=EPA-HQ-RCRA-2002-0029-0005.

Spiegel, S. and Veiga, M., 2006. “Interventions to Reduce Mercury Pollution in Artisanal Gold Mining Sites - lessons from the UNDP/GEF/UNIDO Global Mercury Project”, NIMD Forum 2006 II, Minamata City, Ministry of the Environment, Japan, pp. 1-18. Available at: http://nimd.env.go.jp/english/kenkyu/nimd\_forum/nimd\_forum\_2006\_II.pdf.

Steffen, A. et al., C. 2007. “A Synthesis of Atmospheric Mercury Depletion Event Chemistry Linking Atmosphere, Snow and Water”, *Atmospheric Chemistry and Physics Discussions*, vol. 7, pp. 10837-10931.

Tajima, S., 1970. “Studies on the Formation of Methylmercury Compounds. 1. Preparation of Monomercurated Acetaldehyde XHgCH2CHO and Formation of Methylmercury Compounds from Monomercurated Acetaldehyde” [in Japanese], *Kumamoto Igakkai Zasshi*, vol. 44, pp. 873-886.

Takahashi, Nakamura, Mizoiri, and Shoji, 2004. “Mercury Behaviour in Chuo Bohatei Sotogawa Landfill” [in Japanese], Annual Report of the Tokyo Metropolitan Research Institute for Environmental Protection 2004, pp. 165-171.

Tanel, B., Reyes-Osorno, B. and Tansel, I.N., 1998. “Comparative Analysis of Fluorescent Lamp Recycling and Disposal Options”, *Journal of Solid Waste Technology and Management*, vol. 25, pp. 82-88.

United Nations, 2019a. *United Nations Recommendations on the Transport of Dangerous Goods - Model Regulations* (21st revised edition). Available at: https://www.unece.org/trans/danger/publi/unrec/rev21/21files\_e.html.

United Nations, 2019b. *Globally Harmonized System of Classification and Labelling of Chemicals*. Available at:

https://www.unece.org/trans/danger/publi/ghs/ghs\_welcome\_e.html.

UNDP, 2010. *Guidance on the Cleanup, Temporary or Intermediate Storage, and Transport of Mercury Waste from Health Care Facilities*. Available at: <http://www.gefmedwaste.org/downloads/Guidance%20on%20Cleanup%20Storage%20and%20Transport%20of%20Mercury%20from%20Health%20Care%20July%202010.pdf>.

UNECE, 1998.Available at: <http://www.unece.org/env/treaties/welcome.html>.

Available at: http://www.unep.org/chemicalsandwaste/Mercury/PrioritiesforAction/ArtisanalandSmallScaleGoldMining/Reports/tabid/4489/language/en-US/Default.aspx.

UNEP, 2002. *Global Mercury Assessment*, Geneva, Switzerland. Available at: http://www.eurocbc.org/final-assessment-report-25nov02.pdf.

UNEP, 2005. *Toolkit for Identification and Quantification of Mercury Releases*. Available at: http://web.unep.org/globalmercurypartnership/toolkit-identification-and-quantification-mercury-releases.

UNEP, 2006a. Strategic Approach to International Chemicals Management (SAICM). Available at: http://www.saicm.org/images/saicm\_documents/saicm%20texts/SAICM\_publication\_ENG.pdf.

UNEP, 2006b. *Guide for Reducing Major Uses and Releases of Mercury*. Available at: https://wedocs.unep.org/bitstream/handle/20.500.11822/11588/Hg\_Sector\_Guide\_July\_2006.pdf?sequence=1&%3BisAllowed=.

UNEP, 2006c. *Summary of Supply, Trade and Demand Information on Mercury*Geneva, Switzerland. Available at: https://wedocs.unep.org/bitstream/handle/20.500.11822/11610/HgSupplyTradeDemandJM.pdf?sequence=1&isAllowed=y.

UNEP, 2008a. *Global Atmospheric Mercury Assessment: Sources, Emissions and Transport*. Available at:

https://wedocs.unep.org/bitstream/handle/20.500.11822/13769/UNEP\_GlobalAtmosphericMercuryAssessment\_May2009.pdf?sequence=1&%3BisAllowed=.

UNEP, 2008b. *Report on the Major Mercury Containing Products and Processes, Their Substitutes and Experience in Switching to Mercury Free Products and Processes*. Available at:

http://www.mercuryconvention.org/Portals/11/documents/meetings/oewg2/English/2\_7.pdf.

UNEP, 2008c. *Summary Report on UNEP Mercury Inventory Activities*. Available at: http://www.mercuryconvention.org/Portals/11/documents/meetings/oewg2/English/2\_INF14.pdf.

UNEP, 2008d. *[Mercury] awareness raising package*. Available at: http://www.unep.org/chemicalsandwaste/Mercury/MercuryPublications/ReportsPublications/AwarenessRaisingPackage/tabid/4022/language/en-US/Default.aspx.

Available at: <http://www.mercuryconvention.org/Portals/11/documents/meetings/inc2/English/INC2_4_para29.pdf>.

UNEP, 2013. *Guideline on environmentally sound material ecovery and recycling of end-of-life computing equipment*. Available at: http://www.basel.int/Implementation/TechnicalAssistance/Partnerships/PACE/PACEGuidelines,ManualandReports/tabid/3247/Default.aspx.

UNEP, 2014. *Report on the status of projects funded under the Quick Start Programme as of February 2014*. Available at: http://www.saicm.org/images/SAICM.EB.9.4.rev1%20Report%20on%20projects%20funded%20under%20the%20QSP.pdf.

UNEP, 2015a. *Methodological guide for the development of inventories of hazardous wastes and other wastes under the Basel Convention*. Available at: http://www.basel.int/Implementation/Publications/GuidanceManuals/tabid/2364/Default.aspx#.

UNEP, 2015b. *Manual for the Implementation of the Basel Convention*. Available at: www.basel.int.

UNEP, 2015c. *Guide to the Control System*. Available at: http://www.basel.int.

UNEP, 2016. *Lessons from Countries Phasing Down Dental Amalgam Use*. Available at: https://web.unep.org/globalmercurypartnership/lessons-countries-phasing-down-dental-amalgam-use.

UNEP, 2017. *Global mercury supply, trade and demand*. Available at: https://wedocs.unep.org/bitstream/handle/20.500.11822/21725/global\_mercury.pdf?sequence=1&isAllowed=y.

UNEP, 2018. *Guidelines on the environmentally sound interim storage of mercury other than waste mercury*. Available at:

http://www.mercuryconvention.org/Portals/11/documents/forms-guidance/English/2\_5\_e\_Rev1\_storage.pdf.

UNEP, 2019a. *Global Mercury Assessment 2018*. Available at: https://www.unenvironment.org/resources/publication/global-mercury-assessment-2018.

UNEP, 2019b. Toolkit for Identification and Quantification of Mercury Releases. Available at: https://wedocs.unep.org/bitstream/handle/20.500.11822/30684/HgTlktRef.pdf?sequence=1&isAllowed=y.

UNEP, 2019c. *Guidance on best available techniques and best environmental practices*. Available at: http://mercuryconvention.org/Portals/11/documents/publications/BAT\_BEP\_E\_interractif.pdf.

UNEP, 2019d. *Guidance on the management of contaminated sites*. Available at: www.mercuryconvention.org/Portals/11/documents/forms-guidance/English/Guidance\_Contaminated\_Sites\_EN.pdf.

UNEP, 2022a. *General Technical Guidelines for the Environmentally Sound Management of Wastes Consisting of, Containing or Contaminated with Persistent Organic Pollutants*. Available at: <http://www.basel.int/Implementation/TechnicalMatters/TechnicalGuidelines/tabid/8025/Default.aspx>

UNEP, 2022b. *Technical guidelines on the environmentally sound disposal of hazardous wastes and other wastes in specially engineered landfill (D5).* Available at: <http://www.basel.int/Implementation/TechnicalMatters/TechnicalGuidelines/tabid/8025/Default.aspx>

UNEP, 2022c. *Technical guidelines on the environmentally sound incineration of hazardous wastes and other wastes as covered by disposal operations D10 and R1*. Available at: <http://www.basel.int/Implementation/TechnicalMatters/TechnicalGuidelines/tabid/8025/Default.aspx>

UNEP and WHO, 2008. *Identifying Populations at Risk*. Available at:

https://www.who.int/foodsafety/publications/chem/mercuryexposure.pdf?ua=1.

UNEP and SETAC, 2009. *Life Cycle Management*. Available at: http://www.unep.fr/shared/publications/pdf/DTIx1208xPA-LifeCycleApproach-Howbusinessusesit.pdf.

UNIDO, 2018. *Waste mercury perspective*. Available at: https://www.unido.org/sites/default/files/files/2019-02/UNIDO%20Mercury%20Waste%20Forecast%202025.pdf.

U.S. Department of Energy, 2009. *US Department of Energy Interim Guidance on Packaging, Transportation, Receipt, Management, and Long-Term Storage of Elemental Mercury*. Available at: https://www.federalregister.gov/documents/2009/11/16/E9-27395/notice-of-availability-of-the-us-department-of-energy-interim-guidance-on-packaging-transportation.

U.S. Department of Transportation, Transport Canada, Secretariat of Communications and Transportation of Mexico (SCT), 2012. *Emergency Response Guidebook*. Available at: https://www.tc.gc.ca/eng/canutec/guide-menu-227.htm.

USEPA, undated. Treatment Standards for Hazardous Wastes Subject to Land Disposal Restrictions. Available at: https://www.epa.gov/hw/treatment-standards-hazardous-wastes-subject-land-disposal-restrictions.

USEPA, 1994. *US EPA Method 7470 A: Mercury in Liquid Waste Manual Cold-Vapor Technique)*. Available at: https://www.epa.gov/sites/production/files/2015-12/documents/7470a.pdf.

USEPA, 1996. *US EPA Method 0060: Determination of Metals in Stack Emissions*. Available at: https://www.epa.gov/sites/production/files/2015-12/documents/0060.pdf.

USEPA, 1997a. *Locating and Estimating Air Emissions from Sources of Mercury and Mercury Compounds*. Available at: http://www.epa.gov/ttn/chief/le/mercury.pdf.

USEPA, 1997b. *Sensitive Environments and the Siting of Hazardous Waste Management Facilities*. Available at: https://archive.epa.gov/wastes/wyl/web/pdf/sites.pdf.

USEPA, 2000. *Proceedings and Summary Report - Workshop on Mercury in Products, Processes, Waste and the Environment: Eliminating, Reducing and Managing Risks from Non-Combustion Sources*. Available at: https://cfpub.epa.gov/si/si\_public\_record\_report.cfm?Lab=NRMRL&dirEntryId=64075.

USEPA, 2001. *Mercury Response Guidebook (for Emergency Responders)*. Available at: http://www.epa.gov/mercury/spills/index.htm.

USEPA, 2007a. Mercury Treatment Technologies. Available at: http://www.clu-in.org/contaminantfocus/default.focus/sec/Mercury/cat/Treatment\_Technologies.

USEPA, 2007b. *Treatment Technologies for Mercury in Soil, Waste and Water*. Available at: https://www.epa.gov/remedytech/treatment-technologies-mercury-soil-waste-and-water.

USEPA, 2007c. *Spills, disposal and site clean-up*. Available at: http://www.epa.gov/mercury/spills/index.htm.

USEPA, 2007d. *US EPA Method 7471B: Mercury in Solid or Semisolid Waste (Manual Cold-Vapor Technique)*. Available at: https://www.epa.gov/sites/production/files/2015-12/documents/7471b.pdf.

USEPA, 2007e. *US EPA Method 7473: Mercury in Solids and Solutions by Thermal Decomposition, Amalgamation, and Atomic Absorption Spectrophotometry*. Available at: <https://www.epa.gov/sites/production/files/2015-12/documents/7473.pdf>.

USEPA, 2008. Manual for the Construction of a Mercury Collection System for Use in Gold Shops. Available at: http://www.epa.gov/oia//toxics/asgm.html.

USEPA, 2017a. *Method 1313 - Liquid-Solid Partitioning as a Function of Extract pH Using a Parallel Batch Extraction Procedure*. Available at:

<https://www.epa.gov/sites/production/files/2017-10/documents/method_1313_-_final_8-3-17.pdf>.

USEPA, 2017b. *Method 1314 - Liquid-Solid Partitioning as a Function of Liquid-Solid Ratio for Constituents in Solid Materials Using an Up-Flow Percolation Column Procedure*. Available at:

https://www.epa.gov/sites/production/files/2017-10/documents/method\_1314\_-\_final\_8-3-17.pdf.

USEPA, 2017c. *Method 1315 - Mass Transfer Rates of Constituents in Monolithic or Compacted Granular Materials Using a Semi-Dynamic Tank Leaching Procedure*. Available at: https://www.epa.gov/sites/production/files/2017-10/documents/method\_1315\_-\_final\_8-3-17.pdf.

USEPA, 2017d. *Method 1316 - Liquid-Solid Partitioning as a Function of Liquid-Solid Ratio Using a Parallel Batch Extraction Procedure*. Available at: https://www.epa.gov/sites/production/files/2017-10/documents/method\_1316\_-\_final\_8-3-17.pdf.

USEPA, 2019. *Leaching Environmental Assessment Framework (LEAF) How-To Guide*. Available at: https://www.epa.gov/sites/production/files/2019-05/documents/final\_leaching\_environmental\_assessment\_framework\_leaf\_how-to\_guide.pdf.

Vosteen, Bernhard W., Rico Kanefke and Heinz Köser, 2006. “Bromine-enhanced Mercury Abatement from Combustion Flue Gases - Recent Industrial Applications and Laboratory Research.” *International Journal for Electricity and Heat Generation* 86, no. 3/2006: 70-75. Available at: https://www.vosteen-consulting.de/fileadmin/user\_upload/Publikationen/2006/2006\_03\_Vosteen\_Consulting\_Special\_print\_VGB\_PT03\_06\_U.pdf.

Waples, Jacob S. et al., 2005. “Dissolution of cinnabar (HgS) in the presence of natural organic matter”, *Geochimica et Cosmochimica Acta*, vol. 69 No. 6, pp. 1575-1588.

Waste and Resources Action Programme, 2010. *Demonstration of Flat Panel Display recycling technologies*. Available at: http://www.wrap.org.uk/sites/files/wrap/Flat%20Panel%20Display%20recycling%20technology%20report.pdf.

WHO, 1972. *WHO Food Additives Series, No.4: Evaluation of Mercury, Lead, Cadmium and the Food Additives Amaranth, Diethylpyrocarbonate, and Octyl Gallate*. Available at: http://www.inchem.org/documents/jecfa/jecmono/v004je07.htm.

WHO, 1990. *Environmental Health Criteria 101: Methylmercury*. Available at: http://www.inchem.org/documents/ehc/ehc/ehc101.htm.

WHO, 1991. *Environmental Health Criteria 118: Inorganic Mercury*. Available at: http://www.inchem.org/documents/ehc/ehc/ehc118.htm.

WHO, 2003. *Elemental Mercury and Inorganic Mercury Compounds: Human Health Aspects*. Available at: http://www.who.int/ipcs/publications/cicad/en/cicad50.pdf.

WHO, 2006. *Guidelines for drinking-water quality, third edition, incorporating first and second addenda*. Available at: https://reliefweb.int/report/world/guidelines-drinking-water-quality-third-edition-incorporating-first-and-second-addenda.

WHO, 2010. *Future Use of Materials for Dental Restoration*. Available at: https://www.who.int/oral\_health/publications/dental\_material\_2011.pdf.

WHO, 2011. *Replacement of mercury thermometers and sphygmomanometers in health care Technical guidance*. Available at: https://apps.who.int/iris/bitstream/handle/10665/44592/9789241548182\_eng.pdf?sequence=1&isAllowed=y.

WHO Regional Office for Europe, 2000. *Air Quality Guidelines* - *Second Edition*. Available at: http://www.euro.who.int/\_\_data/assets/pdf\_file/0004/123079/AQG2ndEd\_6\_9Mercury.PDF.

Wood, J.M., 1974. “Biological Cycles for Toxic Elements in the Environment”, *Science*, vol. 15, pp. 1043-1048.

World Chlorine Council, 2020. *World Chlorine Council Report to UNEP Global Mercury Partnership – Area on Mercury cell chlor-alkali production - Data 2019*. Available at:

<https://wedocs.unep.org/bitstream/handle/20.500.11822/32241/WCC_Chlor_alkali_2019.pdf?sequence=1&isAllowed=y>

World Nuclear Association, 2020. Storage and Disposal of Radioactive Waste. Available at: https://world-nuclear.org/information-library/nuclear-fuel-cycle/nuclear-waste/storage-and-disposal-of-radioactive-waste.aspx.

Yanase R., Hirato, O. and Matsufuji, Y., 2009. “Behaviour of Mercury from Used Batteries in Landfills over 20 Years”, *Journal of the Japan Society of Material Cycles and Waste Management*, vol. 20 No. 1, pp. 12-23.

Zero Mercury Working Group et al., 2009. *Mercury Rising: Reducing Global Emissions from Burning Mercury-Added Products*. Available at: https://www.no-burn.org/wp-content/uploads/MercuryRising\_Feb2009.pdf.

Zhang, L., Cao S., Zhu X., Zhou P., and Zhao Y., 2019. *Meeting Minamata: A Cost-Effective Compliance Roadmap for Mercury Pollution Control in China’s Non-Ferrous Metal (Zinc, Lead and Copper) Smelters* prepared for Natural Resources Defense Council.

Zhao C., and Luo K., 2017. “Sulfur, arsenic, fluorine and mercury emissions resulting from coal-washing byproducts: A critical component of China's emission inventory”, *Atmospheric Environment,* vol.152, pp. 270-278.

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1. \* The resumed fourth meeting of the Conference of the Parties to the Minamata Convention on Mercury is to convene in person in Bali, Indonesia, and is tentatively scheduled for the first quarter of 2022. [↑](#footnote-ref-2)
2. \*\* UNEP/MC/COP.4/1. [↑](#footnote-ref-3)
3. UNEP/CHW.12/5/Add.8/Rev.1. [↑](#footnote-ref-4)
4. UNEP/CHW.15/6/Add.6. [↑](#footnote-ref-5)
5. “Mercury” means elemental mercury (Hg(0), CAS No. 7439-97-6) (see Article 2 (d) of the Minamata Convention on Mercury (hereinafter referred to as “Minamata Convention”). [↑](#footnote-ref-6)
6. “Mercury compound” means any substance consisting of atoms of mercury and one or more atoms of other chemical elements that can be separated into different components only by chemical reactions (see Article 2 (e) of the Minamata Convention). [↑](#footnote-ref-7)
7. Decisions VIII/33, IX/15, BC-10/7, BC-11/5, BC-12/4, 14/8 and 15/[X], of the Conference of the Parties to the Basel Convention on the Control of Transboundary Movements of Hazardous Wastes and Their Disposal, decisions VI/20, VII/7 and OEWG-9/4 of the Open-ended Working Group of the Basel Convention, and decisions MC-2/2 and MC-3/5 of the Conference of the Parties to the Minamata Convention. [↑](#footnote-ref-8)
8. See paragraph 23 as well. [↑](#footnote-ref-10)
9. Taking into account the provisions of the Minamata Convention, the term “recover mercury from wastes” is used in this document both for cases where such recovery falls under an operation included in section B of Annex IV of the Basel Convention or an operation included in section A of Annex IV of the Basel Convention. See also paragraph 172. [↑](#footnote-ref-11)
10. The Conference of the Parties has decided that no thresholds need to be established for mercury waste falling under Art. 11(2)(a) and (b) of the Convention, and that, at present, there is no need to develop thresholds for overburden and waste rock from mining other than primary mercury mining, while further work was mandated to establish thresholds for mercury waste falling under Art. 11(2)(c) of the Convention and for tailings from industrial scale non-ferrous metal mining other than primary mercury mining (see decision MC-3/5). See Section II.B.1 for further details. [↑](#footnote-ref-14)
11. “Mercury-added product” means a product or product component that contains mercury or a mercury compound that was intentionally added (see Article 2 (f) of the Minamata Convention). [↑](#footnote-ref-15)
12. Further information on mercury, including its chemical properties, sources, behaviour in the environment and effects on human health, as well as mercury risks and pollution, is available at the following sources (see bibliography below for full references):

    * For chemical properties: Japan Public Health Association, 2001; Steffen et al., 2007; WHO, 2003; Spiegel and Veiga, 2006; ILO, 2014 and 2019; Oliveira et al., 1998; and Tajima, 1970.
    * For sources of anthropogenic emissions: UNEP, 2008a; and the Zero Mercury Working Group, 2009;
    * For behaviour in the environment: Japan Public Health Association, 2001; and Wood, 1974;
    * For human health risks: Ozonoff, 2006; Sanborn and Brodberg, 2006; Sakamoto et al., 2005; WHO, 1990; Kanai and Endou, 2003; Kerper, Ballatori, and Clarkson , 1992; Mottet et al., 1985; Sakamoto et al., 2006; Oikawa et al., 1983; Richardson, 2003; Richardson and Allan, 1996; Gay, Cox, and Reinhardt, 1979; Van Boom et al., 2003; Hylander and Meili, 2005; Bull, 2006; WHO, 1972, 1990, 1991, and 2003; Japan Public Health Association, 2001; Canadian Centre for Occupational Health and Safety, undated; Asano et al., 2000; and UNEP and WHO, 2008.
    * For mercury pollution: Ministry of the Environment of Japan, 1997 and 2002; Amin-Zaki et al., 1978; Bakir et al., 1973; Damluji and Tikriti, 1972; UNEP, 2002; Lambrecht, 1989; GroundWork, 2005; The School of Natural Resources and Environment, University of Michigan, 2000; and Butler, 1997.

    [↑](#footnote-ref-16)
13. For information on the status of individual Parties in relation to the amendment, please see the Status of Ratifications page on the Basel Convention website. [↑](#footnote-ref-18)
14. The following papers are available at the Basel Convention website:

    “Work on hazard characteristics - Approach to Basel Convention hazard characteristic H11: characterization of chronic or delayed toxicity”

    “Interim guidelines on the hazardous characteristic H12-Ecotoxic”

    http://basel.int/Implementation/TechnicalMatters/DevelopmentofTechnicalGuidelines/TechnicalGuidelines/tabid/8025/Default.aspx [↑](#footnote-ref-19)
15. This entry does not include scrap assemblies from electric power generation. [↑](#footnote-ref-21)
16. PCBs are at a concentration level of 50 mg/kg or more. [↑](#footnote-ref-23)
17. See full text of Y48 in the Convention. [↑](#footnote-ref-24)
18. “Outdated” means unused within the period recommended by the manufacturer. [↑](#footnote-ref-26)
19. [↑](#footnote-ref-27)
20. For further information, http://web.unep.org/globalmercurypartnership/ [↑](#footnote-ref-29)
21. <https://wedocs.unep.org/bitstream/handle/20.500.11822/9839/-Practical_Sourcebook_on_Mercury_Waste_Storage_and_Disposal-2015Sourcebook_Mercruy_FINAL_web.pdf.pdf?sequence=3&isAllowed=y> [↑](#footnote-ref-30)
22. http://web.unep.org/globalmercurypartnership/catalogue-technologies-and-services-mercury-waste-management [↑](#footnote-ref-31)
23. www.basel.int/Implementation/CountryLedInitiative/EnvironmentallySoundManagement/ESMFramework/tabid/3616/Default.aspx [↑](#footnote-ref-32)
24. Further guidance on Basel Convention regulatory frameworks can be found in the following documents: Model National Legislation on the Management of Hazardous Wastes and Other Wastes as well as on the Control of Transboundary Movements of Hazardous Wastes and Other Wastes and their Disposal (UNEP, 1995a), Manual for Implementation of the Basel Convention (UNEP, 2015b) and Basel Convention: Guide to the Control System (UNEP, 2015c). [↑](#footnote-ref-34)
25. Norwegian Product Regulations (Section 2.3 on mercury and mercury compounds in Chapter 2 on regulated substances, preparations and products)in Norwegian, available at:  [https://lovdata.no/dokument/SF/forskrift/2004-06-01-922/KAPITTEL\_2#KAPITTEL\_2/](http://www.miljodirektoratet.no/en/Legislation1/Regulations/Product-Regulations/Chapter-2/)  
    Special exemptions apply:

    * Limited use (concentration limits specified) in packaging, batteries, some components in vehicles and in some electrical and electronic equipment according to the European Union Regulations implemented in Norway.
    * Substances/preparations and solid processed products where the content of mercury or mercury compounds is lower than 0.001 per cent by weight.

    [↑](#footnote-ref-36)
26. For information on the status of individual Parties in relation to the amendment, please see the Status of Ratifications page on the Basel Convention website. [↑](#footnote-ref-37)
27. See, e.g., *United Nations Recommendations on the Transport of Dangerous Goods* – *Model Regulations* (21st revised edition) (United Nations, 2019a). [↑](#footnote-ref-39)
28. For bilateral agreements, see:

    <http://www.basel.int/Countries/Agreements/BilateralAgreements/tabid/1517/Default.aspx>;

    for multilateral agreements, see:

    http://www.basel.int/Countries/Agreements/MultilateralAgreements/tabid/1518/Default.aspx [↑](#footnote-ref-41)
29. Reports and publications available at: <http://web.unep.org/globalmercurypartnership/publication_resources?title=&field_gmp_partnership_area_value=product&field_resource_type_value=All> [↑](#footnote-ref-43)
30. See http://www.newmoa.org/prevention/mercury/imerc/notification [↑](#footnote-ref-44)
31. https://mercurylearn.unitar.org/ [↑](#footnote-ref-46)
32. https://www.unido.org/sites/default/files/files/2019-02/UNIDO%20Mercury%20Waste%20Forecast%202025.pdf [↑](#footnote-ref-47)
33. For example, the Czech Republic PRTR (available at <http://www.irz.cz>; in Czech only) collects specific data on mercury and mercury compounds transfered to environmental media from wastes, as well as data on how mercury wastes are handled. [↑](#footnote-ref-49)
34. Polyethylene bottles are permeable to mercury and should not be used. For details, see Parker and Bloom, 2005. [↑](#footnote-ref-53)
35. Technical guidance for implementing US EPA leach test methods 1313-1316 (*Leaching Environmental Assessment Framework (LEAF) How-To Guide* (USEPA, 2019)) is available at: https://www.epa.gov/sites/production/files/2019-05/documents/final\_leaching\_environmental\_assessment\_framework\_leaf\_how-to\_guide.pdf. [↑](#footnote-ref-54)
36. *Guidance on developing a national action plan to reduce and, where feasible, eliminate mercury use in artisanal and small-scale gold mining*. http://www.mercuryconvention.org/Portals/11/documents/forms-guidance/English/ASGM\_guidance\_e\_2017.pdf [↑](#footnote-ref-55)
37. https://web.unep.org/globalmercurypartnership/publication\_resources?title=&field\_gmp\_partnership\_area\_value=asgm&field\_resource\_type\_value=All [↑](#footnote-ref-56)
38. See https://matthey.com/en/products-and-services/chemical-processes/licensed-processes/vinyl-chloride-monomer-process. [↑](#footnote-ref-57)
39. See <http://web.unep.org/globalmercurypartnership/publication_resources?title=&field_gmp_partnership_area_value=chlor&field_resource_type_value=Guidance> (this link contains guidance for the chlor-alkali industry) [↑](#footnote-ref-58)
40. According to European Commission (2014b), the measures described in this sentence allow reducing mercury releases to water to 3-15 µg/l expressed as Hg, at the outlet of the mercury treatment unit during decommissioning or conversion. [↑](#footnote-ref-59)
41. See <http://www.mercuryconvention.org/Meetings/Intersessionalwork/tabid/8279/language/en-US/Default.aspx> [↑](#footnote-ref-60)
42. See <https://www.unenvironment.org/news-and-stories/story/many-products-still-contain-mercury-these-alternatives-could-replace-them> [↑](#footnote-ref-61)
43. See <https://www.orf.od.nih.gov/EnvironmentalProtection/MercuryFree/Pages/Alternatives.aspx> [↑](#footnote-ref-62)
44. European Union Directive 2008/98/EC provides that any natural or legal person who professionally develops, manufactures, processes, treats, sells or imports products has extended producer responsibility. [↑](#footnote-ref-64)
45. Available at: http://www.oecd.org/env/tools-evaluation/extendedproducerresponsibility.htm. [↑](#footnote-ref-66)
46. https://www.canada.ca/en/environment-climate-change/services/managing-reducing-waste/overview-extended-producer-responsibility/inventory-recycling-programs.html [↑](#footnote-ref-68)
47. https://www.recycfluo.ca/en/about-recycfluo [↑](#footnote-ref-69)
48. Information is available at https://www.keco.or.kr/en/core/operation\_extended/contentsid/1980/index.do [↑](#footnote-ref-70)
49. By way of example, guidelines are available at: http://www.newmoa.org/prevention/mercury/imerc/labelinginfo.cfm [↑](#footnote-ref-72)
50. Available at: http://www.newmoa.org/prevention/mercury/imerc/labelinginfo.cfm [↑](#footnote-ref-75)
51. Available at: https://www.nema.org/Standards/Pages/The-Labeling-of-Mercury-Containing-Lamps.aspx . [↑](#footnote-ref-79)
52. Appliance Labeling Rule, 75 Fed. Reg. 41696 (July 19, 2010). [↑](#footnote-ref-80)
53. Available at: <http://www.ftc.gov/os/2010/06/100618lightbulbs.pdf> (last visited on 29 May 2011). For information about recycling etc., see: http://www.epa.gov/cfl/cflrecycling.html. [↑](#footnote-ref-81)
54. The United States Department of Energy provides detailed guidance on the safe handling and storage of mercury, which is available at: <http://energy.gov/em/services/waste-management/waste-and-materials-disposition-information>. [↑](#footnote-ref-83)
55. See USEPA, “Cleaning up a broken CFL,” available at: http://www.epa.gov/cfl/cflcleanup.html; Mercury Policy Project, 2008, *Shedding Light on Mercury Risks from CFL Breakage*, available at: http://ledgobeyondgreen.com/images/mercury-risks.pdf; and German Environment Protection Agency, available at: http://umweltbundesamt.de/energie/licht/hgf.htm (in German). [↑](#footnote-ref-85)
56. See Glenz, Brosseau, and Hoffbeck, 2009. [↑](#footnote-ref-87)
57. Materials should be stored outdoors because many commonly available containers such as plastic bags are permeable to mercury vapour. See, Maine Department of Environmental Protection, 2008. [↑](#footnote-ref-89)
58. UNDP (GEF Global Healthcare Waste Project), *Guidance on the Clean Up, Temporary or Intermediate Storage, and Transport of Mercury Waste from Healthcare Facilities*. Available at: <http://www.gefmedwaste.org/guidanace-documents>. [↑](#footnote-ref-91)
59. For information on storage pending disposal operations (operations R13 and D15), see section III, F, 6. [↑](#footnote-ref-93)
60. Examples include pre-processing such as sorting, crushing, drying, shredding, conditioning or separating. [↑](#footnote-ref-95)
61. Exchange of wastes is interpreted to cover pre-treatment operations unless another R code is appropriate. [↑](#footnote-ref-97)
62. Such stowage of mercury waste is currently possible only in Germany. [↑](#footnote-ref-99)
63. See Article 11(3)(b) of the Minamata Convention. Furthermore, note that Article 3(5)(b) of that Convention prevents recycling of excess mercury (but not mercury wastes) from the decommissioning of chlor-alkali facilities. [↑](#footnote-ref-101)
64. See section III.J for details on worker safety precautions. [↑](#footnote-ref-103)
65. The first large-scale thermal desorption unit for the treatment of mercury-containing wastes was constructed for the remediation of the Marktredwitz Chemical Factory (CFM) in Wölsau, Germany. The operation commenced in October 1993, including the first optimizing phase. Some 50,000 tons of mercury-contaminated solid wastes were treated successfully between August 1993 and June 1996. Thermal desorption units were also used to decontaminate the old chlor-alkali plant in Usti nad Labem in the Czech Republic and to decontaminate the soil in Taipei (Chang and Yen, 2006). [↑](#footnote-ref-105)
66. Seehttps://www.remondis-qr.de/fileadmin/user\_upload/nqr/downloads/Anlagenbeschreibung\_Drehrohr\_RE%20QR\_engl\_160406.pdf. [↑](#footnote-ref-107)
67. https://batrec.ch/recycling-services/#active-carbon [↑](#footnote-ref-108)
68. See [www.gmr-leipzig.de/gbverfahren.htm](http://www.gmr-leipzig.de/gbverfahren.htm). [↑](#footnote-ref-110)
69. http://www.bethlehemapparatus.com/mercury-retirement [↑](#footnote-ref-111)
70. See https://www.remondis-qr.de/fileadmin/user\_upload/nqr/downloads/Anlagenbeschreibung\_engl\_HgS\_RE%20QR\_160406.pdf. [↑](#footnote-ref-113)
71. <https://www.remondis-qr.de/uploads/tx_3sdownloadlibrary/Anlagenbeschreibung_engl_HgS_RE_QR_160406_02.pdf> [↑](#footnote-ref-114)
72. [https://batrec.ch/recycling-services/?cn-reloaded=1#stabilisation-of-mercury](https://batrec.ch/recycling-services/?cn-reloaded=1" \l "stabilisation-of-mercury) [↑](#footnote-ref-115)
73. This section contains information provided by the National Technological Centre for Mercury Decontamination (CTNDM) (Spain). For further information, please contact: [info@ctndm.es](mailto:info@ctndm.es) or visit: http://www.ctndm.es. [↑](#footnote-ref-118)
74. There is a standard definition for sulphur polymer cement in ASTM C1159-98. [↑](#footnote-ref-119)
75. The project of an industrial plant in Spain to develop this process is already made and the budget for its construction is approved; the commercial availability of this plant is expected at the end of 2015. [↑](#footnote-ref-121)
76. This section contains information provided by Cement International Technologies S.L. For further information, please contact: [info@cementinternationaltechnologies.com](mailto:info@cementinternationaltechnologies.com" \o "mailto:info@cementinternationaltechnologies.com), or consult their website at: <http://www.cemintech.com>/en. [↑](#footnote-ref-123)
77. Germany, for example, has significant experience with underground storage of hazardous waste. [↑](#footnote-ref-126)